



1 **Concomitant ocean acidification and increasing total alkalinity at a coastal site in the NW**
2 **Mediterranean Sea (2007-2015)**

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15 **Abstract.** Monitoring global ocean change is necessary in coastal zones due to their physical and
16 biological complexity. Here, we document changes in coastal carbonate chemistry at the time-
17 series station, Point B, in the NW Mediterranean Sea, from 2007 through 2015, at 1 and 50 m.
18 The rate of surface ocean acidification (-0.0028 ± 0.0003 units $\text{pH}_T \text{ yr}^{-1}$) was faster-than-
19 expected based on atmospheric carbon dioxide forcing alone. Changes in carbonate chemistry
20 were predominantly driven by an increase in total dissolved inorganic carbon (C_T , $+2.97 \pm 0.20$
21 $\mu\text{mol kg}^{-1} \text{ yr}^{-1}$), > 50 % of which was buffered by a synchronous increase in total alkalinity (A_T ,
22 $+2.08 \pm 0.19 \mu\text{mol kg}^{-1} \text{ yr}^{-1}$). The increase in A_T was unrelated to salinity and its cause remains to
23 be identified. Interestingly, concurrent increases in A_T and C_T were most rapid from May to July.
24 Changes at 50 m were slower compared to 1 m. It seems therefore likely that changes in coastal
25 A_T cycling via a shallow coastal process gave rise to these observations. This study exemplifies
26 the importance of understanding coastal ocean acidification through localized biogeochemical
27 cycling that extends beyond simple air-sea gas exchange dynamics, in order to make relevant
28 predictions about future coastal ocean change and ecosystem function.
29
30 **Keywords** – global ocean change, ocean acidification, time-series, pH, alkalinity, dissolved
31 inorganic carbon, pCO_2 , Mediterranean Sea, near-shore



32 1. Introduction

33 Maintaining time-series of oceanographic data is essential for understanding
34 anthropogenic change in the ocean (Tanhua et al., 2013). On land, fossil fuel burning, cement
35 production and land use changes have contributed ~580 Gt carbon to the atmosphere during the
36 period 1750-2013 (Le Quéré et al., 2015). An estimated 29 % of this anthropogenic carbon is
37 absorbed by the ocean in the form of carbon dioxide (CO₂; Le Quéré et al., 2015), and causing
38 global changes to the ocean carbonate system. Absorption of CO₂ by seawater produces carbonic
39 acid, which decreases seawater pH, and is of great concern for biological processes and marine
40 ecosystems (Doney et al., 2009; Gattuso and Hansson, 2011; Pörtner et al., 2014). Since the
41 preindustrial era, global mean ocean pH has declined by 0.1 (Rhein et al., 2013). Due to the
42 declining trend of ocean pH with increasing anthropogenic CO₂, the process is termed ‘ocean
43 acidification’, but this expression represents a suite of chemical changes, including increases in
44 total dissolved inorganic carbon (C_T) and partial pressure of CO₂ (pCO₂) and decrease in calcium
45 carbonate saturation states (Ω , aragonite and calcite; Dickson, 2010). Rates of ocean
46 acidification differ by region and range from -0.0013 units pH yr⁻¹ (South Pacific) to -0.0026
47 units pH yr⁻¹ (Irminger Sea, North Atlantic) and are reviewed in Bates et al. (2014). Such time-
48 series remain spatially limited, especially in coastal regions which provide valuable ecosystem
49 services (Barbier et al., 2011; Costanza et al., 1997) and are under high anthropogenic impact
50 (Halpern et al., 2008). Here, we present the first ocean acidification time-series at weekly
51 frequency for a coastal site in the Mediterranean Sea.

52 Compared to the global ocean, marginal seas serve a critical role in anthropogenic CO₂
53 storage via enhanced CO₂ uptake and export to the ocean interior (Lee et al., 2011). As a
54 marginal sea, the Mediterranean Sea has a naturally high capacity to absorb but also buffer



55 anthropogenic CO₂ (Álvarez et al., 2014; Palmiéri et al., 2015). This is primarily due to the high
56 total alkalinity (A_T) of Mediterranean waters and overturning circulation (Lee et al., 2011;
57 Palmiéri et al., 2015; Schneider et al., 2010). In the Mediterranean Sea, the salinity- A_T
58 relationship is driven by the addition of river discharge and Black Sea input, which are generally
59 both high in A_T (Copin-Montégut, 1993; Schneider et al., 2007). Combined with evaporation,
60 this results in higher A_T and salinity in the Mediterranean Sea compared to the Atlantic
61 Mediterranean source water (Jiang et al., 2014). On average, Mediterranean Sea A_T is 10 %
62 higher than in the global ocean (Palmiéri et al., 2015). The surface ocean acidification rate,
63 estimated at ΔpH_T of -0.08 since 1800, is comparable to that of the global ocean despite a 10%
64 greater anthropogenic carbon inventory (Palmiéri et al., 2015). Due to its important role in
65 carbon sequestration and ecological sensitivity to global ocean change with economic impacts
66 (Lacoue-Labarthe et al., 2016), the Mediterranean Sea is a key location for time-series
67 measurements.

68 Over the last few years, numerous studies have estimated ocean acidification rates across
69 the Mediterranean Sea (Table 1). Together, these studies cover various study periods with a
70 range of techniques yielding different results. For example, estimates of change in pH of bottom
71 waters since the preindustrial era range between -0.005 to -0.06 (Palmiéri et al., 2015) and as
72 much as -0.14 (Touratier and Goyet, 2011). Techniques for estimating ocean acidification in the
73 Mediterranean Sea thus far include: (1) hind-casting, using high-resolution regional circulation
74 models (Palmiéri et al., 2015), the TrOCA approach as applied to cruise-based profile data
75 (Krasakopoulou et al., 2011; Touratier and Goyet, 2011; Touratier et al., 2016) and others
76 (Howes et al., 2015), (2) partially reconstructed time-series (Marcellin Yao et al., 2016), (3)
77 comparative study periods (Luchetta et al., 2010; Meier et al., 2014), and (4) sensor-based



78 observations over a short study period (Flecha et al., 2015). Ocean acidification time-series of
79 consistent sampling over many years are lacking for the Mediterranean Sea (The MerMex
80 Group, 2011), especially along the coast where river discharge influences the carbonate system
81 (Ingrosso et al., 2016).

82 Compared to the open ocean, shallow coastal sites exhibit natural variability in carbonate
83 chemistry over annual timeframes (Hofmann et al., 2011; Kapsenberg and Hofmann, 2016;
84 Kapsenberg et al., 2015), complicating the detection and relevance of open ocean acidification in
85 isolation of other processes (Duarte et al., 2013). Variability stems from both physical (e.g.,
86 upwelling, river discharge; Feely et al., 2008; Vargas et al., 2016) and biological processes (e.g.,
87 primary production, respiration, net calcification). Within watersheds, coastal carbonate
88 chemistry is affected by eutrophication (Borges and Gypens, 2010; Cai et al., 2011),
89 groundwater supply (Cai et al., 2003), and land use and rain influence on river alkalinity
90 (Raymond and Cole, 2003; Stets et al., 2014). Over longer periods, pH can also be influenced by
91 atmospheric deposition (Omstedt et al., 2015). Introduction of nutrient-rich upwelled- or fresh-
92 water masses influences biological processes and carbonate chemistry at higher frequencies.

93 Through primary production and respiration, coastal ecosystems produce pH fluctuations over
94 hours (e.g., seagrass, kelp) to months (e.g., phytoplankton blooms; Kapsenberg and Hofmann,
95 2016). Due to existing pH variability in coastal seas, it is necessary to quantify high-frequency
96 pH variability in order to interpret the pH changes inferred from lower-frequency sampling, at
97 time-series stations.

98 In this study, we present the first complete time-series data quantifying the present-day
99 ocean acidification rate for a coastal site in the Mediterranean Sea, based on weekly
100 measurements of A_T and C_T sampled from 2007 through 2015. For a subset of this time-series,



101 we documented pH variability using a SeaFET™ Ocean pH Sensor in order to assess hourly pH
102 variability. For comparison and consistency with other ocean acidification time-series around the
103 world, we report rates of change based on anomalies (Bates et al., 2014).

104

105 **2. Materials and methods**

106 **2.1. Site description**

107 A carbonate chemistry time-series was initiated in 2007 in the NW Mediterranean Sea at
108 the entrance of the Bay of Villefranche-sur-Mer, France (Fig. 1): Point B station (43.686° N,
109 7.316° E, 85 m bottom depth). A second site, Environment Observable Littoral buoy (EOL,
110 43.682° N, 7.319° E, 80 m bottom depth), was used for pH sensor deployment starting in 2014.
111 These two sites are 435 m apart. The site Point B is an historical sampling point, since 1957,
112 regarding several oceanographic parameters. A full site description and research history has been
113 detailed by De Carlo et al. (2013). Briefly, the Bay is a narrow north-south facing inlet with steep
114 bathymetry and estimated volume of 310 million m³. The surrounding region is predominately
115 composed of limestone with a series of shallow, submarine groundwater karst springs (Gilli,
116 1995). The North current, a major and structuring counter-clockwise current in the Ligurian Sea,
117 can sometimes flow close to Point B. The Bay can also be, on occasion, influenced by local
118 countercurrents. Both of these hydrodynamics movements have signatures of riverine discharge,
119 which for the Mediterranean Sea are generally high in A_T (Copin-Montégut, 1993; Schneider et
120 al., 2007). For example, the Paillon River, 4 km west of Point B and whose plume on occasion
121 reaches into the Bay (L. Mousseau, pers. obs.), was sampled on 18 Aug 2014 and had a A_T of
122 $1585 \pm 0.1 \mu\text{mol kg}^{-1}$ ($N=2$, J.-P. Gattuso, unpubl.). Due to low primary productivity, seasonal



123 warming drives the main annual variability in carbonate chemistry at this location (De Carlo et
124 al., 2013).

125

126 **2.2. Point B data collection, processing, and analysis**

127 To document long-term changes in ocean carbonate chemistry at Point B, seawater was
128 sampled weekly starting in January 2007. Samples were collected at 1 and 50 m, using a 12-L
129 Niskin bottle at 9:00 local time. Seawater was transferred from the Niskin bottle to 500 mL
130 borosilicate glass bottles and fixed within an hour via addition of saturated mercuric chloride for
131 preservation of carbonate parameters, following recommendations by Dickson et al. (2007).
132 Duplicate samples were collected for each depth. For each sampling event, CTD casts were
133 performed either with a Seabird 25 or Seabird 25+ profiler whose sensors are calibrated at least
134 every two years. Accuracy of conductivity (SBE4 sensor) and temperature (SBE3 sensor)
135 measurements from CTD casts were 0.0003 S m^{-1} and 0.001°C , respectively.

136 Within 6 months of collection, bottle samples were analyzed for C_T and A_T via
137 potentiometric titration following methods described by Edmond (1970) and DOE (1994), by
138 *Service National d'Analyse des Paramètres Océaniques du CO_2* , at the Université Pierre et Marie
139 Curie in Paris, France. Precision of C_T and A_T was less than $3 \mu\text{mol kg}^{-1}$, and the average
140 accuracy was 2.6 and $3 \mu\text{mol kg}^{-1}$, as compared with seawater certified reference material
141 (CRM) provided by A. Dickson (Scripps Institution of Oceanography). Only obvious outliers
142 were omitted from the analyses: three C_T values at 1 m ($> 2300 \mu\text{mol kg}^{-1}$), one A_T value at 1 m
143 ($> 2900 \mu\text{mol kg}^{-1}$), one A_T value at 50 m ($< 2500 \mu\text{mol kg}^{-1}$). The C_T and A_T measurements on
144 replicates bottle samples were averaged for analyses.



Calculations of the carbonate system parameters were performed using the R package seacarb version 3.1 with C_T , A_T , temperature and salinity as inputs (Gattuso et al., 2016). Total concentrations of silicate (SiOH_4) and phosphate (PO_4^{3-}) were used when available from Point B (L. Mousseau, unpubl.). Detection limits for nutrients were $0.05 \mu\text{M}$ for SiOH_4 and 0.003 to $0.006 \mu\text{M}$ for PO_4^{3-} ; relative precision of these analyses is 5-10 % (Aminot and K  rouel, 2007). Total boron concentration was calculated from salinity using the global ratio determined by Lee et al. (2010). The following constants were used: K_1 and K_2 from Lueker et al. (2000), K_f from Perez and Fraga (1987), and K_s from Dickson (1990). Reported measured parameters are temperature, salinity, A_T , and C_T , and derived parameters are pH_T (total hydrogen ion scale), pH_T normalized to 25°C (pH_{T25}), pCO_2 , and aragonite (Ω_a) and calcite (Ω_c) saturation states. Salinity-normalized changes in A_T (nA_T) and C_T (nC_T) were calculated by dividing by salinity and multiplying by 38. Except for pH_{T25} , all parameters are reported at *in situ* temperatures.

The average uncertainties of the derived carbonate parameters were calculated according to the Gaussian method (Dickson and Riley, 1978) implemented in the “errors” function of the R package seacarb 3.1 (Gattuso et al., 2016). The uncertainties are $\pm 2.7 \times 10^{-10} \text{ mol H}^+$ (about 0.015 units pH_T), $\pm 15 \mu\text{atm pCO}_2$, and ± 0.1 unit of the aragonite and calcite saturation states.

To quantify interannual changes in carbonate parameters, the data were detrended for seasonality by subtracting monthly means from the time-series following methods in Bates et al. (2014). The resulting anomalies were analyzed using a linear regression. All analyses were performed in R (R Core Team, 2016).

165

2.3. Deconvolution of pH_T and pCO_2



167 To identify proportional contributions of various drivers to ocean acidification trends at
 168 Point B, deconvolution of time-series pH_T and pCO_2 was performed following methods from
 169 García-Ibáñez et al. (2016) for observations at 1 m. The equation is described below for pH_T ,
 170 where changes in pH_T are driven by changes in temperature (T), salinity (S), A_T , and C_T , over
 171 time (t), according to the following model:

$$172 \quad \frac{d\text{pH}_T}{dt} = \frac{\partial \text{pH}_T}{\partial T} \frac{dT}{dt} + \frac{\partial \text{pH}_T}{\partial S} \frac{dS}{dt} + \frac{\partial \text{pH}_T}{\partial A_T} \frac{dA_T}{dt} + \frac{\partial \text{pH}_T}{\partial C_T} \frac{dC_T}{dt} \quad (1)$$

173 Here, $\frac{\partial \text{pH}_T}{\partial \text{var}} \frac{d\text{var}}{dt}$ represents the slope contribution of changing var to the estimated change
 174 in pH_T , where var is either T , S , A_T , or C_T . The rate of pH change due to var ($\frac{\partial \text{pH}_T}{\partial \text{var}} \frac{d\text{var}}{dt}$) was
 175 estimated by calculating pH_T using the true observations of var but monthly mean values of the
 176 other three variables, and regressing it to time. The calculation was repeated for pCO_2 ($\frac{d\text{pCO}_2}{dt}$) in
 177 order to compare the rate of increase with that of atmospheric CO_2 .

178 As a sub-component of $\frac{\partial \text{pCO}_2}{\partial C_T} \frac{dC_T}{dt}$, the rate of anthropogenic CO_2 increase was estimated
 179 from atmospheric CO_2 concentrations nearest to Point B (Plateau Rosa, Italy, courtesy of the
 180 World Data Center for Greenhouse Gases, <http://ds.data.jma.go.jp/gmd/wdcgg/>). For these data,
 181 missing daily values were linearly interpolated. A linear regression was performed where the
 182 slope represents the rate of CO_2 increase in the atmosphere. Finally, to help identify different
 183 processes that might have contributed to the observed trends, linear regressions were performed
 184 on change in A_T and C_T by month and on the salinity- A_T relationship by year.

185

186 **2.4. SeaFET data collection, processing, and analysis**

187 To capture pH variability at higher-than-weekly sampling frequencies, a SeaFET™
 188 Ocean pH sensor (Satlantic) was deployed on the EOL buoy (435 m from the Point B sampling



189 site) starting in June 2014, at 2 m depth. Autonomous sampling was hourly and deployment
190 periods ranged between 1 and 3 months. Field calibration samples for pH were collected weekly,
191 using a Niskin bottle next to SeaFET within 15 min of measurement. This sampling scheme was
192 sufficient for this site as there is no large high-frequency pH variability. Unlike Point B
193 sampling, SeaFET calibration samples were processed for pH using the spectrophotometric
194 method (Dickson et al., 2007) with purified m-cresol purple (purchased from the Byrne lab,
195 University of South Florida). *In situ* temperature, salinity, and A_T measured at Point B, within 30
196 min of the SeaFET sampling, were used to calculate *in situ* pH_T of the calibration samples.
197 SeaFET voltage was converted to pH_T using the respective calibration samples for each
198 deployment period, following the methods and code described in Bresnahan et al. (2014) but
199 adapted for use in R.

200 The estimated standard uncertainty in SeaFET pH_T is ± 0.01 and was calculated as the
201 square root of the sum of each error squared. The sources of errors are: measurement error of
202 spectrophotometric pH (± 0.004 , $N=68$ mean SD of 5 replicate measurements per calibration
203 sample for samples collected between 16 July 2014 and 3 May 2016), spatio-temporal mismatch
204 sampling at EOL (± 0.007 , mean offset of pH_T of the calibration samples from calibrated time-
205 series), and variability in purified m-cresol dye batch accuracy as compared to Tris buffer CRM
206 pH (± 0.006 , mean offset of pH_T of the spectrophotometric measurement of Tris buffer from the
207 CRM value).

208

209 3. Results

210 3.1. Time-series trends



At Point B from January 2007 to December 2015, >400 samples were collected for carbonate chemistry at both 1 and 50 m. Anomaly trends detected at 1 m (Fig. 2, Table 2) were also significant at 50 m (Table S1, Fig. S1), with the exception that salinity increased at 50 m (0.0063 ± 0.0020 units yr^{-1} , $P = 0.002$). Changes in carbonate chemistry were faster at 1 m compared to 50 m, with the exception of salinity and temperature. The warming rate at 50 m was 22 % greater compared to 1 m, mostly due to increasing summer temperatures since 2007. Analyses for 50 m are available (Table S1, Fig. S1) but here, we focus on results from 1 m unless explicitly specified otherwise (Fig. 2, Table 2). For time-series anomalies at 1 m, carbonate chemistry anomalies were significant for pH_T (-0.0028 ± 0.0003 units yr^{-1} , $N=412$), A_T ($+2.08 \pm 0.19$ $\mu\text{mol kg}^{-1} \text{yr}^{-1}$, $N=417$), C_T ($+2.97 \pm 0.20$ $\mu\text{mol kg}^{-1} \text{yr}^{-1}$, $N=416$), pCO_2 ($+3.53 \pm 0.39$ $\mu\text{atm yr}^{-1}$, $N=412$), and Ω_a (-0.0064 ± 0.0015 units yr^{-1} , $N=412$). At the same time, temperature anomaly increased ($+0.072 \pm 0.022$ $^{\circ}\text{C yr}^{-1}$, $N=413$), but this significance was lost with the exclusion of the year 2015 (analysis not shown; changes in other carbonate chemistry parameters remained significant). No significant change in the salinity anomaly was detected ($P=0.702$, $N=417$).

Strong seasonal cycles in carbonate parameters were present at Point B at 1 m (Fig. 3). Calculated monthly means (2007-2015) are described briefly and listed in Table S2. Mean temperature range was 11.2 $^{\circ}\text{C}$ with a maximum at 24.77 ± 1.35 $^{\circ}\text{C}$ in August and minimum of 13.58 ± 0.41 $^{\circ}\text{C}$ in February. The range in A_T was $+19$ $\mu\text{mol kg}^{-1}$ from June to September. The C_T range was 33 $\mu\text{mol kg}^{-1}$ with a peak in late winter and minimum values in August and October. The opposing seasonal cycles of temperature and C_T , due to summer warming coinciding with the period of peak primary productivity (De Carlo et al. 2013), resulted in an annual pH_T range of 0.12 with peak pH in late winter (8.14 ± 0.01 , February and March) and



234 minimum pH in summer (8.02 ± 0.03 , July and August). The corresponding $p\text{CO}_2$ range was
 235 $+128 \mu\text{atm}$ from February to August. The monthly means were subtracted from the time-series to
 236 calculate anomalies.

237

238 3.2. Deconvolution of pH_T and pCO_2

239 Deconvolutions of pH and pCO_2 are presented in Table 3. For both $\frac{d\text{pH}_T}{dt}$ and $\frac{d\text{pCO}_2}{dt}$,
 240 temperature and salinity had contributing (approx. 35% and 5% respectively) but non-significant
 241 effects. The 2007-2015 observation period is likely not long enough to detect significance of the
 242 relatively large temperature contribution to $\frac{d\text{pH}_T}{dt}$. The predominant driver of $\frac{d\text{pH}_T}{dt}$ and $\frac{d\text{pCO}_2}{dt}$ was
 243 the increase in C_T ($P \ll 0.001$), 56 and 60 % of which was compensated for by a significant
 244 increase in A_T ($P = 0.002$), respectively.

245 Atmospheric CO_2 at Plateau Rosa increased by $2.02 \pm 0.03 \text{ ppm yr}^{-1}$ ($F_{1,3285} = 5852.43$, P
 246 $\ll 0.001$, R^2 0.64), with an anomaly rate of $2.08 \pm 0.01 \text{ ppm yr}^{-1}$ ($F_{1,3285} = 46649.79$, $P \ll 0.001$,
 247 R^2 0.93) during the study period 2007-2015, and represents the anthropogenic CO_2 forcing of
 248 ocean acidification. Considering the error associated with deconvolution of pCO_2 and assuming
 249 air-sea CO_2 equilibrium, atmospheric CO_2 increase represents 31 to 44 % of the total C_T
 250 contribution ($\frac{\partial \text{pCO}_2}{\partial C_T} \frac{dC_T}{dt}$) to $\frac{d\text{pCO}_2}{dt}$. This leaves 56 to 69 % of the total C_T contribution to pCO_2
 251 unaccounted for. As A_T is not influenced by addition of anthropogenic CO_2 to seawater, the next
 252 question was, thus, whether or not the changes in A_T and C_T were process-linked. Regressions of
 253 annual monthly observations of A_T and C_T revealed similar seasonal cycles for both parameters
 254 (Fig. 4). The fastest increases occurred simultaneously for both parameters, which peaked in
 255 June. The smallest changes occurred in January. Significant increases in A_T occurred from May



256 to July. Significant increases in C_T occurred from February-August and in November. Results of
 257 regression analyses on monthly changes are listed in Table S3.

258

259 **3.3. Salinity and A_T relationships**

260 Given the coastal locale and the increase in A_T , it is interesting to look at the interannual
 261 variability of the relationship between salinity and A_T (Fig. 5). With the exception of 2007,
 262 salinity was a poor predictor of A_T . The R^2 value for each annual salinity- A_T linear regression
 263 ranged from 0.00 (in 2013) to 0.87 (in 2007) with y-intercepts (A_{T0} , the freshwater end-member
 264 alkalinity) ranging between $-176 \mu\text{mol kg}^{-1}$ (in 2007) and $2586 \mu\text{mol kg}^{-1}$ (in 2013). The
 265 interannual variability of the salinity- A_T relationship was driven by the variability in A_T observed
 266 at salinity < 38.0 that was present from November through July.

267 The salinity cycle (monthly means) at Point B was small and ranged from 37.64 ± 0.26 to
 268 38.21 ± 0.11 from May to September, following freshwater input in winter and spring and
 269 evaporation throughout summer and fall (Fig. 3). Highest (> 38.0) and most stable salinity
 270 observations were made in August through October, which coincided with the period of
 271 maximum A_T (2562 and $2561 \pm 9 \mu\text{mol kg}^{-1}$ in September and October, respectively). Minimum
 272 A_T ($2543 \pm 14 \mu\text{mol kg}^{-1}$) was observed in June. The reported salinity- A_T relationship is thus
 273 generated from monthly means ($R^2 = 0.74$) where A_T units are $\mu\text{mol kg}^{-1}$ and error terms are SE:

$$274 \quad A_T = 1554.9(\pm 185.9) + 26.3(\pm 4.9) \times S \quad (2)$$

275

276 **3.4. High-frequency pH data**

277 To verify the weekly sampling scheme at Point B, a total of 11 SeaFET deployments
 278 were conducted from June 2014 to April 2016, averaging 58 ± 25 days and 5 ± 2 calibration



279 samples per deployment (Fig. 6). Only 5 % of the data was removed during quality control, due
 280 to biofouling in one deployment and a drained battery in another, yielding 610 days of good data.
 281 The mean offset between calibration samples and the calibrated SeaFET pH time-series was \pm
 282 0.007 units pH_T , indicating a high-quality pH dataset (Fig. 6c). Sensor data corroborated the
 283 seasonal pH and temperature cycle observed at Point B. Event-scale effects (e.g., pH_T change \geq
 284 0.1 for days to weeks, *sensu* Kapsenberg and Hofmann 2016) were absent at this site suggesting
 285 that weekly sampling was sufficient to describe seasonal and interannual changes in carbonate
 286 chemistry at Point B. Magnitude of diel pH_T variability was small (the 2.5th to 97.5th percentiles
 287 ranged between 0.01 and 0.05 units pH_T) and unrelated to seasonal warming or the concentration
 288 of Chlorophyll-a (Fig. S2).

289

290 **4. Discussion**

291 **4.1. Observed changes in carbonate chemistry**

292 Anthropogenic forcing of CO_2 in seawater necessitates long-term monitoring, in order to
 293 understand and study future ecological change in the coastal environment. In the coastal NW
 294 Mediterranean Sea at EOL, near Point B, high-frequency pH data validated that weekly, morning
 295 sampling at Point B was sufficient to capture water mass changes that were independent from
 296 benthic, diel pH variability in seagrass beds inside the bay (Cox et al., 2016). Based on weekly
 297 samples collected from 2007 through 2015, we detected anomaly changes in pH_T of $-0.0028 \pm$
 298 $0.0003 \text{ units yr}^{-1}$ and an increase in A_T of $2.08 \pm 0.19 \mu\text{mol kg}^{-1} \text{ yr}^{-1}$. The corresponding pCO_2
 299 increase of $3.53 \pm 0.39 \mu\text{atm yr}^{-1}$ is 70 % greater than the atmospheric CO_2 anomaly of $+2.08 \pm$
 300 0.01 ppm yr^{-1} at Plateau Rosa, Italy. While Point B is a weak sink for CO_2 (De Carlo et al.,
 301 2013), the increase in anthropogenic CO_2 in the atmosphere alone does not account for all of the



302 observed changes in carbonate chemistry at this location. The increase in C_T was the dominant
 303 driver of pH_T change over the study period. An increase in A_T partially buffered this C_T increase
 304 such that the observed acidification rate is only slightly larger than those reported at other ocean
 305 time-series sites (-0.0026 to -0.0013 units yr^{-1} , Bates et al., 2014). Warming contributed to
 306 approximately 35 % of the pH decline, which agrees well with the 30 % approximation at
 307 DYFAMED, an open-sea site about 50 km offshore from Point B (Marcellin Yao et al., 2016). A
 308 longer time-series of temperature paired with pH observations will be necessary to definitively
 309 characterize this relationship. While anomalously warm summer temperatures in 2015 drove the
 310 warming trends in this study, this region has warmed steadily since 1980 (Parravicini et al.,
 311 2015) and this is observable at 50 m as well (Fig. S1). Finally, the co-evolution of A_T and C_T
 312 changes suggests that changes in C_T are also due to increasing A_T , which includes bicarbonate
 313 (HCO_3^-) and carbonate (CO_3^{2-}) ions (Fig. 4). We assess the spatial extent of these trends and
 314 discuss potential drivers.

315 Fastest rate increases in A_T and C_T occurred from May through July; the period of
 316 seasonal transition (Fig. 3). At this time, various biophysical processes force the seasonality of
 317 the carbonate system. In the NW Mediterranean, the main processes governing seasonal
 318 variability in A_T are evaporation increasing A_T in summer (i.e., June through September at Point
 319 B) and, to a lesser extent, phytoplankton uptake of nitrate (NO_3^-) and phosphate (PO_4^{3-})
 320 increasing A_T from January through March (Cossarini et al., 2015). During the transition of these
 321 processes, salinity decreases to a minimum in May, reflecting freshwater input that dilutes A_T to
 322 minimum values at the start of summer. For C_T , peak values occur in winter when the water
 323 column is fully mixed (L. Mousseau, unpubl.). Mixing occurs down to more than 2000 m depth a
 324 few 10s of km from the study site and C_T is up to $100 \mu mol kg^{-1}$ higher in deep waters (Copin-



325 Montégut and Bégovic, 2002). Following winter, C_T declines due to a combination of
 326 phytoplankton bloom carbon uptake and freshwater dilution, until the onset of summer
 327 stratification. Summer warming leads to pCO_2 off-gassing to the atmosphere (De Carlo et al.,
 328 2013), thereby further decreasing C_T . The increases in A_T and C_T from 2007 through 2015 were
 329 more pronounced at 1 m compared to 50 m, suggesting that the driving processes dominate at the
 330 surface.

331 To estimate if the observed trends from Point B were also occurring offshore, we turn to
 332 time-series station DYFAMED, approx. 50 km from Point B (Fig. 1). The acidification rate at
 333 DYFAMED was estimated at -0.003 ± 0.001 units $pH\ yr^{-1}$ from 1995 to 2011 (Marcellin Yao et
 334 al., 2016), however, the uncertainty is large and makes the comparison with Point B unreliable.
 335 A_T at the DYFAMED did not change significantly during the period 2007-2014 ($F_{1,51} 3.204$, $P =$
 336 0.0794 , $R^2 0.08$, data from L. Coppola). This suggests that the processes driving changes in A_T
 337 and C_T at Point B, in addition to being shallow, dissipate offshore.

338 Another coastal area of the Mediterranean Sea, in the North Adriatic Sea, exhibited
 339 similar changes in carbonate chemistry as those observed at Point B. By comparing cruise data
 340 between the winters of 1983 and 2008, Luchetta et al. (2010) determined an acidification rate of
 341 $pH_T -0.0025$ units yr^{-1} and an increase in A_T of $2.98\ \mu mol\ kg^{-1}\ yr^{-1}$ at depths < 75 m. As coastal
 342 sites, Point B and the Gulf of Trieste in the N Adriatic Sea show a strong positive A_{T0} (i.e.,
 343 freshwater end-member alkalinity) in surface waters (Cantoni et al., 2012; Ingrosso et al., 2016).
 344 In contrast to Point B, the N Adriatic Sea exhibits a negative salinity- A_T relationship and faster
 345 A_T increase (Luchetta et al., 2010), suggesting that rivers may play a stronger role in A_T trends
 346 there compared to Point B. Rivers are significant sources of C_T to the Gulf of Trieste, making up
 347 3-16 % of C_T in 2007 (Tamše et al., 2015). The authors note that 2007 was a year of record low



348 river discharge. Notably, this is the only year at Point B for which A_{T0} was negative and salinity-
349 A_T relationship was highly correlated, also indicating that 2007 was a year of low freshwater
350 input.

351 The correlates between Point B and N Adriatic Sea suggest a common driver of changes
352 in ocean carbonate chemistry at these two sites (possibly linked via shared watersheds of the
353 Alps), and these independent studies may be symptomatic of changes occurring across wider
354 coastal areas of the Mediterranean Sea. Monitoring efforts of carbonate chemistry in the eastern
355 Mediterranean Sea would offer an important contrast, as pH of eastern waters is expected to be
356 more sensitive to atmospheric CO_2 addition due to their ability to absorb more anthropogenic
357 CO_2 than either the western Mediterranean or Atlantic waters (Álvarez et al., 2014).

358

359 **4.2. Potential drivers of changes in carbonate chemistry**

360 Coastal ocean acidification rates vary greatly across different regions. In the NW Pacific
361 coast, rapid acidification of surface waters ($\Delta\text{pH}_T -0.058 \text{ units yr}^{-1}$) at Tatoosh Island has been
362 documented in the absence of concomitant changes in known drivers of local pH variability (e.g.,
363 upwelling, eutrophication, and more; Wootton and Pfister, 2012; Wootton et al., 2008). Further
364 inshore, in the Hood Canal sub-basin of the Puget Sound, only 24-49 % of the estimated pH
365 decline from pre-industrial values could be attributed to anthropogenic CO_2 (Feely et al., 2010).
366 The excess decrease in pH was attributed to increased remineralization (Feely et al., 2010).
367 Acidification rates documented along the North Sea Dutch coastline and inlets were highly
368 variable in space, with some exceeding the expected anthropogenic CO_2 rate by an order of
369 magnitude while others exhibited an increase in pH (Provoost et al., 2010). These sites
370 experience much greater sub-annual variability than Point B. As a coastal ocean acidification



371 monitoring site, the relative simplicity of the Point B pH variability regime may therefore
372 provide an opportunity to further investigate the underlying drivers of rapid coastal ocean
373 acidification in the absence of additional noise from otherwise overshadowing processes.

374 While ocean acidification is often described in pH_T , at Point B, pH is largely a product of
375 underlying changes in A_T and C_T . Sediment dissolution is unlikely to contribute to the observed
376 increase in A_T as both aragonite and calcite were supersaturated throughout the study period;
377 conditions that are not conducive to large rates of dissolution. An overall reduction in calcium
378 carbonate (CaCO_3) precipitation rates is unlikely, as the dominant ecosystem in the Bay of
379 Villefranche-sur-Mer is seagrass meadows, which harbor relatively few calcifying organisms and
380 there has been no obvious changes in the abundance of calcifiers (J.-P. Gattuso, pers. obs.). All
381 the same, the sheer volume of the Bay would likely dilute any signature of changes in
382 calcification or dissolution of sediment or organisms. For the Mediterranean Sea, in general, the
383 influence of biogenic CaCO_3 on A_T is small compared to the influence of river and Black Sea A_T
384 (Copin-Montégut, 1993). Increased input from the eastern Mediterranean Sea waters could
385 increase A_T , but this is questionable, because while eastern waters are higher in A_T , salinity and
386 pH are also greater and C_T is a little lower, compared to the western waters (Álvarez et al., 2014;
387 Touratier and Goyet, 2011).

388 For C_T , there are a few additional drivers to discuss. First, anthropogenic CO_2 partly
389 accounted for the observed increase in C_T , independent of A_T . In summer, pCO_2 peaks and Point
390 B becomes a weak source of CO_2 to the atmosphere (De Carlo et al., 2013). As atmospheric CO_2
391 is increasing it may contribute to summertime C_T retention, but this effect is likely too small for
392 the observed increases in C_T . Second, increased remineralization rates could contribute to C_T ,
393 particularly in summer when the water is warmer, hence promoting respiration, and highly



394 stratified and isolated from C_T -rich bottom waters. However, this would require either (1) a
395 change in substrate (local or regional) for remineralization but no clear trend in particulate
396 organic carbon was observable in the study period (L. Mousseau, unpubl.), or (2) a change in net
397 community metabolism, but warming trends up through 2014 were not significant. Lastly,
398 Chlorophyll-a biomass, a proxy of primary production, has decreased since 1995 and blooms
399 have shifted towards earlier dates in the year, at Point B (Irisson et al., 2012). Both of these
400 processes could influence the increasing rate of C_T in summer but would not account for the
401 increase in A_T .

402 The lack of salinity change excludes additional processes as drivers of carbonate
403 chemistry change at Point B. For example, increased summertime evaporation (concentration
404 effect) and reduced riverine discharge (decreased dilution effect) would both be expected to
405 cause an increase in salinity, which was not observed and so these are not suspected drivers.
406 However, changes in the content of freshwater sources is conceivable.

407 The observed changes in A_T and C_T could be achieved via augmented limestone
408 weathering increasing the A_T input from land to the sea via rivers and submarine groundwater
409 springs. Increased A_T in freshwater was documented in North American rivers (Raymond and
410 Cole, 2003; Stets et al., 2014) and groundwater (Macpherson et al., 2008). For the Mediterranean
411 Sea, rivers are a significant source of both A_T and C_T (Copin-Montégut, 1993). Riverine
412 contributions of A_T originate from erosion and are correlated with bedrock composition (e.g.,
413 McGrath et al., 2016). The annual variability in salinity- A_T relationships at Point B does suggest
414 influence of river discharge, as has been observed elsewhere in the Mediterranean Sea (Cantoni
415 et al., 2012; Turk et al., 2010). Signatures of limestone erosion can be observed in the A_T of local
416 rivers near Point B (Var, Paillon, and Roya) and range between 1000 to 2000 $\mu\text{mol kg}^{-1}$ (data



417 from *Agence de l'Eau Rhône-Méditerranée-Corse*, <http://sierm.eaurmc.fr>). Even a rainwater
418 outfall at the entrance of Port de la Darse harbor, inside the Bay of Villefranche-sur-Mer, had an
419 A_T of $607 \pm 5 \mu\text{mol kg}^{-1}$ ($N=2$) following a precipitation event. However, local precipitation was
420 not correlated with salinity or A_T , suggesting that rain runoff is not a driving factor of Point B
421 carbonate chemistry (Fig. S3). Lastly, submarine groundwater springs are a significant source of
422 nutrients, A_T , and C_T to the ocean (Cai et al., 2003; Slomp and Van Cappellen, 2004). Submarine
423 springs have been identified along the Point B coastline (Gilli, 1995), but their carbonate
424 chemistry contributions are currently unknown.

425 Rivers as a potential driver of A_T and C_T trends at Point B could be achieved if the A_T
426 content discharged by rivers was changing, as was proposed for the Adriatic Sea (Luchetta et al.,
427 2010). For example, terrestrial organic matter cycling influences riverine C_T (Vargas et al.,
428 2016), so changes in soil respiration could be expected to change A_T of rivers. Increasing river A_T
429 has been documented in North America and occurs via a number of processes including: (1) the
430 interplay of rainfall and land-use (Raymond and Cole, 2003), (2) anthropogenic limestone
431 addition (a.k.a., liming) used to enhance agriculture soil pH (Oh and Raymond, 2006; Stets et al.,
432 2014) and freshwater pH (Clair and Hindar, 2005), and (3) potentially indirect effects of
433 anthropogenic CO_2 on groundwater CO_2 -acidification and weathering (Macpherson et al., 2008).
434 Such, and other, processes are hypothesized to have driven A_T changes in the Baltic Sea (Müller
435 et al., 2016). There, A_T increased at a rate of $+3.4 \mu\text{mol kg}^{-1} \text{ yr}^{-1}$ during the period 1995 to 2014
436 (mean salinity = 7). In contrast to Point B, where salinity is about 38, the increase in Baltic Sea
437 A_T was not noticeable at salinity > 30 (Müller et al., 2016). This contrast clouds the perspective
438 that changes in A_T of freshwater sources can influence A_T at Point B, and a more in depth study
439 will be necessary to address this.



440 Given the discussion above, the simplest plausible mechanisms causing changes in
441 carbonate chemistry at Point B would be through (1) increasing anthropogenic atmospheric CO₂,
442 and (2) increasing A_T of freshwater sources (i.e., rivers, groundwater). Freshwater has a shallow
443 and coastal influence and is also dominant in the N Adriatic Sea, which exhibits similar trends as
444 those observed at Point B. If so, there is a lag effect, as freshwater influence peaks in May but A_T
445 and C_T increased fastest from May through July. Consequently, this hypothesis needs further
446 investigation. The influence of a coastal boundary processes influencing seawater A_T and C_T
447 presents a potentially major difference between coastal and offshore ocean acidification rates.
448 Until the source of A_T increase is properly identified, use of this observation (e.g., in modeling)
449 should be implemented with great caution.

450

451 **5. Conclusion**

452 Predictions of coastal ocean acidification remain challenging due the complexity of
453 biogeochemical processes occurring at the ocean-land boundary and the lack of long-term
454 monitoring. At the Point B coastal monitoring station in the NW Mediterranean Sea, the ocean
455 acidification trend is greater than expected from assuming atmospheric equilibrium. We
456 postulate that the enhanced acidification trend could stem from changes in freshwater inputs
457 from land which are also the source of interannual variability in A_T at this site. This study
458 highlights the importance of considering other anthropogenic influences in the greater land-sea
459 region that may (1) contribute to coastal biogeochemical cycles (*sensu* Duarte et al. 2013) and
460 (2) inform projections of anthropogenic change in near-shore waters and experimental design in
461 global ocean change biology.

462



463 **Data availability** – Time-series data from Point B are available at Pangaea[®] (doi:
464 10.1594/PANGAEA.727120)

465

466 **Author contribution** – JPG initiated the study, LM supervised data collection, SA performed
467 SeaFET deployments and calibration, JPG and LK designed and JPG conducted statistical
468 analyses, and LK prepared the manuscript with contributions from all authors.

469

470 **Competing interests** - The authors declare that they have no conflict of interest.

471

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491

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Table 1. Ocean acidification studies estimating or documenting pH change in the Mediterranean Sea. TrOCA is the ‘Tracer combining Oxygen, inorganic Carbon, and total Alkalinity’ method, N.R. means ‘not reported’, and PI is ‘pre-industrial era’. *indicates studies where the reported pH change was assumed to be at *in situ* temperatures.

Region	Site	Method	Study period	pH scale	°C	$\Delta pH \text{ yr}^{-1} \pm SE$	Total ΔpH	Reference
NW	Point B	time-series, anomaly	2007-2015	total	<i>in situ</i>	-0.0028 ± 0.0003	-0.0252	This study
NW	Point B	time-series, anomaly	2007-2015	total	25	-0.0017 ± 0.0002	-0.0153	This study
NW	Point B	model	1967-2003	total	<i>in situ</i>	-0.0014	-0.05	Howes et al. (2015)
NW	DYFAMED	time-series, observed	1995-2011	seawater	17.34	-0.003 ± 0.001	-0.051	Marcellin Yao et al. (2016)
NW	DYFAMED	time-series comparison	1998-2000, 2003-2005	seawater	<i>in situ</i> *	-	-0.02	Meier et al. (2014)
NW	Gulf of Lion	TrOCA	PI-2011	NR	<i>in situ</i> *	-	-0.15 to -0.11	Touratier et al. (2016)
East	N Adriatic Sea	cruise comparison	1983, 2008	total	25	-0.0025	-0.063	Luchetta et al. (2010)
East	Otranto Strait	TrOCA	PI-1995	seawater	25	-	< -0.1 to -0.05, ± 0.014	Krasakopoulou et al. (2011)
Total	Full profile	TrOCA	PI-2001	NR	<i>in situ</i> *	-	-0.14 to -0.05	Touratier and Goyet (2011)
Total	Bottom waters	model	1800-2001	total	<i>in situ</i> *	-	-0.06 to -0.005	Palmiéri et al. (2015)
Total	Surface waters	model	1800-2001	total	<i>in situ</i> *	-	-0.084 ± 0.001	Palmiéri et al. (2015)
Gibraltar Strait	Espartel sill	pH, pCO ₂ sensors	2012-2015	total	25	-0.0044 ± 0.00006	-	Flecha et al. (2015)



Table 2. Time-series (observed values and anomalies) regression analyses on seawater carbonate chemistry, at Point B, 1 m, for salinity (S), temperature (T, °C), dissolved inorganic carbon (C_T , $\mu\text{mol kg}^{-1}$), total alkalinity (A_T , $\mu\text{mol kg}^{-1}$), pH_T , 25 °C-normalized pH_T (pH_{T25}), calcite (Ω_c) and aragonite (Ω_a) saturation state, and salinity-normalized A_T (nA_T) and C_T (nC_T). Slope is change yr^{-1} . $P < 0.001$ indicate p-values far smaller than 0.001.

	<i>Variable</i>	<i>Slope</i> \pm <i>SE</i>	<i>Intercept</i> \pm <i>SE</i>	<i>F</i>	<i>df</i>	<i>Slope P</i>	<i>R</i> ²
Observed	S	0.0031 \pm 0.0054	31.7 \pm 10.8	0.326	1,415	0.568	0.001
	T	0.12 \pm 0.08	-226 \pm 157	2.429	1,411	0.12	0.006
	C_T	2.71 \pm 0.31	-3208 \pm 626	75.671	1,414	<<0.001	0.155
	A_T	2.24 \pm 0.21	-1945 \pm 418	115.895	1,415	<<0.001	0.218
	pH_T	-0.0031 \pm 0.0009	14.3 \pm 1.8	11.674	1,410	0.001	0.028
	pH_{T25}	-0.0012 \pm 0.0004	10.5 \pm 0.9	6.861	1,410	0.009	0.016
	pCO_2	3.78 \pm 0.98	-7214 \pm 1976	14.82	1,410	<<0.001	0.035
	Ω_c	-0.0043 \pm 0.0052	13.9 \pm 10.5	0.684	1,410	0.409	0.002
	Ω_a	-0.0017 \pm 0.004	6.74 \pm 7.98	0.177	1,410	0.674	0
	nA_T	2.01 \pm 0.31	-1489 \pm 632	41.067	1,410	<<0.001	0.091
	nC_T	2.57 \pm 0.47	-2918 \pm 944	29.927	1,410	<<0.001	0.068
Anomaly	S	-0.0017 \pm 0.0044	3.38 \pm 8.82	0.147	1,415	0.702	0
	T	0.072 \pm 0.022	-145 \pm 44	10.999	1,411	0.001	0.026
	C_T	2.97 \pm 0.20	-5965 \pm 400	221.87	1,414	<<0.001	0.349
	A_T	2.08 \pm 0.19	-4189 \pm 379	122.429	1,415	<<0.001	0.228
	pH_T	-0.0028 \pm 0.0003	5.72 \pm 0.66	74.205	1,410	<<0.001	0.153
	pH_{T25}	-0.0017 \pm 0.0002	3.46 \pm 0.43	64.204	1, 410	<<0.001	0.1354
	pCO_2	3.53 \pm 0.39	-7105 \pm 776	83.927	1,410	<<0.001	0.17
	Ω_c	-0.0109 \pm 0.0022	22.0 \pm 4.5	24.08	1,410	<<0.001	0.055
	Ω_a	-0.0064 \pm 0.0015	12.9 \pm 3.1	17.33	1,410	<<0.001	0.041
	nA_T	2.20 \pm 0.28	-4425 \pm 560	62.34	1,410	<<0.001	0.132
	nC_T	3.12 \pm 0.29	-6275 \pm 579	117.486	1,410	<<0.001	0.223

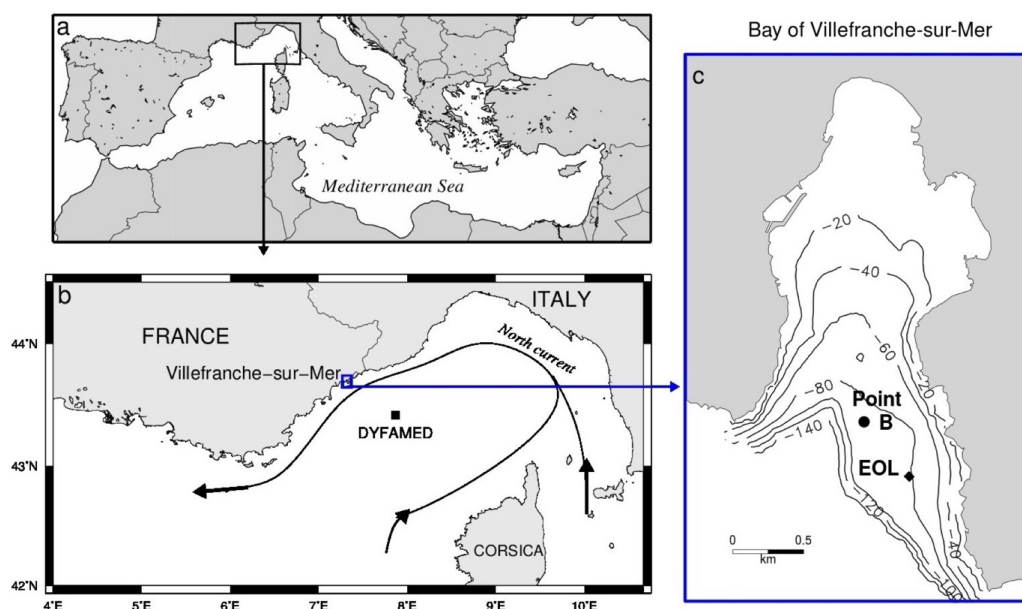


Table 3. Contribution of temperature (T), salinity (S), total alkalinity (A_T), and dissolved inorganic carbon (C_T) to observed changes in pH_T and pCO_2 (μatm) yr^{-1} . The sum of the slopes ($\frac{dpH_T}{dt}, \frac{dpCO_2}{dt}$) is slightly inflated compared to the observed trends reported in Table 2 (-0.0037 vs. -0.0031 yr^{-1} for pH_T , and 4.26 vs. 3.78 μatm yr^{-1} for pCO_2). These differences are negligible relative to the error associated with the slope estimates. Incomplete sum of % contributions are due to rounding. P of $<<0.001$ indicate p-values far smaller than 0.001 .

	<i>Variable</i>	<i>Slope \pm SE</i>	<i>% contribution</i>	<i>Slope P</i>
$\frac{\partial pH_T}{\partial var} \frac{dvar}{dt}$	T	-0.0013 ± 0.0009	35	0.15
	S	-0.0002 ± 0.0008	5	0.793
	A_T	0.0028 ± 0.0009	-76	0.002
	C_T	-0.0050 ± 0.0009	135	$<<0.001$
$\frac{dpH_T}{dt}$		-0.0037	99	
$\frac{\partial pCO_2}{\partial var} \frac{dvar}{dt}$	T	1.43 ± 0.97	34	0.143
	S	0.24 ± 0.89	6	0.792
	A_T	-2.94 ± 0.94	-69	0.002
	C_T	5.53 ± 0.96	130	$<<0.001$
$\frac{dpCO_2}{dt}$		4.26	101	



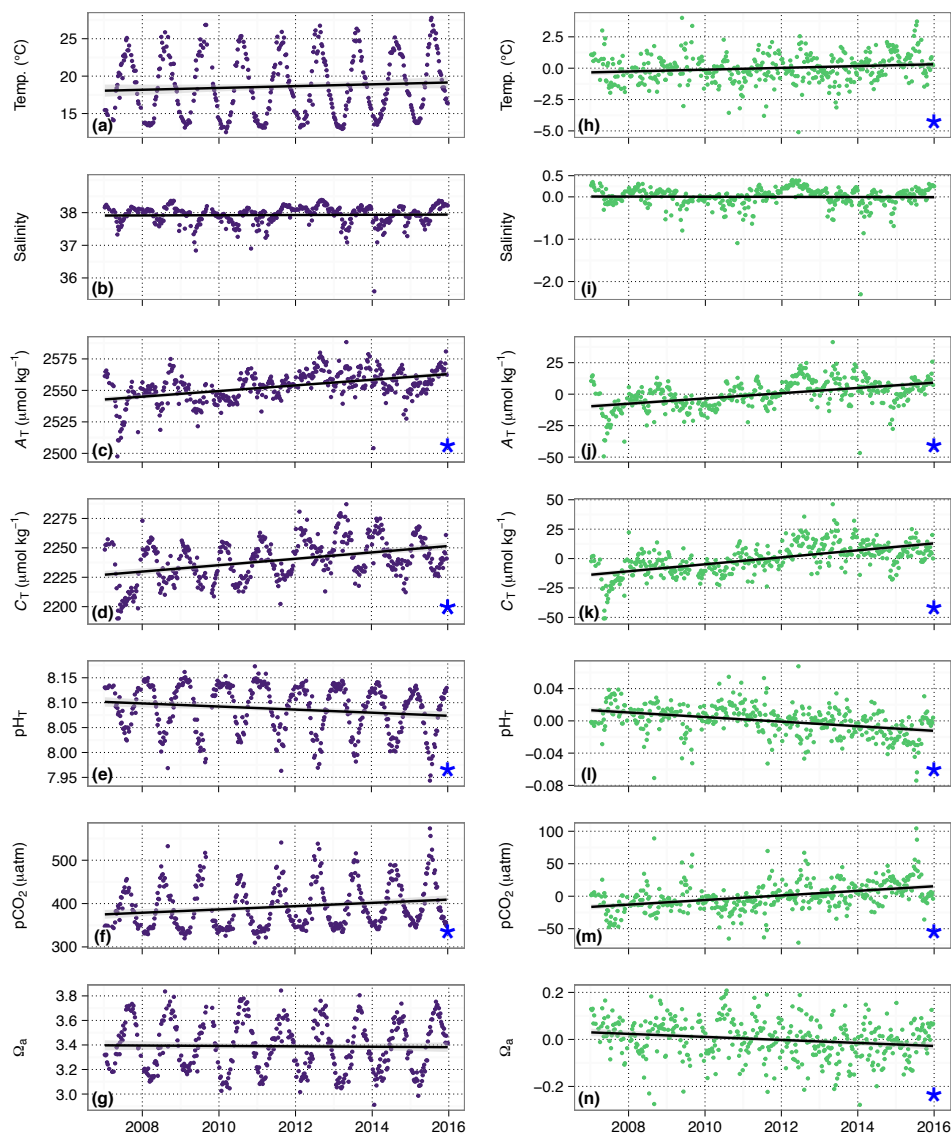
Figure 1. Map of study region in the NW Mediterranean Sea (a) along the North current (b) in
 the Bay of Villefranche-sur-Mer, France (c). Point B station, EOL buoy, and offshore time-series
 station DYFAMED are marked. Bathymetric line units are m (c).



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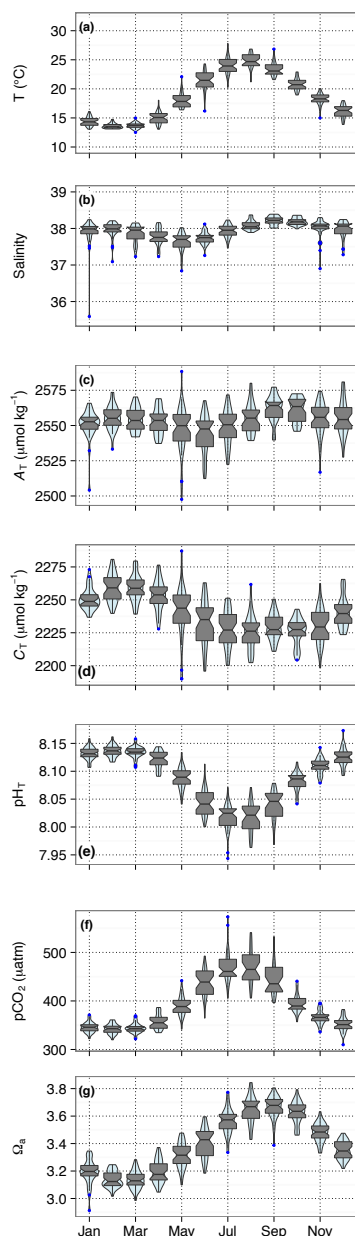
782 **Figure 2.** Time-series observations (a-g) and anomalies (h-n) of temperature, salinity, and
 783 seawater carbonate chemistry at Point B, 1 m. Regression slopes are drawn \pm SE (in grey) and
 784 noted with a star for significance at $\alpha=0.05$.



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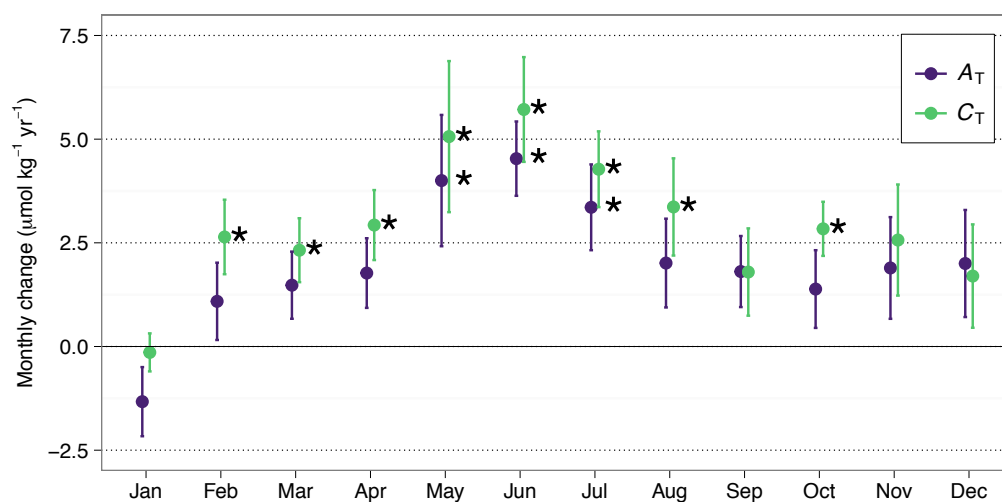
786 **Figure 3.** Monthly distribution of seawater carbonate chemistry at Point B, 1 m., using a
 787 combination of a violin plot showing the relative frequency of the observations (shaded blue
 788 area) and a boxplot showing the median, first and third quartiles, as well as outliers (blue).



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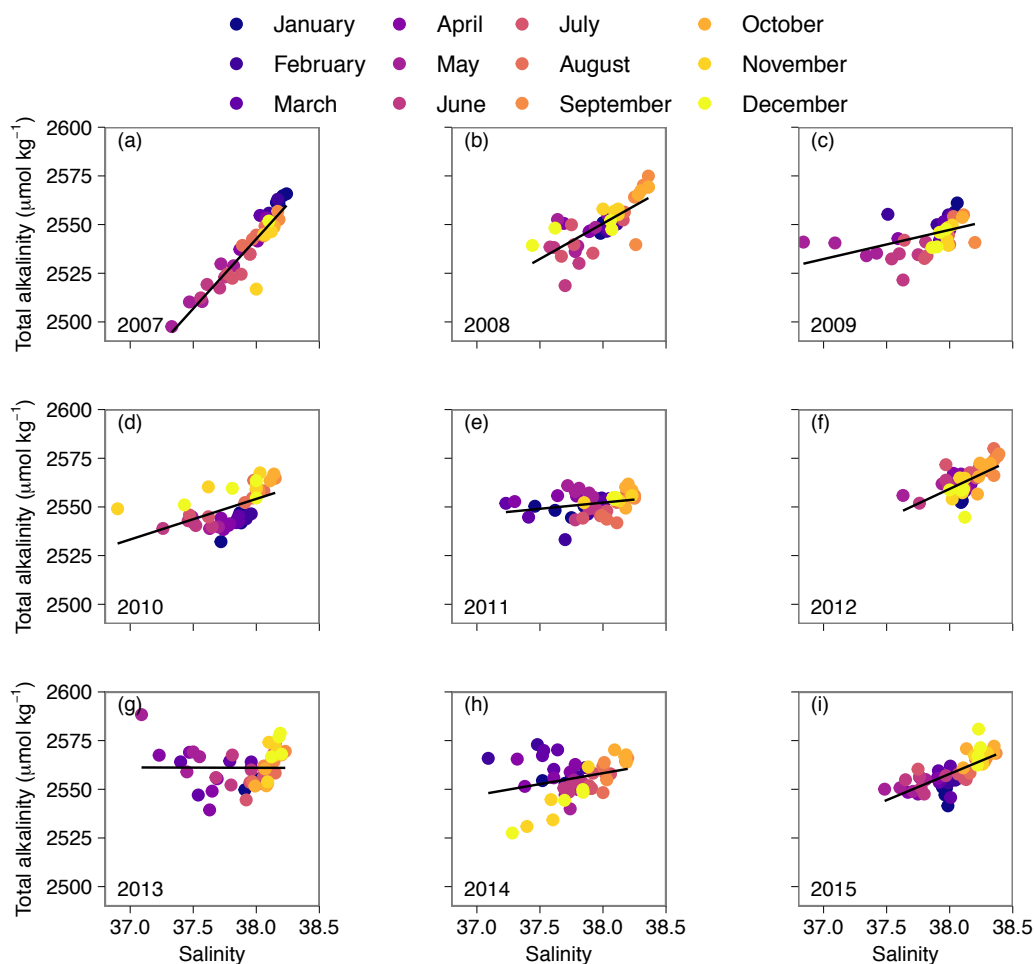
790 **Figure 4.** Increase in total alkalinity (A_T , purple) and dissolved inorganic carbon (C_T , green) by
 791 month for the period 2007-2015. Errors bars are \pm SE of the slope estimate and significance is
 792 noted (*) at $\alpha=0.05$.



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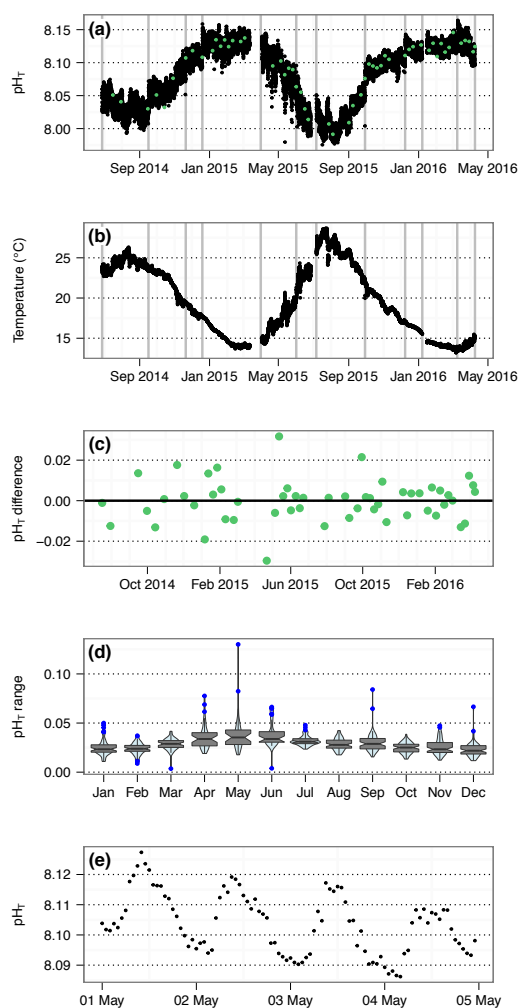


794 **Figure 5.** Salinity and total alkalinity relationships by year for the period 2007-2015, at Point B,
 795 1 m. Data points are colored for month.





797 **Figure 6.** Time-series pH (a) and temperature (b) from autonomous SeaFET pH sensor at EOL
 798 buoy, 2 m. Discrete calibration samples are noted in green, and grey vertical lines bracket
 799 deployment periods (a). Mean offset of calibration samples from processed pH time-series was
 800 $\text{pH}_T \pm 0.007$ (c). Diel pH range was small and peaked in April and May (d) and exhibits a clear,
 801 small, diel cycle (e, representative example from May 2015).



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