

High-resolution under-water laser spectrometer sensing provides new insights to methane distribution at an Arctic seepage site

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Abstract. Methane (CH₄) in marine sediments has the potential to contribute to changes in the ocean- and climate system. Physical and biochemical processes that are difficult to quantify with current standard methods such as acoustic surveys and discrete sampling govern the distribution of dissolved CH₄ in oceans and lakes. Detailed observations of aquatic CH₄ concentrations are required for a better understanding of CH₄ dynamics in the water column, how it can affect lake- and ocean acidification, the chemosynthetic ecosystem, and mixing ratios of atmospheric climate gases. Here we present pioneering high-resolution in situ measurements of dissolved CH₄ throughout the water column over a 400 m deep CH₄ seepage area at the continental slope west of Svalbard. A new fast-response under-water membrane-inlet laser spectrometer sensor demonstrates technological advances and breakthroughs for ocean measurements. We reveal decametre-scale variations of dissolved CH₄ concentrations over the CH₄ seepage zone. Previous studies could not resolve such heterogeneity in the area, assumed smoother distribution and therefore lacked both details and insights to ongoing processes. We show good repeatability of the instrument measurements, which are also in agreement with discrete sampling. New numerical models, based on acoustically evidenced free gas emissions from the seafloor, support the observed heterogeneity and CH₄ inventory. We identified sources of CH₄, undetectable with echosounder, and rapid diffusion of dissolved CH₄ away from the sources. Results from the continuous ocean laser-spectrometer measurements, supported by modelling, improve our understanding of CH₄ fluxes and related physical processes over Arctic CH₄ degassing regions.

1 Introduction

Methane (CH₄) release from gas bearing ocean sediments has been of high interest for many years (e.g. Westbrook et al., 2009; Ferré et al., 2012; Ruppel and Kessler, 2016; Jørgensen et al., 1990; Boetius and Wenzhöfer, 2013; Myhre et al., 2016; Platt et al., 2018). Once released and dissolved in the water column, the CH₄ gas diffuses and is partly oxidized in the water column (Reeburgh, 2007), contributing to minimum oxygen zones (Boetius and Wenzhöfer, 2013) and possibly to ocean acidification (Biaostoch et al., 2011). Chemosynthetic life on the seabed depends on the supply of methane as an energy resource (e.g. Boetius and Wenzhöfer, 2013). Supply of nutrient rich bottom water, by means of local upwelling, may enhance biological productivity, induce drawdown of CO₂ from the atmosphere, potentially making shallow CH₄ seepage sites sinks for this critical greenhouse gas (Pohlman et al., 2017). Warming of ocean bottom waters, active tectonics and ice sheet build up and retreat could, at different time scales, lead to CH₄ gas release from the seabed (e.g. Portnov et al., 2016). The magnitude and trend of such a phenomenon are still under debate (e.g. Hong et al., 2018; Ruppel and Kessler, 2016; Andreassen et al., 2017) and accurate methods to measure methane concentration from its source are needed. At shallow seepage sites, such as the East Siberian Arctic Shelf, CH₄ can potentially reach the atmosphere and amplify greenhouse warming (Shakhova et al., 2010; Shakhova et al., 2014). However, most studies of shallow CH₄ seepage sites have found no or little CH₄ flux to the atmosphere (e.g. Miller et al., 2017; Platt et al., 2018; Myhre et al., 2016; Gentz et al., 2014).

In the past, most CH₄ measurements relied on indirect or discrete sample measurements (e.g. Damm et al., 2005; Westbrook et al., 2009; Gentz et al., 2014). Bubble catcher and mapping with multibeam echosounder (Sahling et al., 2014), hydro-acoustic imaging together with bubble size and bubble rising speed measurements (Sahling et al., 2014; Weber et al., 2014; Veloso et al., 2015; Greinert et al., 2006; Ostrovsky, 2003) have been used to derive CH₄ flow rates. The acoustic method effectively maps CH₄ seepage from acoustically detectable sources, and camera equipped remotely operated vehicles (ROVs) can investigate their properties. However, these methods cannot detect CH₄ from sources other than free gas seepage and do not provide information about the distribution of dissolved CH₄. Discrete sampling with Niskin bottles allows 3D mapping of dissolved CH₄, but is limited by its labour intense nature, with resulting low resolution, which in turn may lead to smoothing and inaccurate estimates of CH₄ inventories. The combination of bubble catcher and multibeam echosounder is very efficient once the bubble seepage has been properly categorised, but uncertainties arise while extrapolating bubble catcher flow rates to acoustically evidenced bubble streams (flares). Present commercial underwater CH₄ sensors do not have the required response time for accurate high-resolution mapping. For this reason, Gentz et al. (2014) deployed an underwater membrane inlet mass spectrometer (UWMS) with a fast response time for mapping of CH₄ at shallow (10 m) depths. Boulart et al. (2013) used an in situ, real time sensor in the Baltic Sea. The instrument response time of 1–2 minutes and detection limit of 3 nmol l⁻¹ represent limitations for fast profiling and near surface concentration studies linked to atmospheric exchange. Sommer et al. (2015) used a pump-fed membrane inlet mass spectrometry installation at a blowout location in the North Sea. They achieved a response time of 30 minutes and a detection limit of 20 nmol l⁻¹. Wankel et al. (2010), deployed a deep-sea graded in situ mass spectrometer over a brine pool in the gulf of Mexico, where they measured high (up to 33 mM) concentrations of CH₄. They do not specify their detection limit or the response time of the instrument, but state an uncertainty of 11%. Boulart et al. (2017) mapped hydrothermal activity while deploying an in situ mass spectrometer (ISMS) over the southeast Indian Ridge. The ISMS has the advantage of measuring several dissolved gases simultaneously, but only CH₄ was reported because of the high detection limit of H₂. The ISMS response time and detection limits were not specified.

Here we present the first in-situ, high-resolution ocean laser spectroscopy mapping of dissolved CH₄ in seawater over active CH₄ seepage in the Arctic. The data was collected by deploying a patent based (Triest et al. patent France No. 17 50063) membrane inlet laser spectrometer (MILS) (Grilli et al., 2018). The high-resolution measurements, together with echosounder data, discrete water sampling, and newly developed control volume and 2-dimensional (2D) models improve our understanding of CH₄ fluxes from the seabed into oceans and lakes, and potentially to the atmosphere.

2 Materials and methods

2.1 Study area

The survey was performed on board RV Helmer Hanssen, UiT, The Arctic University of Norway, in October 2015 (CAGE 15-6 cruise) west of Prins Karls Forland located offshore western Svalbard. Over a period of three days (21–23 October), we

surveyed an area of $\sim 18 \text{ km}^2$ at water depths between 350 and 420 m, using continuous under-water laser spectroscopy as well
70 as traditional discrete sampling for dissolved CH_4 , and echosounding for bubble detection and gas seepage quantification. The
study area is located at $78^\circ 33' \text{ N}$, $9^\circ 30' \text{ E}$ over an active CH_4 venting area (Fig. 1a). Here, more than 250 flares (acoustic
signature of bubble streams in echograms) exist along the shelf break (e.g. Sahling et al., 2014; Westbrook et al., 2009; Damm
et al., 2005; Graves et al., 2015; Berndt et al., 2014). The northward flowing West Spitsbergen Current (WSC), which transports
Atlantic Water (AW, $S > 34.9$, $T > 3^\circ \text{ C}$) (Schauer et al., 2004), controls the hydrography of the study area. The East Spitsbergen
75 Current (ESC), flows south-westward along the eastern Spitsbergen coast, and northward along the western Svalbard margin,
carrying Arctic Surface Water (ASW, $34.4 \leq S \leq 34.9$) and Polar Water (PW, $S < 34.4$) (Skogseth et al., 2005). The Coastal
Current (CC), extension of the ESC (Loeng, 1991; Skogseth et al., 2005), contributes a transient addition of ASW and PW on
the shelf and the continental slope as the WSC meanders on- and offshore (Steinle et al., 2015). The Lower Arctic Intermediate
Water (LAIW, $S > 34.9$, $T \leq 3^\circ \text{ C}$) flows below the Atlantic Water (Ślubowska-Woldengen et al., 2007).

80 2.2 Hydrocasts with discrete water sampling

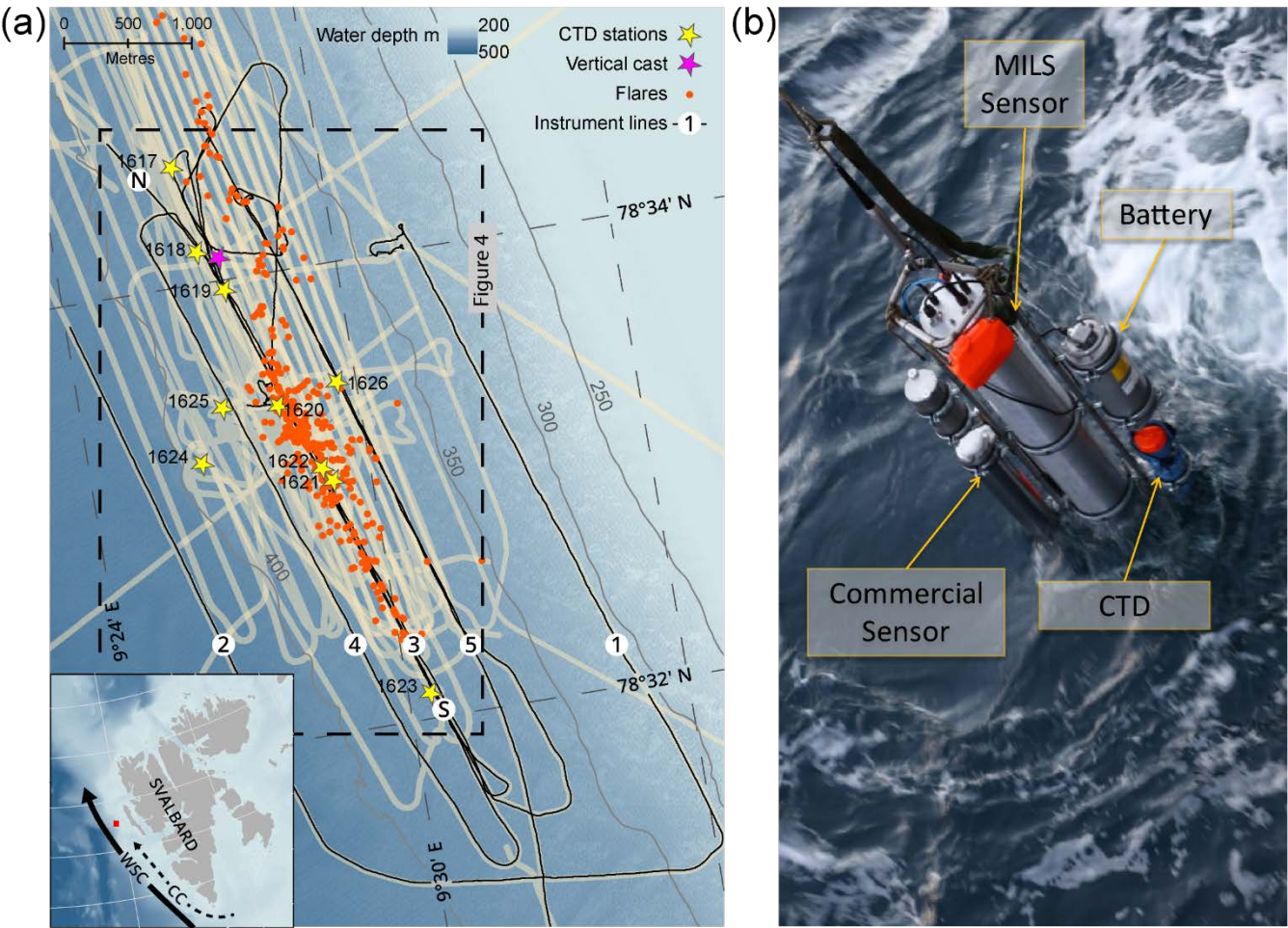
Vertical oceanographic profiles were recorded at 10 stations (Fig. 1a) using a Seabird SBE 911 plus CTD (Conductivity,
Temperature, and Depth) mounted on a rosette, which carried twelve 5-liter Niskin bottles. In January 2015, the CTD was
fitted with new sensors. An SBE 4 Conductivity Sensor and an SBE 3plus Premium CTD Temperature Sensor, with initial
accuracies of $\pm 0.001^\circ \text{ C}$ and $\pm 0.3 \text{ mS m}^{-1}$. At 24 Hz sampling, the resolutions are 0.0003° C and 0.04 mS m^{-1} .
85 The Niskin bottles were closed during the up-casts, collecting seawater at different depths for further dissolved CH_4 analysis.
Headspace equilibration followed by gas chromatography (GC) analysis was carried out in the laboratory at the Department
of Geoscience at UiT, The Arctic University of Norway, using the same technique as Grilli et al. (2018). The resulting
headspace mixing ratios (ppmv) were converted to in situ concentrations (nmol l^{-1}), using Henry's solubility law, with
coefficients calculated accordingly with Wiesenburg and Guinasso (1979). The sample dilution from addition of a reaction
90 stopper (1 ml of 1M NaOH solution replacing 1 ml of each 120 ml sample), and the removal of sample water while introducing
headspace gas (5 ml of pure N_2 replacing 5 ml of sample water) was accounted for. The overall error for the headspace GC
method was 4%, based on standard deviation of replicates.

2.3 Methodology and technology for high-resolution laser spectrometer CH_4 sensing

A stainless steel frame attached to a cable connected to an on-board winch served as a platform to which the MILS, an
95 Aanderaa, Seaguard TD262a CTD, a standard commercial CH_4 sensor, and a battery pack were mounted. This instrument
assembly, hereafter called "probe", has a total height of $\sim 1.8 \text{ m}$, a total weight in air of $\sim 160 \text{ kg}$ and a negative buoyancy of
 $\sim 52 \text{ kg}$. We towed the probe for a total of 28 hours, providing unsurpassed high-resolution in situ CH_4 measurements with a
sampling rate of 1 s^{-1} , together with dissolved oxygen data, as well as pressure, temperature and salinity. The autonomy of
the MILS was ~ 12 hours at 50 W power consumption. The sensors fitted to the Aanderaa CTD, a Conductivity Sensor 4319,

100 a Temperature Sensor 4060, and an Oxygen Optode 4330, has initial accuracies of $\pm 0.03^{\circ}\text{C}$, $\pm 5 \text{ mS m}^{-1}$, and $< \pm 8 \mu\text{M}$ and resolutions of 0.001°C , 0.2 mS m^{-1} , and $< 1 \mu\text{M}$, respectively.

Lowering and heaving of the probe in the water column allowed for vertical casts, while towing the probe behind the moving ship at varying heights above the seafloor generated near-horizontal trajectories. The main horizontal trajectories, acquired at a ship speed of 1.5 ± 0.15 knots, comprise five lines (Fig. 1a), where the desired distance from the seafloor was attained by
 105 monitoring the pressure in real time while adjusting the cable payout. The battery-powered MILS (Fig. 1b, see Grilli et al. (2018) for more details) has a membrane inlet system, linked to an optical feedback cavity-enhanced absorption spectrometer and an integrated PC for control and data storage. Cabled real-time communication with the instruments allowed instant decision-making, and ensuring optimal sensor operation during the deployments.



110 **Figure 1: Map of the surveyed area and photo of the instrument assembly.** a) Survey lines and sampling locations over the study area at the Svalbard continental margin. Black lines show the ship trajectory with line numbers assigned in the order they were surveyed. Beige areas appearing as thick lines indicate echosounder beam coverage from this campaign and previous cruises (AOEM 2010, and CAGE 13-7 in 2013). The start- and end- locations of line 3 are indicated with N and S respectively. Known flare locations from this survey and surveys in 2010 and 2013 are marked with orange dots. CTD stations with discrete water sampling are marked with yellow stars and the vertical instrument cast with a purple star. The inset image shows an overview of Svalbard with the survey location indicated with a red square. The controlling currents are shown with solid (WSC) and dashed (Coastal current) black lines.
 115 b) Instrument assembly. The main central tube is the prototype MILS sensor. The stainless steel frame acts as a platform and allows attachment of instrument battery (top right side), CTD (blue at the bottom right) and a commercial CH₄ sensor and its battery pack (left side).

Sensors with membrane inlets can be sensitive to fluctuating water flow over the membrane, which can result in artificial variability of measured concentrations. The SBE5T pump, which provided a steady water flow of 0.8 l min^{-1} during all deployments, was positioned about 25cm away from the membrane inlets and connected with short $\frac{1}{2}$ " hose sections and a T piece. By shielding the inlet and outlets and mounting them at the same height with an open flow-path, pressure changes due to movement through the water column were minimised. The water pump inlet has a fine mesh filter and a shield to avoid entry of free gas bubbles and artefacts from gas bubbles entering the sampling unit and reaching the membrane surface.

All parameters from the MILS sensor, including gas flow, pressure, sample humidity, and internal temperature were logged to process and evaluate the quality of the data. A dedicated ship-mounted GPS logged positional data for accurate synchronization of the probe and ship position. A position correction, accounting for the lag between the probe and the ship synchronizes the towed instrument data with simultaneously acquired echosounder data. The Matlab routine "Mooring Design and Dynamics" (Dewey, 1999) simulated the towing scenario, for which we used a simplified instrument assembly composed by a cylinder 1.68 m long, 0.28 m diameter with a negative buoyancy of 52 kg, corresponding to the volume and buoyancy of the whole instrument assembly. A polynomial speed-factor ($x_* = -0.2211u^5 + 1.355u^4 - 3.0126u^3 + 2.6741u^2 - 0.1609u$) was derived to account for the combined ship- and water current velocities (u in m s^{-1}). The distances of the probe behind the ship and the corresponding required time-shifts were calculated by multiplying the non-dimensional speed-factor(x_*) with the instrument depth at each data point. This approach allowed for dynamic correction of data positions, accounting for towing with or against the water current and a near-stationary ship during vertical profiling. Correction for tidal currents was neglected since tides constituted less than 5% of the WSC of $\sim 0.2 \text{ m s}^{-1}$ during our deployments, according to the tide model TPXO (Egbert and Erofeeva, 2002).

A time lag of 15 sec for the MILS was calculated based on the volume of the gas line between the extraction system and the measurement cell and the gas flow rate ($6.5 \pm 0.02 \text{ cm}^3 \text{ min}^{-1}$). We expect that concentrations profiles obtained from down- and up-casts align when this time lag is applied. The response time of the MILS is given by the flushing time of the measurement cell, and for this campaign, the T_{90} was 15 sec.

Mixing ratios of CH_4 (ppmv) measured by the MILS were converted into aqueous concentrations (nmol l^{-1}) using Henry's law, where the solubility coefficients were determined accordingly with Wiesenburg and Guinasso (1979), while accounting for in situ pressure, temperature, and salinity. The uncertainty of the dissolved CH_4 , measured with the MILS is $\pm 12\%$ (Grilli et al., 2018).

2.4 Acoustic mapping and quantification of seafloor CH_4 emissions

Gas bubbles in the water column are efficient sound scatterers and ship-mounted echosounders can therefore be used for identifying and quantifying gas emissions (Weber et al., 2014; Veloso et al., 2015; Ostrovsky et al., 2008). The target strength, defined as 10 times the 10-base logarithmic measurements of the frequency dependent acoustic cross sections (Medwin and Clay, 1997), quantifies the existence of sound scattering objects in the water column. Time series of target strength are

displayed in so-called echograms (Greinert et al., 2006;Judd and Hovland, 2009). During the cruise, the 38 kHz channel of the ship-mounted single beam Simrad EK-60 echosounder recorded acoustic backscatter continuously. Flares can be identified in the echograms and distinguished from other acoustic scatter from fish schools, dense plankton aggregations, and strong water density gradients. We identify flares as features in echograms, which exceed the background backscatter by more than 10 dB, with a vertical extension larger than their horizontal, and which are attached to the seafloor.

We used the methodology developed and corrected by Veloso et al. (2015);Veloso et al. (2019a) and the prescribed FlareHunter software for mapping and quantifying gas release. For the flow rate calculations performed with the Flare Flow Module of FlareHunter, we used a bubble size spectrum with a Gaussian distribution peaking at 3 mm equivalent radius, previously observed in the area (Veloso et al., 2015). Temperature, salinity, pressure, and sound velocities, all required for correct quantification, were provided by the CTD casts. The resulting flow rates and seepage positions allow for mass balance calculation in the control volume model and in the two-dimensional (2D) model, as described in Sect. 2.5 and 2.6, respectively.

2.5 Control volume model

The temporal evolution (dC/dt) of a solute's concentration C within a certain volume V , which is fixed in space, and with water flowing through it can, using mass conservation, be written as:

$$\frac{dC}{dt} = \frac{Q_{IN} \times C_B}{V} - \frac{Q_{OUT} \times C}{V} + \frac{F}{V} + k \nabla^2 C \quad (1)$$

Equation (1) is a second order differential equation, from which an analytical steady state solution can be derived by following these assumptions: The volumetric flow of water in and out of the control volume, Q_{IN} and Q_{OUT} are balanced and are given by a steady water current in the x-direction across the width (Δy) and height (Δz) of the control volume. The diffusion is kept homogenous and constant by applying a constant diffusion coefficient k . The background concentration C_B is fixed in time and space and F represents the persistent flow of the solute (in this case bubble mediated CH_4) into the volume. The CH_4 dissolves completely within the volume, and the diffusion occurs across the domain (in the y-direction). Using the central difference approximation of the second derivative (∇^2 in Eq. (1) and the above assumptions yield that the aqueous CH_4 within the volume reaches the steady state concentration:

$$C_{t=\infty} = \left(\frac{Q_{IN} \times C_B}{V} + \frac{F}{V} + \frac{2k \times C_B}{(\Delta y)^2} \right) \times \left(\frac{Q_{OUT}}{V} + \frac{2k}{(\Delta y)^2} \right)^{-1} \quad (2)$$

Finally, by averaging measured CH_4 concentration within a defined volume, and assuming that it represents a steady state concentration, the bubble flow rate is retrieved from Eq. (2).

$$F = (\bar{C} - C_B) \times \left(Q + \frac{V \times 2k}{(\Delta y)^2} \right) \quad (3)$$

Where $\bar{C}$ represents the measured average concentration, and $Q = Q_{IN} = Q_{OUT}$.

The dimensions of the control volume with volume $V = \Delta x \times \Delta y \times \Delta z$, were chosen to match the length of line 3 ($\Delta x = 4.5$ km), extended 25 m perpendicularly on each side of the line ($\Delta y = 50$ m), and 75 m vertically ($\Delta z = 75$ m). A graphic describing the control volume is supplied in Fig. SI 1.

2.6 Two-dimensional model

In order to gain insight to the physical processes behind the observed CH₄ variability, we constructed a two-dimensional (2D) numerical model, resolving the evolution of dissolved CH₄ in the water column, which results from CH₄-bubble emissions, advection with water currents and diffusion. The model domain was made 400 m high in the z-direction, 4.5 km long in the x-direction, and oriented along line 3 (Fig. 1a). The navigation data along this line is linearly interpolated to form the basis for a 2-metre gridded model domain starting at 78°34.54' N, 9°25.92' E and ending at 78°32.1' N, 9°30.58' E as indicated by N and S, in Fig. 1a. FlareHunter derived flow rates within 50 m from line 3 were projected into the model domain, and the source of dissolved CH₄, mediated by bubbles, was distributed vertically by applying a non-dimensional source-function similar to the approach of Jansson et al. (2019): $S_0(z) = 6.6 \times 10^{-2} \times e^{-0.066 \times z}$, where z is the vertical distance from the seafloor in metres. We calculated source distribution functions $S(z)$ by scaling $S_0(z)$ with the flare flow rates, and distributed the resulting source into current-corrected x/z nodes with volumes $\delta V = \delta x \times \delta y \times \delta z$, where $\delta x = \delta y = \delta z = 2$ m. The model domain comprises 12 extra cells on each side in the y-direction in order to avoid fast diffusion out of the domain while the background concentration is held constant. The 2D model simulated CH₄ diffusion and advection with water currents, and was run to steady state using different diffusion coefficients, within the range suggested by Sundermeyer and Ledwell (2001). A graphic representation of the 2D model is shown in Fig. SI 1.

3 Results

3.1 Water properties

The measurements from the Seabird CTD during our survey indicate well-mixed water within 150 metres above the seafloor, and continuously stratified water upwards to 50 metres below the sea level (mbsl) (Fig. 2a) with a squared buoyancy frequency of $\sim N^2 < 4 \times 10^{-5} \text{ s}^{-2}$. A pycnocline exists at ~ 30 mbsl (Fig. 2a) with N^2 up to 10^{-4} s^{-2} , marking the transition between surface water and AW below (Fig. 2b and 2c). Temperatures close to the seafloor range from 4.2–4.4 °C, which is more than 1 °C above the CH₄ hydrate stability limit (Tishchenko et al., 2005), for a salinity of 35.1 as indicated in Fig. 2a. The velocity of the WSC was between 0.1 and 0.3 m s⁻¹ (Fig. 2d) inferred from the inclination of flare spines (Veloso et al., 2015), which was calculated from the echosounder data, obtained throughout the whole survey. The current followed the isobaths, which is consistent with previous findings (Graves et al., 2015; Gentz et al., 2014). The mean salinity and temperature acquired with the Andreaa CTD, in different layers, with their corresponding standard deviations according to the water masses classification of Skogseth et al. (2005) and Ślubowska-Woldengen et al. (2007) are shown in Fig. 2b,c. The temperature/ salinity distribution suggests a clear dominance of AW during the survey, overlaid with fresher and colder ASW and PW.

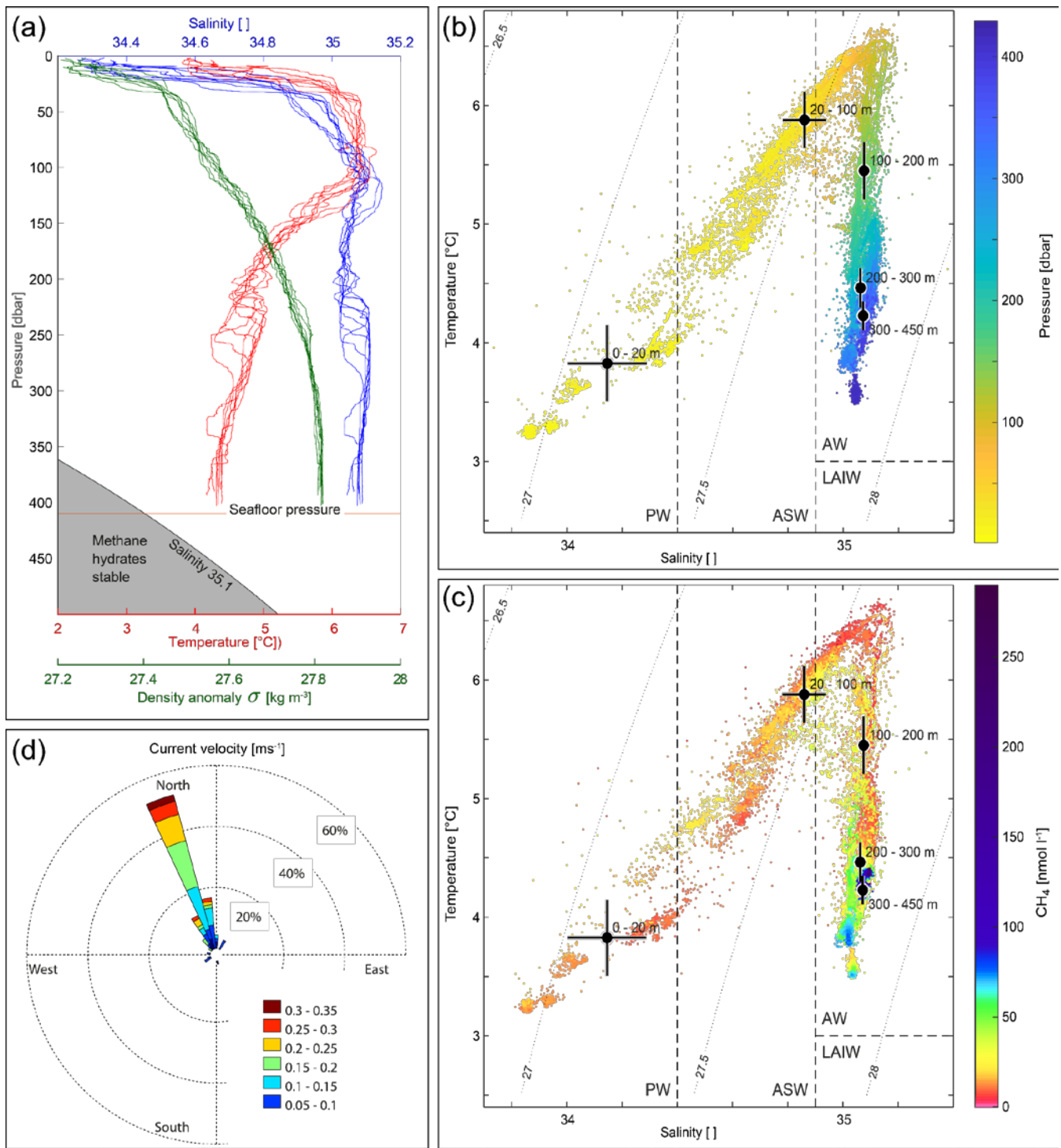


Figure 2. Hydrography during the survey. a) CTD casts 1617- 1626 showing temperature (red), salinity (blue) and density anomaly (green) calculated with Gibbs seawater package (McDougall and Barker, 2011). b) Temperature and salinity diagram coloured by pressure (dbar). Grey curved lines in the background indicate isopycnals (constant density (σ) lines). AW indicates Atlantic Water, PW is Polar Water, ASW is Arctic Surface Water and LAIW is Lower Arctic Intermediate Water. Water mass definitions are described in the text. Black dots indicate the mean water properties for the different layers and crosses indicate the corresponding standard deviations. c) Temperature and salinity diagram coloured by CH_4 concentrations (nmol l⁻¹) measured with the MILS. Black dots depict average temperature and salinity at water depth intervals, and the error bars indicate the corresponding standard deviations. d) Water currents inferred from inclination of flare spines (Veloso et al., 2015) with a mean bubble rising speed of 23 cm s⁻¹.

3.2 Measured and modelled CH₄ distribution

The high-resolution dissolved CH₄ concentration profiles resulting from towing the MILS along five lines, approximately 15 meters above the seafloor (masf) show high variability (Fig. 3), especially over line 3, which geographically matches the clustering of bubble plumes (Fig. 1a).

On the landward side (lines 1 and 5), the concentration is relatively smooth with an average of $\sim 55 \text{ nmol l}^{-1}$, but along line 5, which is closer to the main seepage area, the concentration is influenced by the nearby seepage, inferred from the concentration peaks reaching up to 105 nmol l^{-1} at $78^\circ 33.5' \text{ N}$. On the offshore side, the mean concentrations are 15 and 36 nmol l^{-1} along lines 4 and 2, respectively with elevated CH₄ concentrations of up to $\sim 70 \text{ nmol l}^{-1}$, lacking hydroacoustic evidence of CH₄ seep sources. The peak in line 4 may be explained by its proximity to the main bubble seep cluster, but the CH₄ concentrations show more variability along line 2, the offshore-most horizontal trajectory of the survey, which may indicate undetected CH₄ seepage located deeper than 400 mbsl.

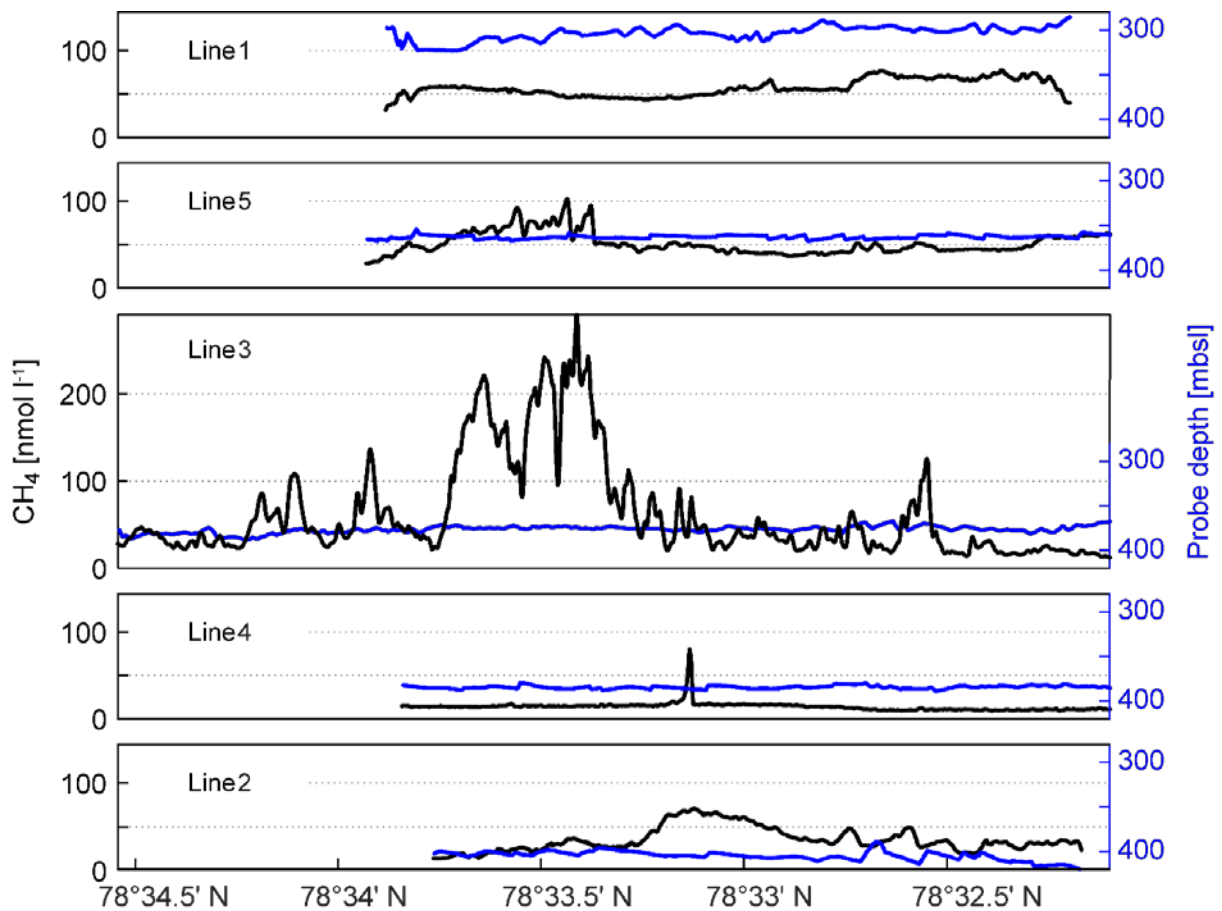


Figure 3. MILS measurements along the five lines $\sim 15 \text{ m}$ from the seafloor. Panels show data acquired along lines 1–5, shown in order of proximity to the shore, with line 1 closest to the shore and line 2 furthest offshore. See Fig. 1a for line locations. Black lines show CH₄ concentrations, blue lines show the probe depth.

A 25-minute down- and upward sequence obtained from the vertical MILS cast at station 1616 (Fig. 4) shows excellent repeatability after correcting for the instrument time lag of 15 seconds. The sensor showed no memory effects, i.e. different response times between increased and decreased CH₄ concentrations.

Analysis of discrete samples (DS) from CTD casts 1618 and 1619 and the vertical MILS cast 1616 give further insights to the
240 heterogeneity and temporal variation of the dissolved CH₄ distribution (Fig. 4). Discrete measurements from CTDs 1618 and
1619 reveal a qualitative match with the MILS measured concentrations extracted from line 3 near these stations (red and
green symbols in Fig. 4). Discrepancies between the MILS cast 1616 and the DS from CTDs 1618 and 1619 close to the
seafloor is likely due to the difference in sampling location, as the MILS vertical cast 1616 was ~150 and ~180 m away from
CTDs 1618 and 1619, respectively.

245 The exponential “dissolution” function, which represents the expected trace of dissolved CH₄ in the water column, resulting
from bubble dissolution, was compared to the entire MILS dataset by plotting CH₄ concentrations against height above the
seafloor, determined from position corrected pressure and previously acquired multibeam data (Fig. 4).

Elevated CH₄ concentrations at ~160 and ~220 mbsl revealed by the MILS vertical profile 1616 was not identified with DS
from the nearby CTD cast 1619, and DS from CTD 1618 reveal only a small fraction of the CH₄ anomaly, because of too
250 sparse sampling (Fig. 4). The MILS data collected 15 masf along line 3 reveals 50 nmol CH₄ l⁻¹ while the vertical profile only
30 metres away (MILS-cast 1616), measured ~200 nmol CH₄ l⁻¹ (Fig. 4). This emphasizes the strong spatiotemporal variability
of the CH₄ distribution in the area.

Despite the high CH₄ variability in the horizontal profiles (Fig. 3), further analysis of the data may be obtained by focusing on
line 3, towed in north-south direction at ~0.8 m s⁻¹ directly over the bubble streams. Based on a mean depth of 390 m and the
255 depth of the towed CTD, the height above the seafloor of the towed probe along line 3, was 13.4 ± 3.8 m. The fast response
time of the MILS sensor ($T_{90} = 15$ s) revealed decametre-scale variations of the dissolved CH₄ concentrations with high values
well correlated with the echosounder signal, after correcting for the towed instrument position (Fig. 5).

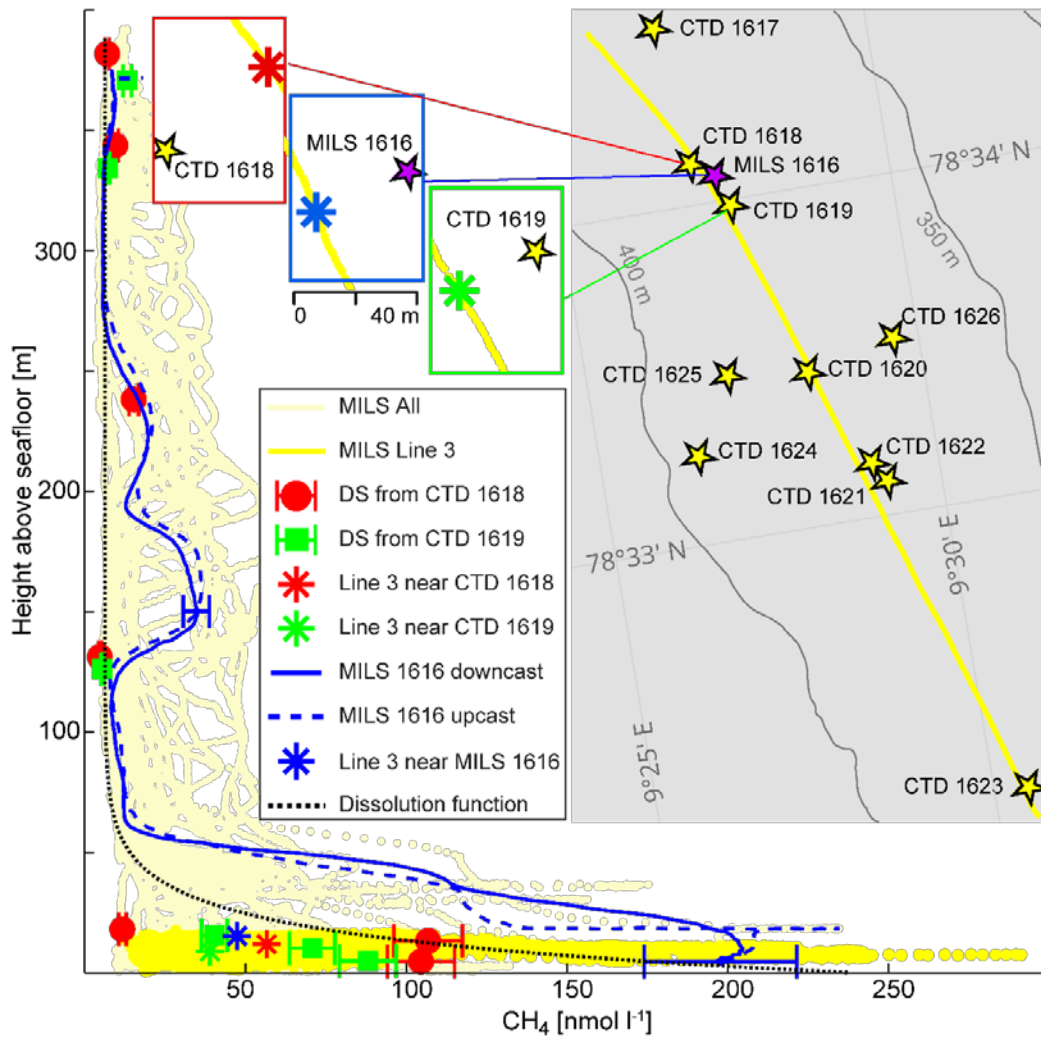


Figure 4. High-resolution CH_4 concentrations and discrete samples. Light yellow lines show CH_4 concentrations acquired with the MILS during the entire survey and the bright yellow derives from line 3 at ~15 masf only. Solid and dashed blue lines represent continuous down- and upward profiles acquired at station 1616 after correction for instrument response time. The blue error bars indicate the instrument uncertainty of 12%. Discrete sample data is shown as red dots (CTD 1618) and green squares (CTD 1619) with error bars that indicate the discrete sampling/ headspace GC method uncertainty of 4%. The asterisks indicate MILS data points from the towing along line 3, closest to the vertical cast 1616 (blue), to CTD 1618 (red) and CTD 1619 (green). The black dotted line indicates the exponential dissolution function described in the text. The inset map shows the locations of the CTDs with discrete sampling (stnr1617–1623) (yellow stars) as well as line 3, which is indicated with a yellow line. The blue rectangle shows the location of the vertical MILS profile from station 1616 (purple star) and the data point from line 3, which is closest to the deepest location of the vertical cast (blue asterisk). The green rectangle shows the location of CTD 1619 and the closest point on line 3 (green asterisk), while the red rectangle shows the location of CTD 1618 and the corresponding point on line 3 (red asterisk).

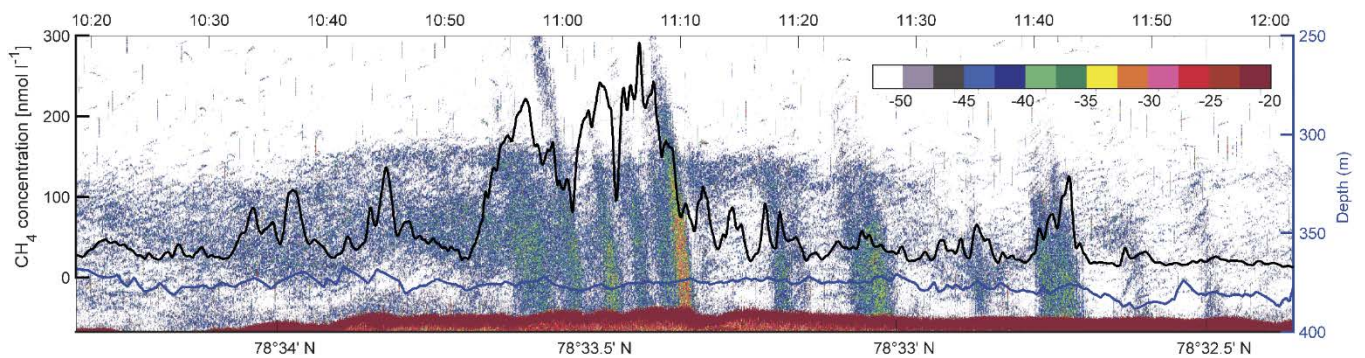


Figure 5. Towed MILS data overlaying echo-sounder data. The black line shows the CH_4 concentration along line 3 (see Fig. 1 for location) at ~15 m from the seafloor. The blue line indicates the depth of the probe. The echogram, displaying target strength values (colour bar shows intensity (-dB)) from the 38 kHz-channel of the EK60, is shown in the background.

275 A close analysis of the measured concentration reveals that the up- and down- stream gradients are equally distributed (bar chart in Fig. SI 2c). This symmetry suggests that CH₄ disperses fast and equally in all horizontal directions around the bubble plumes while being advected away from the source.

The measured CH₄ concentrations along line 3 changed significantly (5% or more) on sub-response times (<15 s) in only two instances and over a total time of 26 s, out of 1h 42 min, as indicated with red dots in Fig. SI 2a. This suggests that the MILS
280 resolved 99.6% of the gradients and that the response time of the MILS did not limit the resolution of the CH₄ distribution. The mean absolute gradient, assessed from steadily increasing or decreasing concentrations (grey vertical lines in Fig. SI 2a show the position of the selected slopes), was 1.5 nmol l⁻¹ m⁻¹, corresponding to 1.2 nmol l⁻¹ s⁻¹. The minimum and maximum lateral gradients were -5.0, and 4.8 nmol l⁻¹ m⁻¹, which corresponds to -4.1 and 4.6 nmol l⁻¹ s⁻¹. Correlations of CH₄ concentrations versus depth and speed changes were low (R= 0.0133, -0.0001, -0.0094, 0.0028 for ship speed, ship
285 acceleration, vertical instrument speed and instrument acceleration, respectively), showing the stability of the instrument during rapid movements and disproving artefacts due to water flow fluctuations at the membrane.

Sources of CH₄ constraining the control volume and 2D model were obtained from the acoustic mapping and quantification described in section 2.4. During the entire survey, we identified 68 unique groups of bubble plumes, with an average flow rate of 48 (SD = 50) ml min⁻¹. Within 50 metres of line 3, we acoustically identified 31 flares with an average flow rate of 60 (SD
290 = 65) ml min⁻¹ amounting to a total flow rate of 1.87 l min⁻¹. These flow rates were taken as sources in the control volume and 2D model. FlareHunter calculates the flow rates in a layer 5–10 m above the seafloor. In order to calculate flow rates from the seafloor, we upscaled the FlareHunter flow rates by 40% to compensate for bubble dissolution near the seafloor, in accordance with the dissolution profile.

The 2D model was run to steady state with different diffusion coefficients, $k \in [0.3 - 4.9 \text{ m}^2 \text{ s}^{-1}]$, adopted from dye-
295 experiments offshore Rhode Island (Sundermeyer and Ledwell, 2001). These coefficients are in agreement with the ones obtained from the Celtic Sea ($k \in [0.8 - 4.4 \text{ m}^2 \text{ s}^{-1}]$) (Stashchuk et al., 2014), but much higher than the coefficient applied by Graves et al. (2015) ($k = 0.07 \text{ m}^2 \text{ s}^{-1}$). The best fit between the 2D model and the MILS data (R = 0.68) was achieved during a simulation with $k = 1.5 \text{ m}^2 \text{ s}^{-1}$. Because the high-end coefficients of Sundermeyer and Ledwell (2001) and Stashchuk et al. (2014) were derived during wavy conditions, and because our model mainly resolves the near-bottom region, away from
300 wave action, we interpret that our best-fit diffusion coefficient is relatively high. The resulting range of model outputs and the best fit-model simulation are visualized and compared with high-resolution measurements in Fig. 6. Despite applying a high diffusion coefficient, the 2D model shows a residual downstream tailing, which is not seen in the MILS data. We attribute this to the fact that the model does not resolve small scale eddies, but only diffusion across the domain and diffusion/ advection along the domain.

305 The salinity and temperature profiles of the towed CTD indicate well-mixed water, particularly over the most prominent gas flares. Here, the relative standard deviation of the salinity and temperature drops by factors of 10, and 58 respectively, as

highlighted by the dashed-line box in Fig. 6. The depth stability of the probe is also better in the area. Its relative standard deviation dropped by a factor of 3, which is not enough to justify the larger factors observed for the temperature and salinity. We interpret that this is caused by turbulent mixing enhanced by the bubble streams.

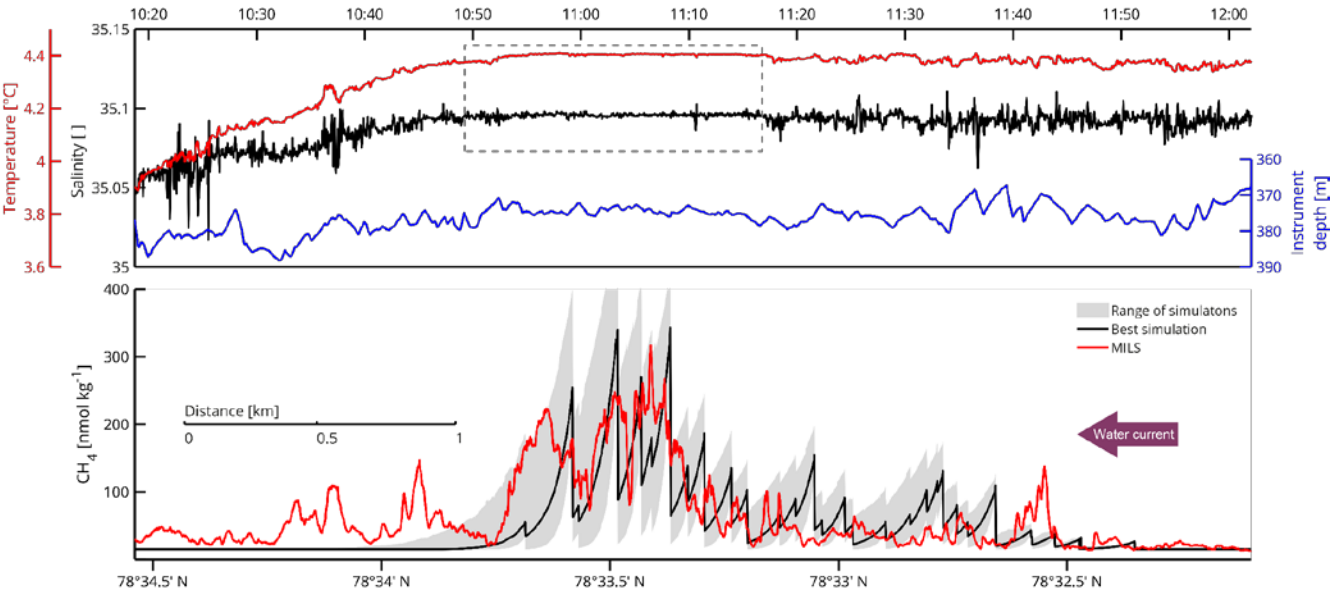


Figure 6. Water properties and comparison between modelled and measured dissolved CH₄ concentrations along line 3. Top panel shows temperature and salinity data together with the depth of the towed instruments. The dashed-line box highlights the area of intense mixing. In the lower panel, the red line shows the dissolved CH₄ measured by the MILS. The grey area indicates the range of CH₄ concentrations from the 2D model simulations. The black line depicts output of the model simulation with the best match with the measured concentrations.

3.3 Methane inventory

The method, dimensions, and resolution chosen for calculating CH₄ inventories may strongly influence the resulting content and average concentrations. This may have serious implications when the results are used for upscaling. To highlight this, we applied different inventory calculation methods on the same water volume.

Averages along line 3 were calculated from: a) Concentrations from discrete sampling, based on different sampling depths. b) Discrete data from different depths, linearly interpolated along the line. c) High-resolution data obtained from the MILS data ~15 masf. d) Concentrations extracted from the 2D model output at steady state at 15 masf.

Average concentrations were calculated in a “box” volume equivalent to MILS line 3 (4.5 km long (x-direction), 50 m wide (y-direction), equivalent to the echosounder beam width, 75 m high (z-direction) corresponding to the most dynamic and CH₄ enriched zone (e.g. McGinnis et al., 2006; Jansson et al., 2019; Graves et al., 2015). Box averages were derived as follows: The volume was divided into 1 m cubic cells. Cells located in the y-centre and in z-positions vertically matching the underlying data (DS or MILS) were populated with the MILS, or interpolated DS profiles. The remaining cells were populated by perpendicular and vertical extrapolation following the typical horizontal gradient of 1.5 nmol l⁻¹ m⁻¹, and vertical dissolution profiles scaled by the measured or interpolated concentrations. The mean concentrations from the 2D model was delimited by the height of the box. The control volume model provided only one value for the entire box.

The underlying data and its interpolation is seen in Fig. 7 and the resulting averages are reported in Table 1.

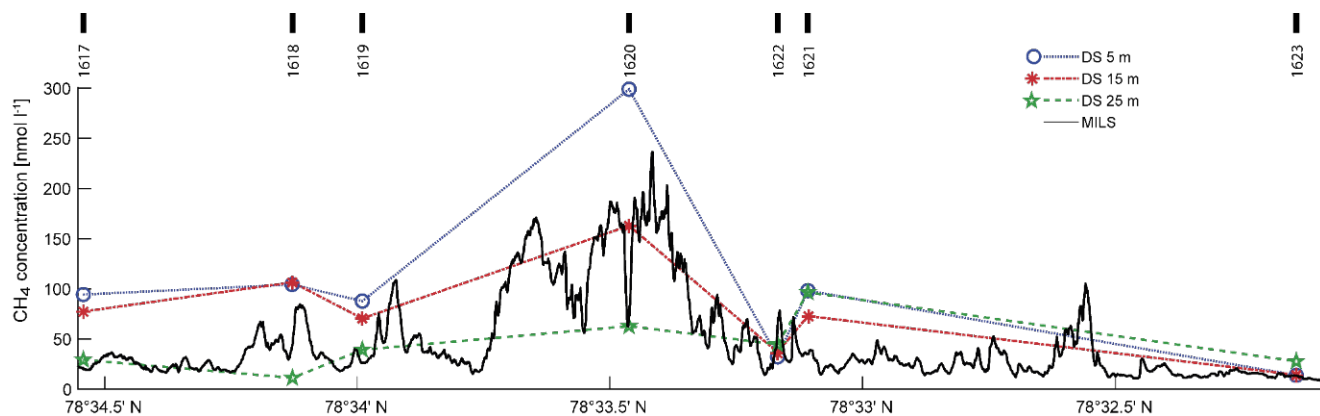


Figure 7: Underlying CH₄ concentration data for inventory calculations. Solid black line shows the continuous MILS profile ~15 masf along line 3. DS concentrations at various depths are shown as blue circles, red asterisks and green stars for 5, 15 and 25 masf respectively, and blue dotted, red dash-dotted, and green dashed lines represent the corresponding linear horizontal interpolations. CTD cast numbers are marked with thick black lines at the top of the graph.

The average CH₄ concentration in the box volume based on continuous data is similar to the average obtained from discrete data at 15 m above the seafloor. We obtained 47 vs 77 nmol l⁻¹ for the high-resolution line and the interpolated DS, while the box averages for the high-resolution and interpolated DS were 22 vs 29 nmol l⁻¹. The 2D model yielded a line average of 60 nmol l⁻¹, while it was 22 nmol l⁻¹ for the box. The control volume model predicted a steady state concentration of 23 nmol l⁻¹ when the diffusion coefficient of 1.5 m² s⁻¹, inferred from the 2D model was applied.

4 Discussion

During our survey, the mean flow rate at the seafloor per flare within 50 m of line 3, was 84 (SD = 91.6) ml min⁻¹, (min = 15.8, max = 355.6 ml min⁻¹). This is comparable with the flow rate per flare of 125 ml min⁻¹, estimated by Sahling et al. (2014) who assumed that an acoustic flare consists of 6 bubble streams, each with a flow rate of 20.9 ml min⁻¹. The authors found 452 flares in the area, for which they assumed similar flow rates, and thereby calculated a total flow in the area of ~57 l min⁻¹. Our study encompasses a smaller area, where we only detected 68 flares (31 flares within 50 m of line 3) and the total flow rate from these 68 flares was 4.56 l min⁻¹. This total flow translates to 65.7 t CH₄ y⁻¹ assuming constant ebullition. Considering the sparse beam coverage and relatively small area, this may be compared to CH₄ seepage of ~550 t CH₄ y⁻¹ estimated for a larger area, covered by 9 surveys (Veloso et al., 2019b), and ~400 t CH₄ y⁻¹ (Sahling et al., 2014), in a study area covering ours, but also extending northwards, where additional gas venting occurs. A comparison of studies from the same area, using different methods, shows a large range of yearly CH₄ emissions to the water column. Flow rates of CH₄ per distance along the continental shelf from previous studies given by the authors (900 kg m⁻¹ y⁻¹ (Westbrook et al., 2009); 141 kg m⁻¹ y⁻¹ (Reagan et al., 2011); 13.8 (6.9–20.6) t m⁻¹ y⁻¹ (Marín-Moreno et al., 2013); 2400 (400–4500) mol m⁻¹ y⁻¹ (Sahling et al., 2014); 748 (561–935) t m⁻¹ y⁻¹ (Berndt et al., 2014)) yield emissions of 4050, 635, 992, 173 and 54000 t CH₄ y⁻¹ over the 4500 m section of the continental shelf which corresponds to our line 3.

The MILS data collected 15 masf along line 3 did not reveal the high concentrations ($\sim 200 \text{ nmol l}^{-1}$) measured during the vertical cast only 30 m away, emphasizing the heterogeneous CH_4 distribution, and highlighting the need for high-resolution
360 sensing, rather than sparse discrete sampling.

The fast response of the MILS helped revealing decametre scale variability of dissolved CH_4 , and we conclude that uncertainties introduced by MILS response time were negligible in this survey. The observed symmetry of CH_4 gradients suggests fast dispersion in all horizontal directions while enriched water is advected away from the sources.

Because the instrument assembly lacked an inertia measurement unit, the stability during towing is unknown, but we did not
365 observe any effect on the measurements from wobble and/ or rotation.

Gentz et al. (2014) and Myhre et al. (2016) suggested that a pronounced pycnocline is a prerequisite to limit the vertical transport of dissolved CH_4 towards the surface. One should note that this hypothesis was based on discrete sample data, rather than high-resolution data. We observed high CH_4 concentrations up to 75–100 masf, which is in agreement with bubble models (e.g. McGinnis et al., 2006; Jansson et al., 2019), highlighting that bubbles of observed sizes ($\sim 3 \text{ mm}$ average equivalent radius)
370 are fully dissolved within this range. Density stratification plays an important role in the vertical distribution of dissolved CH_4 because turbulent energy is required to mix solvents across isopycnals. Vertical mixing is therefore inhibited even without the presence of a strong pycnocline. We suggest that the observed height limit is a result of rapid bubble dissolution and inefficient vertical mixing, regardless of the existence of a pronounced pycnocline.

We observed CH_4 concentrations of up to 100 nmol l^{-1} without the acoustic signature of flares north of the active flare zone
375 (Fig. 5). Echograms from the CAGE 15-6 survey (this work) and previous surveys conducted in 2010 (AOEM 2010 cruise, University of Tromsø, with RV Jan Mayen) and 2013 (CAGE 13-7 cruise, with RV Helmer Hanssen (e.g. Portnov et al., 2016)) reveal that the nearest bubble stream is located $\sim 300 \text{ m}$ northeast of this CH_4 anomaly. Several hypotheses may explain this CH_4 enrichment: a) nearby presence of CH_4 enriched water seepage (hypothetically from dissociating hydrates) from the seafloor; b) presence of bubble streams with bubbles too small to be detected by the echosounder (the detection limit (target
380 strength $< -60 \text{ dB}$) of a single bubble was 0.42 mm for this survey); c) advection of CH_4 -enriched water from an upstream bubble plume source, not detected by the echosounder. In our case, the temperature- and salinity anomaly, which coincides with the increased CH_4 , reveals mixing of AW with colder and fresher water (Fig. 6). Because mixing lines drawn in the Temperature and salinity diagrams (Fig. 2b and 2c) point towards PW rather than a pure fresh water source, our data supports hypothesis c, namely that AW mixed with PW was transported, and enriched in CH_4 while passing over a bubble plume before
385 reaching the location of the measurement. Lateral eddies or bottom Ekman transport may have been responsible for the intrusion of fresh, cold, CH_4 enriched water.

The 2D model relies on acoustically detected bubble plume locations, and the difference between measured and modelled CH_4 is obvious along line 3 from 10:30 to 10:50 as seen in Fig. 6. The CH_4 signal from high-resolution data, not thoroughly resolved by the model, underscores that mapping and modelling based on echosounder data are not enough for a correct quantitative

390 estimate of the CH₄ inventory. The 2D model required a high diffusion coefficient in order to reproduce the variability of measurements, which is supported by high turbulence in the area, caused by the strong currents. Downstream tailing of CH₄ concentrations seen in the 2D model was not observed with the MILS. In fact, MILS data reveal equal distribution of down- and upstream concentration gradients. We explain the discrepancy by the fact that the 2D model does not resolve eddies and the CH₄ source is placed in discrete cells, following a theoretical straight bubble line, and not accounting for diffusion along
395 the bubble paths.

The relatively high midwater (120–260 mbsl) CH₄ concentrations revealed by the vertical MILS cast 1616 was only partly observed in the discrete sampling and was not inferred from echosounding. We suggest that this discrepancy is attributed to seepages at the corresponding depth interval, not previously mapped. The closest known seepages are a few km away from the location, at the shallow shelf (50–150 mbsl) and at the shelf-break (~250 mbsl) (Veloso et al., 2015), but it is doubtful that
400 water masses from these locations can reach the surveyed area, as the WSC is persistently northbound. Unless horizontal eddies transport CH₄ from the shelf-break to this area, this result indicates the existence of undiscovered CH₄ bubble plumes further south, at the depth of the observed anomaly.

The high-resolution data from the MILS results in a significantly lower CH₄ inventory than the one obtained from discrete sampling (47 vs 77 nmol l⁻¹) due to the heterogeneous distribution of dissolved CH₄. The choice of discrete sample locations
405 can significantly affect the resulting average concentration. The average CH₄ concentration (93 nmol l⁻¹) estimated by Graves et al. (2015) from a box with dimensions $\Delta x = 1$ m, $\Delta y = 50$ m, $\Delta z = 75$ m, obtained from a DS transect across the slope, was substantially higher than our box estimates of 20–39 nmol l⁻¹. These two results highlight the need for high-resolution sensing when estimating CH₄ inventories and average CH₄ concentrations.

The optical spectrometer of the MILS can be tuned or replaced to improve its sensitivity or to sample more CH₄ enriched
410 waters. We believe the MILS would be an excellent tool for evaluating CH₄ related water column processes. Grilli et al. (2018) reported a sensitivity of ± 25 ppbv in air, which translates to ± 0.03 nmol l⁻¹ at 20 °C and a salinity of 38, which is low enough for investigations of atmospheric exchange and CH₄ production/ consumption rates.

5 Conclusion

We have presented new methods for understanding the dynamics of CH₄ after its release from the seafloor, coupling for the
415 first time continuous high-resolution measurements from a reliable and fast CH₄ sensor (MILS) with dedicated models. The MILS sensor was successfully deployed as a towed body from a research vessel and provided high-resolution, real-time data of both vertical and horizontal dissolved CH₄ distribution in an area of intense seepage west of Svalbard. For the first time, we observed a more heterogeneous CH₄ distribution than has been previously presumed.

We employed an inverse acoustic model for CH₄ seepage mapping and quantification, which provided the basis for a new 2D
420 model and a new control volume model, which both agreed relatively well with observations. The 2D model did not reproduce

the symmetric gradients observed with the MILS, which suggests a need to improve the model by including turbulent mixing enhanced by the bubbles streams.

Despite the large spatial and temporal variability of the CH₄ concentrations, a comparison between high-resolution (MILS) and DS data showed good general agreement between the two methods.

425 Heterogeneous CH₄ distribution measured by MILS matched acoustic backscatter, except for an area with high CH₄ concentrations without acoustic evidence of CH₄ source. Similarly, high midwater CH₄ concentration was observed by the MILS vertical casts with little evidence of a nearby CH₄ source, further supporting that high-resolution sensing is an essential tool for accurate CH₄ inventory assessment and that high-resolution sensing can give clues to undetected sources.

CH₄ inventories, given by discrete sampling agreed with those from high-resolution measurements, but sparse sampling may
430 over- or underestimate inventories, which may have repercussions if the acquired data is used for predicting degassing of CH₄ to the atmosphere in climate models. The added detail of the fine structure allows for better inventories, elucidates the heterogeneity of the dissolved gas, and provides a better insight to the physical processes that influence the CH₄ distribution. The methods for understanding CH₄ seepage presented here shows potential for improved detection and quantification of dissolved gas in oceans and lakes. Applications for high-resolution CH₄ sensing with the MILS include environmental and
435 climate studies as well as gas leakage detection desired by fossil fuel industry.

6 Data availability

A dataset comprising the MILS sensor data and echosounder files is available from the UiT Open Research Data repository <https://doi.org/10.18710/UWP6LL>.

7 Author contribution

440 PJ and JT contributed equally to the work leading to this manuscript, and are listed alphabetically. JM and JC initiated the collaboration and designed the study. JT, RG, PJ, and JM planned and participated in the field campaign. JT and RG systematized and synchronized the high-resolution MILS data. PJ mapped and quantified gas seepage and water currents by collecting and analysing echosounder data. PJ collected and analysed the CTD data and water samples, and organized the data repository. RG converted mixing ratios to aqueous concentrations. PJ developed the numerical models. AS analysed the water
445 mass properties. JT and PJ calculated CH₄ inventories. All authors contributed to the manuscript.

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455 9 Competing interests

None of the authors reported competing interests.

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Table 1: Average concentrations (nmol l⁻¹) calculated with different methods at different altitudes as indicated in the first column (metres above the seabed, masf). ¹Average of the sparse discrete sampling from CTD casts 1617–1623 . ²Average of high-resolution (MILS) measurements from line 3. ³Average of linearly interpolated concentrations based on discrete measurements (CTDs 1617–1623). ⁴Average concentrations from the 2D model, extracted from depths matching the MILS position along line 3. ⁵Average concentrations within the box (4500 m (L) × 50 m (W) × 75 m (H)) based on the high-resolution measurements or interpolated concentrations along the box, vertical distribution based on the dissolution-profile, and horizontal distribution across the width of the box based on the mean horizontal concentration gradient (Fig. SI 2). ⁶Average of the 2D model (best simulation) result from 0–75 masf and across the domain. ⁷The CV model yields a box value only.

Dataset	Discrete	High-resolution	Box
MILS ~15 masf	-	47 ²	22 ⁵
DS ~5 masf	104 ¹	108 ³	39 ⁵
DS ~15 masf	77 ¹	77 ³	29 ⁵
DS ~25 masf	44 ¹	49 ³	20 ⁵
2D model (~15 ⁴ and 0–75 ⁶ masf)	-	60 ⁴	22 ⁶
CV model	-	-	23 ⁷