



Plastics in the Indian Ocean – sources, transport, distribution, and impacts

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Abstract. Plastic debris is the most common and exponentially increasing human pollutant in the world's ocean. The distribution and impact of plastic in the Pacific and Atlantic oceans have been the subject of many publications but not so the Indian Ocean (IO). Some of the IO rim countries have the highest population densities globally and mismanagement of plastic waste is of concern in many of these rim states. Some of the most plastic-polluted rivers empty into the IO, with all this suggesting that the IO receives a tremendous amount of plastic debris each year. However, the concentration, distribution, and impacts of plastics in the IO are poorly understood as the region is under-sampled compared to other oceans. In this review, we discuss sources and sinks, which are specific to the IO. We also discuss unique atmospheric, oceanographic, and topographic features of the IO that control plastic distribution, such as reversing wind directions due to the monsoon, fronts, and upwelling regions. We identify hotspots of possible plastic accumulation in the IO, which differ between the two hemispheres. In the southern IO, plastics accumulate in a garbage patch in the subtropical gyre. However, this garbage patch is not well defined, and plastics may leak into the southern Atlantic or the Pacific Ocean. There is no subtropical gyre and associated garbage in the northern IO due to the presence of landmasses. Instead, the

majority of buoyant plastics most likely end up on coastlines. Finally, we identify the vast knowledge gaps concerning plastics in the IO and point to the most pressing topics for future investigation.

1 Introduction

Historically, the motivation for developing synthetic materials like plastics was for the conservation of elephants that inhabit countries along the Indian Ocean (IO) rim in southern Asia and Africa (Freinkel, 2011). The first plastic materials were advertised as saviours of the environment because it would no longer be necessary to ransack the environment for scarce natural resources (Meikle, 1997). However, the production of plastic materials has increased exponentially since the 1950s (PlasticsEurope, 2019), and plastics have instead become a ubiquitous environmental pollutant (Law, 2017). Since a large percentage of all plastics are single-use packaging items (PlasticsEurope, 2019), plastic waste has increased at a similar rate (Geyer et al., 2017). Since the 1950s, the global human population has generated an estimated 8300×10^6 t of plastic, of which over 75 % (6300×10^6 t) has been discarded as waste (Geyer et al., 2017). Almost 80 % of

this plastic waste either ends up in landfill or in the environment. As a result, several million tonnes of plastic waste are estimated to enter the global oceans every year from coastlines (Jambeck et al., 2015), from inland sources transported by rivers (Lebreton et al., 2017; Schmidt et al., 2017, 2018), or directly from ocean-based sources such as the global fishing industry (Richardson et al., 2019b), offshore platforms, and commercial and recreational shipping.

Plastic materials are durable and very slow to degrade and can persist in the marine environment for decades to centuries (Barnes et al., 2009). Around 35 % of all plastic materials produced globally have densities higher than that of seawater (PlasticsEurope, 2019) and thus can sink to the seafloor. The remaining 65 % of plastics float in the ocean and can travel enormous distances in the ocean. Photodegradation and other weathering processes at the sea surface lead to fragmentation, increasing micro- and nanoplastic abundance (Andrady, 2011). Plastics also accumulate biofouling while in the ocean, which can change the overall density and lead to plastics moving vertically in the water column (Lobelle and Cunliffe, 2011; Long et al., 2015; Kooi et al., 2017). As a result, plastics are ubiquitous throughout the marine environment and have been found on remote islands (Duhec et al., 2015; Lavers and Bond, 2017), in polar ice (Obbard et al., 2014; Bergmann et al., 2017; Peeken et al., 2018), and in the deep seas (Van Cauwenberghé et al., 2013; Jamieson et al., 2019; Courtene-Jones et al., 2020).

Plastic pollution has several known harmful effects on marine species and ecosystems, including ingestion, entanglement, and the transport of potentially invasive species into foreign remote habitats (Gregory, 2009; Law, 2017). Plastics can also absorb and transfer toxicants (Rochman et al., 2012), potentially leading to the accumulation of toxins across the food web (Engler, 2012). Marine plastic debris also causes significant economic damage to industries and communities, with estimated costs ranging between USD 8 billion (Raynaud, 2014) to USD 2500 billion (Beaumont et al., 2019). Economic costs can be caused by damage to ships and fishing equipment when they collide with big marine plastic fragments (Richardson et al., 2019a) and by soiled landscapes that impair the tourism industry when it relies on clean beaches and coastal waters (Sari et al., 2021).

However, the overall impact of plastic pollution in the ocean is still poorly understood (Law, 2017). There are challenges in investigating ecological impacts and the fate of plastics in the ocean. Estimates of the amount of plastic floating on the global ocean surface are approximately 1 % or less of the estimated amount of plastic that enters the ocean per year (van Sebille et al., 2015). This discrepancy in the global ocean plastic budget highlights a fundamental gap in the understanding of the fate of ocean plastic and suggests that there are unknown sinks of marine plastic debris. Possible sinks include biological sinks, i.e. organisms ingesting plastics (e.g. Davison and Asch, 2011); sea ice cover temporarily accumulating plastics (Obbard et al., 2014; Peeken

et al., 2018); fragmentation and biofouling of plastics leading to settling (e.g. Koelmans et al., 2017); and plastics making landfall (“beaching”) and accumulating on the world’s coastlines (e.g. Lebreton et al., 2019; van der Mheen et al., 2020b).

Plastic debris in the IO has been under-sampled and under-studied compared to other oceans, whilst the fate of plastics has not been investigated in any detail. However, a large percentage of all global ocean plastic is estimated to enter the IO: up to 15 % of all coastal plastic (Jambeck et al., 2015) and 20 % of all riverine plastic (Lebreton et al., 2017; Schmidt et al., 2017, 2018). Half of the top 10 countries contributing most to ocean plastic pollution are located along the IO rim (Indonesia, Thailand, Malaysia, India, and Bangladesh), and two of the largest and most polluting rivers (Ganges and Indus) empty into the IO. In addition, it has been estimated that the IO contains the second-largest plastic load in the ocean after the North Pacific Ocean (Eriksen et al., 2014), although there are insufficient measurements of plastics available to confirm this. In this synthesis paper, we compile existing knowledge about the sources (Sect. 2), observations (Sect. 3), transport and fate (Sect. 4), and impact (Sect. 5) of plastic debris in the IO as well as highlight some of the emerging policies and initiatives (Sect. 6) and knowledge gaps and recommend future research strategies (Sect. 8). We complete the paper by referring to a recent ship accident that released 78 000 t of nurdles (plastic pellets) into the Indian Ocean off Sri Lanka in May 2021 (Sect. 7).

2 Sources

Plastic waste enters the ocean from coastal sources transported by wind and tides (Jambeck et al., 2015), from sources far in the hinterland transported by rivers (Lebreton et al., 2017; Schmidt et al., 2017, 2018), and directly from ocean-based sources (Richardson et al., 2019b). The IO is surrounded by 34 countries with population densities of around 100 people per km² on average, with Australia the least populated country (3 people per km²) and India the most populated country (400 people per km²). Hoornweg and Bhada-Tata (2012) estimated that the total amount of plastic waste produced by all IO rim countries in 2010 was around 41×10^6 t. In comparison, the total amount of plastic waste produced in 2010 by the USA and China alone is estimated to be 38 and 59×10^6 t, respectively. More recently, Kaza et al. (2018) estimated that the total amount of plastic waste produced by IO rim countries in 2016 was around 24×10^6 t, compared to 34×10^6 t by the USA and 39×10^6 t by China. Despite the relatively low plastic waste production around the IO rim, a large percentage ends up in the environment because of poor waste management in most Asian and African countries. Jambeck et al. (2015) estimated that around 73 % of plastic waste is inadequately managed along the IO rim and is released into the environment.

2.1 Land-based sources

Jambeck et al. (2015) estimated the global amount of plastic waste that entered the ocean in 2010 from coastal populations living within 50 km of the coastline. Based on a fixed percentage of mismanaged plastic waste entering the oceans (15 % for the low-range estimates and 40 % for the high-range estimates), they estimated that between 4.8 and 12.7×10^6 t of plastic entered the global oceans. However, it is likely that the estimated amount of plastic waste entering the ocean by Jambeck et al. (2015) for Sri Lanka is incorrect. Jambeck et al. (2015) based their estimate on a reported 5.1 kg of municipal solid waste generated per person per day in Sri Lanka (Hoorweg and Bhada-Tata, 2012). The updated report by Kaza et al. (2018) and dataset available through the World Bank (2021) (“What A Waste Global Database”) indicate that only 0.34 kg of municipal solid waste is generated per person per day in Sri Lanka; this number is also more in line with the amount of waste generated in other developing countries. Using this correction, the amount of plastic waste entering the ocean from Sri Lanka through coastal populations is estimated to be between 0.021 and 0.057×10^6 t in 2010, instead of between 0.24 and 0.64×10^6 t as reported in Jambeck et al. (2015). Using this corrected number, around 15 % of global ocean plastic entered the IO directly through coastal sources (Fig. 1a; showing corrected waste input for Sri Lanka).

Lebreton et al. (2017) and Schmidt et al. (2017) estimated the amount of plastic waste entering the oceans through rivers. Similar to Jambeck et al. (2015), their estimates are based on a percentage of mismanaged plastic waste and include the influence of river catchment geography and river discharge. In addition, these estimates were calibrated based on available measurements of river plastic debris globally, ranging between sizes of 0.3 mm to 0.5 m. Lebreton et al. (2017) estimated that between 1.15 and 2.41×10^6 t of plastic waste enter the global ocean per year, up to 20 % of which enters the IO (Fig. 1b). Schmidt et al. (2017) estimated that between 0.47 and 2.75×10^6 t of plastic waste enter the global ocean per year, of which around 15 % enters the IO. More recently, Meijer et al. (2021) estimated that between 0.80 and 2.7×10^6 t of macroplastics (defined by Meijer et al., 2021, as larger than 5 mm) enter the global ocean per year. In this estimate, Meijer et al. (2021) took into account the spatial variability of mismanaged plastic waste generated within a river basin as well as more advanced climate and terrain characteristics than considered in the estimates of Lebreton et al. (2017) and Schmidt et al. (2017). They calibrated their estimates based on visual sampling of macroplastics at river mouths around the world.

The estimates of the amount of plastic waste entering the oceans through rivers by Lebreton et al. (2017), Schmidt et al. (2017), and Meijer et al. (2021) agree relatively well with each other. Jambeck et al. (2015) estimated that the amount of plastic waste entering the oceans through coasts is an or-

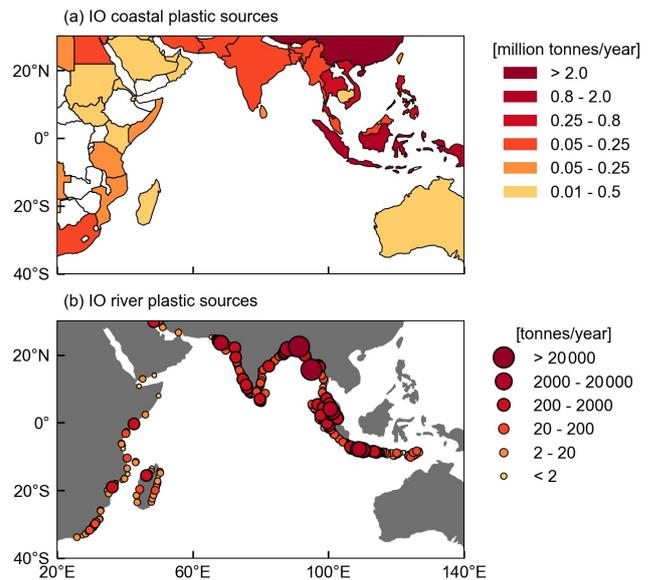


Figure 1. (a) Estimated coastal sources of plastic waste entering the Indian Ocean, based on data from Jambeck et al. (2015). Plastic waste input estimates for Sri Lanka have been reduced by a factor of 10 in this map to correct for a mistake in the Jambeck et al. (2015) data. (b) Estimated river sources of plastic waste entering the Indian Ocean, based on data from Lebreton et al. (2017). The riverine inputs from Australia are negligible due to lack of major rivers. Please note that country boundaries in (a) are from outputs of scientific computer simulations and may not represent political and/or geographical realities.

der of magnitude higher. In even starker contrast, Weiss et al. (2021) re-evaluated the estimates of Lebreton et al. (2017) and Schmidt et al. (2017) and suggested that only 6100 t of microplastics (defined by Weiss et al., 2021, as smaller than 5 mm) enter the ocean through rivers each year, which is 2 to 3 orders of magnitude smaller than previous estimates. These differences highlight the extreme uncertainty involved in determining the amount of plastic waste entering the ocean from land-based sources. These estimates are based on few measurements of plastics entering the ocean (in the case of Jambeck et al., 2015, only on data from the San Francisco Bay; in the case of Lebreton et al., 2017, Schmidt et al., 2017, Meijer et al., 2021, and Weiss et al., 2021, on 30 to 340 samples from 13 to 89 rivers around the world). None of these samples was from IO rim countries or in rivers that empty into the IO. Expanding on these datasets will likely help improve these estimates, especially for the IO. However, as Weiss et al. (2021) demonstrate, to reduce extreme errors it is essential to use comparable sampling methodologies and collect data on the amount of plastics sampled and their weight. Furthermore, Meijer et al. (2021) emphasize the importance of sampling plastics at river mouths to get a more reliable estimate of the amount of plastic that actually enters the ocean. However, sampling plastics further upstream in

addition to the river mouth can help improve models of the probability for plastic to reach the ocean from inland areas.

Based on the currently available estimates, the largest coastal and riverine plastic sources in the IO are in the Northern Hemisphere around the Bay of Bengal and on the eastern side of the Arabian Sea (Fig. 1). The input of plastic waste from rivers mainly depends on the river discharge, which varies seasonally. The highest river discharges occur during the wet season, during the boreal summer in the northern IO. Lebreton et al. (2017) estimated that plastic waste input from rivers in the IO peaks in August. In the Southern Hemisphere, the largest coastal and riverine sources of IO plastic waste are from Indonesia (Fig. 1).

2.2 Ocean-based sources

Plastic waste can also enter the ocean directly from ocean-based sources such as the fishing industry, commercial and recreational shipping, and offshore platforms. In 1988, the International Convention for the Prevention of Pollution from Ships (MARPOL) prohibited waste dumping from vessels. However, accidental losses and illegal dumping still contribute to plastic debris (Richardson et al., 2021; Stelfox et al., 2020a; see Sect. 7 for a case study of a recent incident off the coast of Sri Lanka). Abandoned, lost, or discarded fishing gear (ALDFG) by the fishing industry can produce large quantities of plastic debris, i.e. monofilament lines and nets primarily made from synthetic materials (Sheavly and Register, 2007; Bond et al., 2012). Since 2000, around 13 % of the global marine capture of fishes originated from the IO region (Pauly and Zeller, 2016). It is, therefore, likely that ALDFG is a significant source of plastic debris in the IO. Estimates of plastic waste entering the ocean from fishing vessels (e.g. ghost nets) or offshore platforms do not currently exist.

Finally, plastic waste originating from southeast Asia can also be transported to the IO by the Indonesian Through-flow (Sect. 4.1). However, these sources are currently undocumented and need to be further investigated.

3 Observations

Plastics have been sampled in the IO in coastal waters, the open ocean, sediment, and organisms since the 1980s (Table 1), with most sampling studies being conducted on beaches. However, measurements are relatively scarce, and plastics have been regularly under-sampled in the IO (Fig. 2) compared to the Atlantic and Pacific oceans. Samples of plastic debris consist of different plastic polymers and are generally classified into different types and size categories. Size and type categories can vary widely between authors but it is beyond the scope of this review to discuss these different categories. Instead, we refer to recent review papers by Gigault et al. (2018) and Frias and Nash (2019) discussing

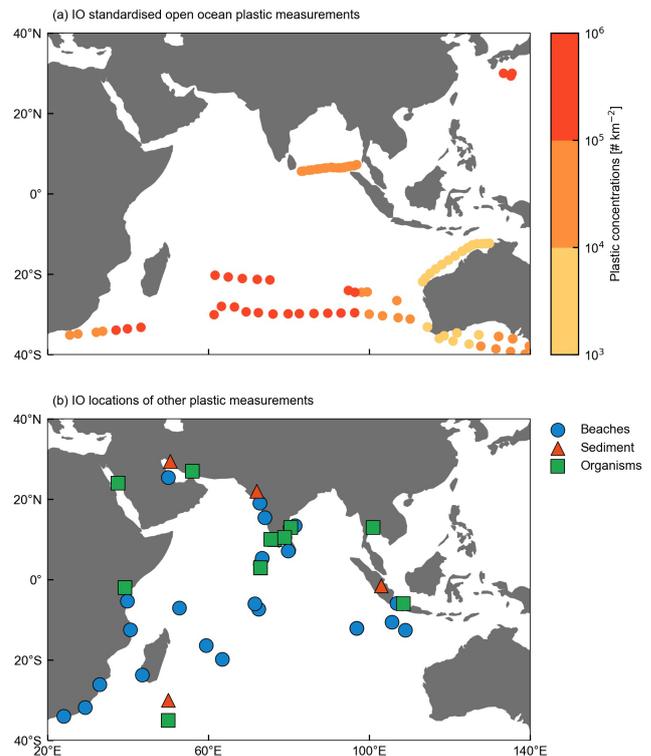


Figure 2. (a) Standardized measured concentrations of plastics in the Indian Ocean. Original samples were performed by Morris (1980), Reisser et al. (2013), Eriksen et al. (2014), and Cózar et al. (2014) and were standardized by van Sebille et al. (2015). (b) Locations of plastic samples taken on beaches, in sediment, and in organisms around the Indian Ocean (references to individual studies listed in Table 1). Because sampling methods vary widely between studies, a quantitative comparison of measured plastic concentrations between these studies is not possible.

plastic size categories and Hartmann et al. (2019) discussing different categories of polymers, sizes, shapes, colours, and origins.

Global open-ocean plastic samples were standardized by van Sebille et al. (2015), and the plastic concentrations from these samples in the IO can be quantitatively compared (Fig. 2a). In contrast, the methods used for sampling plastics on beaches and in sediment vary widely (as illustrated in Table 1) and offer only a qualitative confirmation that plastics have been found on beaches and in sediment throughout the IO (Fig. 2b). As discussed extensively in the review by Serra-Gonçalves et al. (2019), adopting a standardized framework to collect and report on beach debris is essential for these studies to be useful to the wider scientific community. Isobe et al. (2019) discuss the importance of a standardized protocol for laboratory analysis of plastics.

Multiple studies of plastic ingestion by different types of fauna have been done in the IO (Table 1, Sects. 4.3.2 and 5.2).

Table 1. Overview of plastic sampling studies performed in the Indian Ocean.

Location	Latitude [° N]	Longitude [° E]	Observation site	Categories ^a			Reference
				Size [mm]	Shape/ type	Polymers	
<i>Surface waters</i>							
Northeastern Qatar	25.3	52.4	Surface waters	[< 0.125, 15.98]	Granules, fibres	PP, PE, LDPE, PA, ABS	Castillo et al. (2016)
Doha Bay, Qatar	25.5	50.1	Surface waters	[< 0.5, 5]	Fibre, film, fragment	LDPE, LDPP	Abayomi et al. (2017)
Kuala Nerus and Kuantan, Malaysia			Surface waters	< 5 irregular	Filaments, fragments, PP, PS	PES, PE, PA, PVC,	Khalik et al. (2018)
Faafu Atoll, Maldives	3.1	72.97	Surface waters	[0.05, 25]	Fragments, foils, pellets, fibres, foam	PE, PP, PS, PET, PU, PVC	Saliu et al. (2018)
Chabahar Bay, Gulf of Oman (Makran coasts), Iran	25.3	60.4	Surface waters	[0.1, 5]	Fragment, pellets, fibres, paint flakes	PE, PP, PES	Aliabad et al. (2019)
Ashmore Reef, Australia	−12	123	Surface waters	[0.5, 1.5], [1.5, 5.0], [5.0, 15], > 15	Film, fragments, foam	PE, PP	Hajbane et al. (2021)
<i>Surface waters and other</i>							
Along Cilacap coast, Indonesia	[−18.4, −7.7]	109.1	Surface waters, beaches	[2.5, 5]	–	PP, LDPE, HDPE, PVC, PET, PS, PC	Syakti et al. (2017)
Southern coastline of Sri Lanka	–	–	Surface waters, beaches	[1.5, 4.5]	Pellets, fragments, films, filaments, foam	PE, PP, PS	Koongolla et al. (2018)
Kenyan coastline	1.7	39.2	Surface waters, zooplankton	[0.01, 2.4]	Fragments, fibres, pellets, film, foam	PP, LDPE; LDPE	Kosore et al. (2018)
Maldivian archipelago	3.1	72.97	Surface waters, corals	< 5	Fragments, films, filaments, foam	PAHs, PE, PP, PA, PS, PU	Saliu et al. (2019)
Southwest coast of India	[8.3, 12.8]	[74.9, 77.4]	Surface waters, sediment, fish fauna	[0.3, 5]	Pellets, film, fragments, fibres/line, foam	PE, CE, PES, PP, RY	Robin et al. (2020)

Table 1. Continued.

Location	Latitude [° N]	Longitude [° E]	Observation site	Categories ^a			Reference
				Size [mm]	Shape/ type	Polymers	
<i>Beaches</i>							
Prince Edward Island	−46.61	37.96	Beaches	> 10	Foam, bags, packing strips, bottles and containers	–	Ryan (1987)
Marion Island	−46.84	37.85	Beaches	> 10	Foam, bags, packing strips, bottles and containers	–	Ryan (1987)
Heard Island	−53.00	73.48	Beaches	> 10	Foam, bottles, fragments, rope and net fragments	–	Slip and Burton (1991)
Macquarie Island	−54.62	158.80	Beaches	> 10	Foam, bottles, fragments, rope and net fragments	–	Slip and Burton (1991)
former Transkei, South Africa	−31.75	29.38	Beaches	1, 1–10, 11–100, 101–1000, > 1000 cm ²	Plastics, foam, fishing gear	–	Madzena and Lasiak (1997)
Jakarta Bay, Indonesia	−5.86	106.85	Beaches	–	Plastic bags, footwear, foam, plastic bottles, rope and net fragments	–	Unepetty and Evans (1997)
Negombo, Sri Lanka	7.16	79.82	Beaches	> 1 cm ²	–	–	Barnes (2004)
Ari Atoll, Maldives	3.88	72.83	Beaches	> 1 cm ²	–	–	Barnes (2004)
Pemba Island, Tanzania	−5.30	39.85	Beaches	> 1 cm ²	–	–	Barnes (2004)
Diego Garcia	−7.34	72.51	Beaches	> 1 cm ²	–	–	Barnes (2004)
Christmas Island	−10.54	105.59	Beaches	> 1 cm ²	–	–	Barnes (2004)
Cocos (Keeling) Islands	−12.08	96.88	Beaches	> 1 cm ²	–	–	Barnes (2004)
Quirimbas Island, Mozambique	−12.44	40.62	Beaches	> 1 cm ²	–	–	Barnes (2004)
Rodrigues Island	−19.76	63.46	Beaches	> 1 cm ²	–	–	Barnes (2004)

Table 1. Continued.

Location	Latitude [° N]	Longitude [° E]	Observation site	Categories ^a			Reference
				Size [mm]	Shape/ type	Polymers	
Nosy Ve, Madagascar	−23.64	43.50	Beaches	> 1 cm ²	–	–	Barnes (2004)
Inhaca Island, Mozambique	−26.02	33.00	Beaches	> 1 cm ²	–	–	Barnes (2004)
Mumbai, India	18.9	72.9	Beaches	[< 5, > 100]	Fragments, pellets	–	Jayasiri et al. (2013)
The Chagos Archipelago, Chagos (BIOT)	−6	71.5	Beaches	[0.03, 4]	Fragments, fibres	NY, PE, PES, PP, RY	Readman et al. (2013)
Alphonse Island, Western Indian Ocean, Seychelles	−7	52.7	Beaches	–	Footwear, fragments, fishing nets, foam, hard plastic, soft plastic, plastic caps	PET, PVC, PP, HDPE, PE, PS	Duhec et al. (2015)
Southeast coast, South Africa	−34.00	24.00	Beaches	[0.065, 5]	Fibres, fragments, foam	–	Nel and Froneman (2015)
Chennai, India	13.05	80.28	Beaches	–	Plastic bags, food wrappers, cups, containers, bottles and caps, rope and net fragments	–	Arun Kumar et al. (2016)
St. Brandon's Rock, Mauritius	−16.38	59.45	Beaches	> 5	Foam, footwear, plastic, rope	PU, PS, “plastics”	Bouwman et al. (2016)
Chennai coast, southeast coast of India	13	81.3	Beaches	[2, 5]	Ovoid, spheroids, discs, cylindrical rods	PE, PP	Veerasingam et al. (2016a)
Beaches of Goa, India	[15.0, 15.75]	[73.75, 74.25]	Beaches	[1, 5]	Cylindrical, spherical, oval	PE, PP	Veerasingam et al. (2016b)
Coastline of Qatar	25.5	50.1	Beaches	[< 0.5, 5]	Fibre, film, fragment	LDPE, LDPP	Abayomi et al. (2017)
Vavvaru Island, Maldives	5.42	73.35	Beaches	[1, 5], > 5	Fragments, foils, pellets, fibres, foam	PE, PP, PS	Imhof et al. (2017)

Table 1. Continued.

Location	Latitude [° N]	Longitude [° E]	Observation site	Categories ^a			Reference
				Size [mm]	Shape/ type	Polymers	
Tamil Nadu coast, India	[8, 13]	[78, 80]	Beaches	[0.3, 9]	Fragments, foam, pellets, film, fibres	PE, PP, PS, NY	Karthik et al. (2018)
Cocos (Keeling) Islands	−12.08	96.88	Beaches	[2, 5], > 5	Fragments, foam, pellets, bottle caps, plastic bags, packaging, fishing line, crates, footwear, straws	–	Lavers et al. (2019)
Coastal areas, Tamil Nadu, India	[8.1, 13.1]	[77.95, 80.5]	Beaches	[0.5, 3]	Fibre, fragment, foam	PE, PP, PS, NY, PES	Sathish et al. (2019)
<i>Sediments</i>							
Alang-Sosiya ship-breaking yard, Gujarath, India	22	72	Intertidal	[0.0016, 5] sediment	Fragments, fibres, films	PU, NY, PES, PS	Reddy et al. (2006)
Mangrove sediments, northwest coast of Singapore	1.2		Sediments	[< 0.02, 5]	Fibres, films, granules	PE, PP, NY, PVC	Nor and Obbard (2014)
Seamounts, southwest Indian Ocean	−30	[40, 60]	Sediment	[2, 3]	Fibres	RY, ACT, ACR, PES, PA	Woodall et al. (2014)
Coastal sediments of Khark Island, Iran	29.35	50.5	Coastal sediment	[0.1, 5]	Fragments, fibres	–	Akhbarizadeh et al. (2017)
Vembanad Lake, Kerala, India	[9.5, 10.2]	[76.2, 76.4]	Lake and estuarine sediment	< 5	Fibre, pellet, fragment, foam, film	HDPE, LDPE, PP, PS	Sruthy and Ramasamy (2016)
Faafu Atoll, Maldives	3.1	72.97	Sediments	[0.05, 25]	Fragments, foils, pellets, fibres, foam	PE, PP, PET, PS, PA, PAN, PU, PVC	Saliu et al. (2018)
Skudai and Tebrau	1.5	103°24'85'' E	River sediment	[0.1, 5]	Fibres/line, fragments, film, foam, beads/ pellets	–	Sarijan et al. (2018)

Table 1. Continued.

Location	Latitude [° N]	Longitude [° E]	Observation site	Categories ^a			Reference
				Size [mm]	Shape/ type	Polymers	
<i>Fauna</i>							
KwaZulu Natal, South Africa	[−31.2, −29]	[30, 31.5]	Large sharks	–	plastic bags, sheets	–	Cliff et al. (2002)
Between Madagascar and Réunion	−22	52	Turtles	> 5	Hard plastic, soft plastic, plastic caps, fishing line and rope, foam	–	Hoarau et al. (2014)
Southwest Indian Ocean	[−40, −30]	[40, 60]	Benthic invertebrates	–	Fibres	PP, PES, ACR, MACR, VI, NF	Taylor et al. (2016)
Crozet Islands	−46	51	Seabirds (albatross)	–	Fragments	–	Cherel et al. (2017)
Kerguelen Islands	−49	70	Seabirds (albatross)	–	Fragments	–	Cherel et al. (2017)
Intertidal area of eastern coastal Thailand	13	101	Sessile invertebrates	< 0.1	Rod, fragments, fibres	PS, PET, PA	Thushari et al. (2017)
Biawak Island, north coast of Indramayu, Indonesia	−5.9	108.4	Below water surface, fish fauna	[1, 5]	Fragments, fibres	–	Ismail et al. (2018)
Tamil Nadu coast, India	[8, 13]	[78, 80]	Fish	< 0.5	Fragments, fibres	PE, PP,	Karthik et al. (2018)
Réunion	–	–	Seabirds	–	Fragments, fibres, films	–	Cartraud et al. (2019)
Juan de Nova Island	–	–	Seabirds	–	Fragments, fibres, films	–	Cartraud et al. (2019)
Frenchman Bay, Western Australia	−35.08	118	Fish (sardines)	< 1, < 2, > 5	Fibres	PP, NY, PE	Crutchett et al. (2020)
Algoa Bay, South Africa	−34	26	Fish (South African mullet)	–	Fibres, fragments	–	McGregor and Strydom (2020)
Agulhas Bank, South Africa	−36.42 −33.84 −34.58 −36.33 −35.5 −34.78	21.33 26.67 25.44 21.49 21.87 24.07	Fish	< 125, [125, 500], [500, 1000], [1000, 2000], > 2000 µm	Fragments, fibres	–	Sparks and Immelman (2020)
Coastal waters of Kochi, southeastern Arabian Sea, India	9.9	76.6	Benthic invertebrates	–	Particles, fibres	PP	Naidu et al. (2018)

Table 1. Continued.

Location	Latitude [° N]	Longitude [° E]	Observation site	Categories ^a			Reference
				Size [mm]	Shape/ type	Polymers	
Iranian coast of the Persian Gulf, Iran	26.97	55.0	Molluscans	[0.01, 5]	Fragments, fibres, films, pellets	PE, PET, NY	Naji et al. (2018)
Coast of Red Sea, Saudi Arabia	[20, 28]	[35, 40]	Fish fauna	[1, 3]	Fragments, films, fibres	PP, PE, PS, PVC, PAN	Baalkhuyur et al. (2018)
Fishing harbour of Chennai, southeastern Arabian Sea, India	13.2	80.3	Bivalves	[0.005, 0.03]	Particles, fibres, colourants	PP and colourants ^b	Naidu (2019)
Maldivian archipelago	3.1	72.97	Corals	–	–	MEP, DEP, DBP, MEPH, BBzP ^a	Saliu et al. (2019)
<i>Other</i>							
Super markets in Mumbai, India	–	–	Commercial salts	[< 0.021, 7]	Fragments, fibres	PES, PET, PA, PE, PS	Seth and Shrivastav (2018)
Asalouyeh city, Iran	27.53	52.6	Air and street dust	[< 0.1, > 1]	Fragments, fibres, films, spheroids	–	Abbasi et al. (2019)

^a Denotes phthalate esters (BBzP). ^b Colourants (dyes or pigments); acrylonitrile butadiene styrene (ABS); modified acrylic (MACR); polycarbonate (PC); acetate (ACT); acrylic (ACR); acrylonitrile butadiene styrene (ABS); bis(2-ethylhexyl) phthalate (DEHP); di-butyl phthalate (DBP); di-ethyl phthalate (DEP); di-methyl phthalate (MEP); high-density polyethylene (HDPE); low-density polyethylene (LDPE); low-density polypropylene (LDPP); mono (2-ethylhexyl) phthalate (MEPH); natural fibres (NF); nylon (NY); phthalic acid esters (PAHs); polyacrylonitrile (PAN); polyamide (PA); polyester (PES); polyethylene (PE); polyethylene terephthalate (PET); polypropylene (PP); polystyrene (PS); polyurethane (PU); polyvinyl chloride (PVC); rayon (RY); viscose (VI).

4 Transport and fate

Approximately 35 % of all produced plastic materials have densities higher than that of seawater (PlasticsEurope, 2019) and, therefore, will sink to the seafloor after entering the ocean. To the best of our knowledge, there have been no studies in the IO on sinking plastics and the transport of plastics through the water column and along the seafloor. Most plastic materials float in the ocean and are transported by ocean surface dynamics (currents, wind, and waves). However, buoyant plastic debris can be vertically distributed in the upper few metres of the water column because of wind- and wave-induced mixing (Kukulka et al., 2012; Kukulka and Brunner, 2015; Brunner et al., 2015). Depending on the sea state and the characteristics of the plastic debris, sampling studies in the North Atlantic Ocean found plastic debris mixed up to 5 m depth in the water column (Reisser et al., 2015; Kooi et al., 2016). In addition, the density of plastic debris can change under the influence of biofouling and degrada-

tion (Lobelle and Cunliffe, 2011; Long et al., 2015; Kooi et al., 2017). As a result, initially buoyant plastics may then sink (e.g. Koelmans et al., 2017). These plastics can also become de-fouled in the water column because of foraging and decreasing light intensities, leading to plastics oscillating in the water column (Andrady, 2011; Kooi et al., 2017). Fall-outs of plastic into the water column and the deep sea have recently been detected below the subtropical garbage patch in the North Pacific Ocean (Egger et al., 2020). In the IO, mixing of buoyant plastics in the water column and sinking of initially buoyant plastics have not yet been studied. This section focuses only on the transport of buoyant plastics by ocean surface dynamics in the IO.

Different forcing mechanisms transport buoyant plastics drifting at the ocean surface. Recent reviews by Zhang (2017) and van Sebille et al. (2020) discuss these processes and their influence on the transport of plastics in detail. Ocean surface currents play a dominant role in the transport of buoyant objects, but wind (Breivik and Allen, 2008) and surface gravity

waves (Röhrs et al., 2012) can significantly influence an object's behaviour. Ocean surface currents are forced by many different mechanisms such as wind, waves, tides, and density gradients (Talley et al., 2011; van Sebille et al., 2020). Combined with the Coriolis force, these forcing mechanisms result in Ekman currents, geostrophic currents, and Stokes drift that transport plastics. In addition to the indirect influence of the wind (i.e. forcing ocean surface currents and generating waves), the wind can also have a direct influence the transport of buoyant objects through “windage”, where the wind acts on an object's surface area exposed to air (Richardson, 1997). Currents, wind, and waves have large temporal and spatial variations, both horizontally and vertically. Ocean surface currents typically have a pronounced vertical profile, with current speeds decaying rapidly with depth (Laxague et al., 2017) which can significantly influence on the transport of buoyant objects (van der Mheen et al., 2020b). The different forcing mechanisms that transport buoyant plastics are rarely independent and interact with each other in complex ways. In addition, depending on their specific characteristics (e.g. size, shape, density), buoyant objects can react differently to the same forcing conditions (Maximenko et al., 2012).

4.1 Plastic transport and accumulation along ocean fronts

Plastics entering the ocean from land-based sources are subject to many physical processes on the continental shelf before they are transported into the deeper ocean or beach on coastlines (Fig. 3). Several studies investigating plastic debris in the IO have mainly focussed on the biology and biogeochemistry (Roy et al., 2015; Sarma et al., 2015; Sarkar et al., 2018), and only a few studies have addressed the transport of plastics between inshore and offshore regions. Physical processes that lead to convergent flows, where ocean currents flow towards each other, are one of the most important features for the transport of buoyant plastics. Convergent flows promote downwelling, which causes the accumulation of buoyant plastic debris along the convergent flow boundary defined as the front (Fig. 3b).

Convergent flows occur along ocean fronts, defined as the boundary between two distinct water masses or a region where the rate of change in selected physical properties is much greater than the surrounding areas (Bowman and Esaias, 1977; Belkin and Cornillon, 2007; D'Asaro et al., 2018). Fronts occur from hundreds of metres to many thousand kilometres, and some are short-lived, but most are quasi-stationary and emerge at the same location on seasonal timescales (Belkin and Cornillon, 2007). Ocean fronts are considered hotspots of marine life (Belkin et al., 2009) because flow convergence at fronts channels nutrients towards the fronts and stimulates increased production at different trophic levels (Owen, 1981; Baltar et al., 2016; Sarma et al., 2015; Woodson and Litvin, 2015; Sarkar et al., 2018). Aggre-

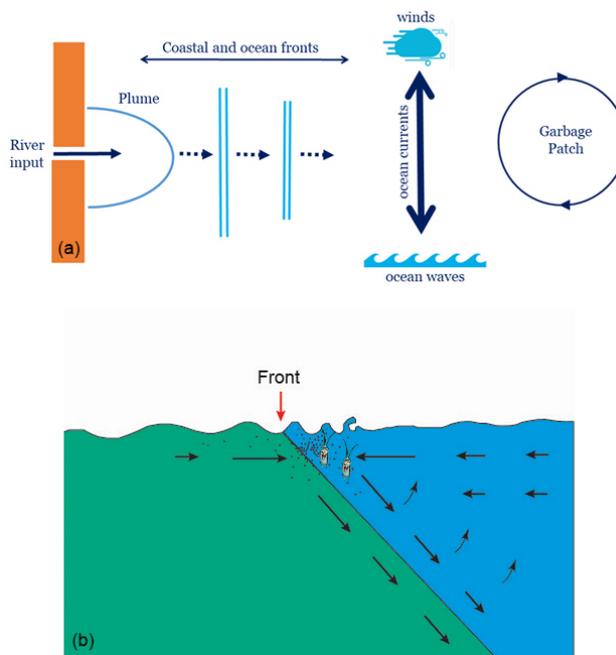


Figure 3. Schematic illustrating the different physical processes that control the transport of buoyant material from rivers to the open ocean (not to scale): (a) bird's-eye view; (b) cross section showing convergence at the front and aggregation of buoyant material. When buoyant material is transported to the open ocean, it is transported by a combination of wind, waves, and ocean currents to the centre of ocean gyres forming the garbage patches.

gations of plankton, larvae, and eggs are often found on the surface (Fig. 3b). Here, as the water sinks at the front due to convergent flow, buoyant material will remain at the surface (Miyao and Isobe, 2016). Predators such as fish and higher-order biota are found above and beneath the front. Due to the accumulation of buoyant debris, including plastics, along fronts, the risk of marine organisms interacting with plastics is high.

There are many types of fronts that are formed through different physical processes. They include river plumes (Luketina and Imberger, 1989; O'Donnell et al., 1998; Karati et al., 2018; Cole et al., 2020), shelf-sea tidal fronts (Simpson and Hunter, 1974; Nahas et al., 2005; Sharples and Simpson, 2020), shelf break fronts (Sharples and Simpson, 2020), upwelling fronts (Brink, 1987), and fronts formed through interaction between flow and topography (Wolanski and Hamner, 1988; Pattiaratchi et al., 1987). Fronts appear as a visible band along the sea surface with differences in temperature and salinity (and, as a result, density) on either side of the front. All these different types of fronts occur in the IO and are extremely important for the accumulation of plastics.

Buoyant plastics entering the IO from land-based sources such as rivers may accumulate at many different frontal systems before being transported offshore (Fig. 3). During the southwest monsoon season (Sect. 4.2), coastal fronts are gen-

erally formed by the interaction of two different water masses moving in opposite directions, where the low-saline runoff waters from rivers are trapped as a salinity front by the high-saline upwelling waters. Examples of such fronts are along the southern coast of Sri Lanka (de Vos et al., 2014). As a result, buoyant plastic debris is trapped in the frontal zone along these water masses (Naidu et al., 2021), referring to locations of future investigations. Naidu et al. (2021) also found that the abundance of plastics, mainly fibres and fragments, was an order of magnitude higher in the frontal zone compared to outside of it off the west coast of India. Similarly, Hajbane et al. (2021) found that close to an offshore reef located in the eastern IO, concentrations of plastics along a coastal front were up to 2 orders of magnitude higher than in surrounding waters. Frontal systems can also be eroded under storm conditions (e.g. Chen et al., 2020), and as a result, plastics accumulated along coastal fronts may beach on coastlines.

4.2 Indian Ocean surface dynamics, plastic transport pathways, and surface accumulation

Compared to other ocean basins, the IO has several unique topographic, atmospheric, and oceanic features that influence the transport of buoyant plastics. To the north, the IO is bounded by the Indian subcontinent, to the west by Africa and the Middle East, to the east by Indonesia and Australia, and in the south by the Southern Ocean. There are two distinctive tropical basins in the northern IO: the Arabian Sea in the west and the Bay of Bengal in the east. The IO is connected to the Pacific Ocean through the Indonesian Archipelago that allows tropical water inflow from the Pacific Ocean. In the southern IO, the eastern boundary formed by the African continent does not extend beyond 35° S, which allows for a connection between the southern IO and the South Atlantic Ocean (Gordon, 2003; Lutjeharms, 2006) and facilitates plastic transport between these oceans leading to debris exchange between the garbage patches (van der Mheen et al., 2019; Sect. 4.3). Because the subtropics in the northern IO are covered by land mass, there is no subtropical gyre. Instead, temperature differences between the northern land mass and the ocean drive the monsoon system, which dominates the atmospheric and oceanic dynamics in the region. These dynamics determine the transport of buoyant plastics in the IO (van der Mheen, 2020) and are described in more detail here.

4.2.1 Northern Indian Ocean surface dynamics, plastic transport pathways, and beaching

The atmospheric dynamics in the IO are characterized by biannually reversing monsoon winds due to seasonal differential heating and cooling of the continental land mass and the ocean (Schott et al., 2009). The southwest (SW) monsoon generally operates between June and October, and

the northeast (NE) monsoon operates between December through April (Tomczak and Godfrey, 2003). The transition periods between the monsoon seasons are the First Inter-Monsoon (May) and Second Inter-Monsoon (November). The influence of the monsoon system is not limited to the northern IO. Unlike in the other oceans, there are no steady equatorial easterly trade winds in the northern IO. Instead, they only have an easterly component during the NE monsoon season and are westerly during the rest of the year. In addition, the SW monsoon season starts with strengthened southeasterly trade winds in the southern IO (Findlater, 1969; Joseph and Sijikumar, 2004).

These strong, seasonally reversing winds drive the ocean surface currents and circulation patterns in the northern IO (Stramma et al., 1996; Shankar and Shetye, 1997; Schott and McCreary, 2001; Shenoi et al., 2004). During the SW monsoon season, the flow in the northern IO is predominantly towards the east (Fig. 4a), from the Arabian Sea into the Bay of Bengal, and the westward-flowing North Equatorial Current (NEC) does not exist (Schott et al., 2009). Along the coastlines of India and Sri Lanka in the Arabian Sea, the West Indian Coastal Current (WICC) flows southwards along the western Indian coastline and joins the eastward-flowing Southwest Monsoon Current (SMC). The SMC flows from the Arabian Sea past Sri Lanka and into the Bay of Bengal (de Vos et al., 2014). After passing the coast of Sri Lanka, the ocean surface currents form an anticlockwise eddy called the Sri Lanka Dome (SLD; Su et al., 2021). The western arm of this eddy drives a southward current along the eastern coast of Sri Lanka; the remainder flows northwards along the eastern Indian coastline as the East Indian Coastal Current (EICC). At the eastern boundary, the South Java Current (SJC) is variable but flows predominantly to the northwest along the Java coast (Sprintall et al., 2009), in the same direction as the Indonesian Throughflow (ITF), which is strongest during the SW monsoon (Sprintall et al., 2009). At the western boundary of the northern IO, the Somali Current (SC) flows northeastwards during this season.

During the NE monsoon season, the ocean surface currents reverse direction and the flow in the northern IO is predominantly towards the west (Fig. 4b), from the Bay of Bengal towards the Arabian Sea (Schott et al., 2009). Both the WICC and the EICC along the Indian coasts reverse direction and the SMC reverses and becomes the Northeast Monsoon Current (NMC), which is weaker than the SMC (de Vos et al., 2014). Together with the East African Coastal Current (EACC), the SC supplies the eastward South Equatorial Counter Current (SECC). The SECC feeds into the SJC, which flows southeastwards along the coasts of Sumatra and Java during this season. Another unique feature in the equatorial IO are the strong eastward-flowing Wyrki jet (Wyrki, 1973) that develop along the Equator during the First and Second Inter-Monsoon. These jets are at their strongest during the Second Inter-Monsoon (Qui and Yu, 2009).

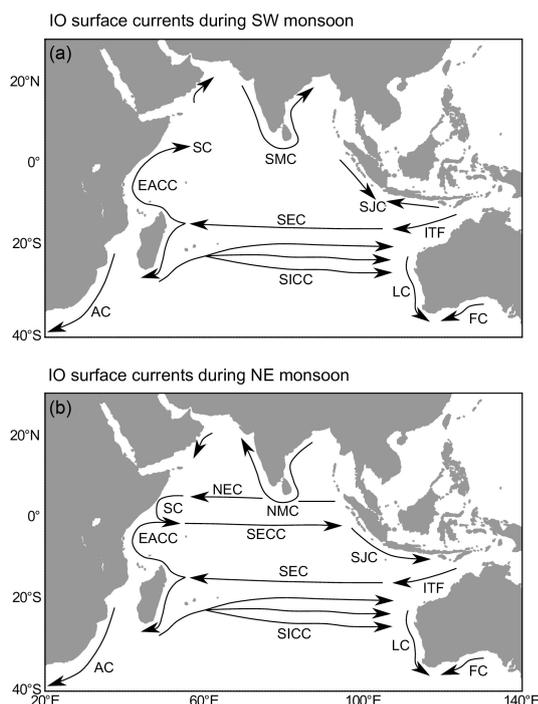


Figure 4. Schematic of ocean surface currents in the Indian Ocean based on Schott et al. (2009), during (a) the SW monsoon season and (b) the NE monsoon season. The following currents are shown and labelled with abbreviations: Southwest Monsoon Current (SMC) and Northeast Monsoon Current (NMC); West Indian Coastal Current (WICC) and East Indian Coastal Current (EICC); Sri Lanka Dome (SLD); South Java Current (SJC); Indonesian Throughflow (ITF); Somali Current (SC); East African Coastal Current (EACC); North Equatorial Current (NEC); South Equatorial Counter Current (SECC); South Equatorial Current (SEC); Agulhas Current (AC); Leeuwin Current (LC); Flinders Current (FC); and South Indian Counter Current (SICC). The numbers in (a) refer to marginal seas (1: Arabian Sea; 2: Bay of Bengal) and countries listed in the text: 3: India; 4: Sri Lanka; 5: Somalia; 6: Madagascar; 7: Sri Lanka; and, 8: Sumatra (Indonesia).

The main features to note from the description of the ocean surface currents in the northern IO are the reversal of major currents with the monsoon season. Particle tracking simulations illustrate that these reversing currents transport buoyant objects between the eastern and western section of the northern IO with Sri Lanka and the Maldives in the central section (Fig. 5; van der Mheen, 2020). In addition, the net surface transport in the northern IO is eastwards (Schott et al., 2009) because of the strong eastwards SMC and the eastward-flowing Wyrki jets between monsoon seasons. This net eastward transport, combined with eddies developing year-round in the Bay of Bengal and potentially trapping plastics, results in plastics being present in the Bay of Bengal throughout the year. In contrast, the Arabian Sea is mostly depleted of plastics during the SW monsoon season (Fig. 5). However, if sources of plastics increase, it is likely that concentrations

will also increase. As plastics are transported back and forth between the Arabian Sea and the Bay of Bengal, they frequently come close to coastlines. Because of this, it is likely that a large amount of plastic beaches and accumulates on northern IO coastlines (van der Mheen, 2020). Simulations by van der Mheen (2020) showed that shorelines of countries bordering the Bay of Bengal were most affected by beaching plastics (Fig. 6). This is most likely due to the large source locations of particles (Fig. 1b) and the ocean dynamics of the Bay of Bengal, resulting in plastics being present in the Bay throughout the year.

The dynamics of beaching plastics are complex and strongly influenced by small-scale coastal ocean dynamics (Isobe et al., 2014), as well as local morphology of the coastline (Zhang, 2017). It is unknown what the influence is of these small-scale effects on basin-scale beaching patterns in the IO. Finally, plastics do not necessarily remain beached indefinitely but can also re-float and re-enter the ocean (Zhang, 2017; Lebreton et al., 2019). Several recent studies highlight the potential of oceanic islands to act as transitory repositories for plastic debris (Monteiro et al., 2018; Pham et al., 2020). As a result, it is unknown how much plastic is stored on coastlines in the IO or how permanent this sink is. Thus, long-term (multi-decadal) monitoring and field studies are necessary along IO coastlines.

4.2.2 Southern Indian Ocean surface dynamics and the subtropical garbage patch

In the southern IO, similar to other ocean basins, there is a wind-driven subtropical gyre (Fig. 4) which in this instance has several unique features. The gyre is bounded in the north by the South Equatorial Current (SEC), which flows westwards and is relatively steady all year round. At the western boundary of the subtropical gyre, the Agulhas Current (AC) flows poleward along the southern African coastline (Beal et al., 2011). The Agulhas Retroflection (AR) and Agulhas Leakage (AL) provide a connection between the southern IO and South Atlantic Ocean (Gordon, 2003; Lutjeharms, 2006). At the eastern boundary, the Leeuwin Current (LC) flows poleward along the western Australian coastline. This is opposite to the equatorward flow expected for a wind-driven subtropical gyre (Pattiaratchi and Woo, 2009). The Flinders Current (FC; Middleton and Cirano, 2002) flows westward from the Pacific Ocean to the southern IO along the southern Australian coastline. In the south, the gyre is bounded by the Antarctic Circumpolar Current (ACC). The South Indian Counter Current (SICC) flows eastward through the centre of the subtropical gyre (Lambert et al., 2016), which is the opposite of the direction expected from Sverdrup theory (Palastanga et al., 2007; Wijeratne et al., 2018). Similar subtropical counter-currents exist in the other oceans but the SICC is unique because it flows across the full width of the basin and splits into three separate branches as it flows be-

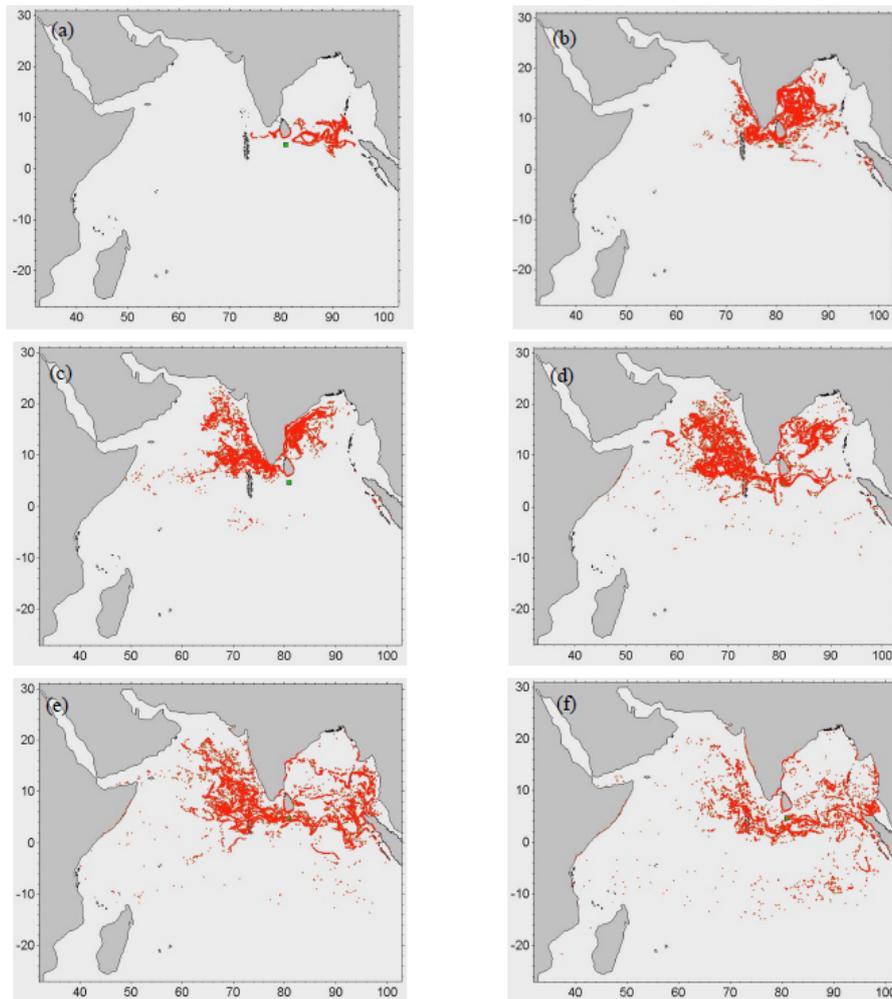


Figure 5. Results of a particle tracking simulation using Ichthyop in the northern IO. In total, 10 000 passive particles were released on 1 September to the south of Sri Lanka (green square) and tracked using hourly surface current output from the HYCOM model to illustrate the influence of the northern IO surface dynamics on the transport of buoyant plastics. Note that the release location does not represent a location where plastics enter the ocean but instead was chosen as a central location that is influenced by the reversal of monsoonal currents. (a) 1 November; (b) 1 January; (c) 1 March; (d) 1 May; (e) 1 July; (f) 1 September.

tween the southern tip of Madagascar and the western coast of Australia (Menezes et al., 2014).

Most plastic waste enters the southern IO from the Indonesian Archipelago (Jambeck et al., 2015; Lebreton et al., 2017; Schmidt et al., 2017, 2018; Meijer et al., 2021; Sect. 2). These plastics are transported westwards across the IO basin by the SEC and into the subtropical southern IO (van der Mheen et al., 2020b). In the subtropics, converging wind-driven Ekman currents lead to downwelling and associated accumulation of buoyant plastics in a subtropical garbage patch. However, because of the unique features of the southern IO subtropical gyre, the dynamics of the garbage patch in the IO are distinctive. The garbage patch is very sensitive to different transport mechanisms (van der Mheen et al., 2019). Van der Mheen et al. (2019) showed that, under the influence of Stokes drift and wind, the subtropical garbage

patch centres on the west of the IO basin, and simulated concentrations are at least a factor 2 smaller than in the garbage patches in the other oceans. In contrast, if no Stokes drift or wind is included in the simulation, a large and stable garbage patch forms that spans almost the entire IO basin width. In this case, the simulated concentrations are of the same order as concentrations in the North Pacific subtropical garbage patch. It is, however, unknown which forcing mechanisms predominantly transport plastics. As a result, it is unclear which of these simulated garbage patches best represents the reality in the subtropical southern IO. Sampling in this region is too scarce to provide any conclusions (see Sect. 3). More measurements are required to determine the dynamics and concentration of plastics in the subtropical IO garbage patch.

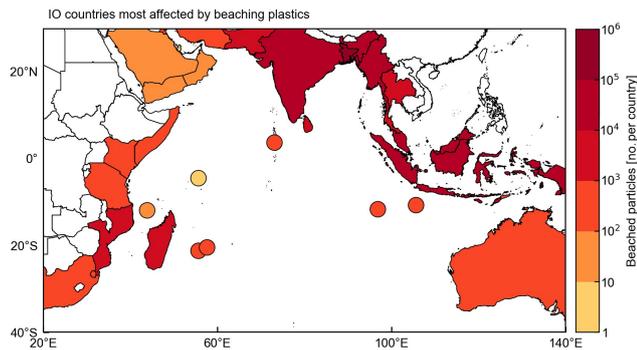


Figure 6. Countries and islands in the Indian Ocean most affected by beaching plastics from river sources. Colours show the number of particles that have beached in each country in Lagrangian particle tracking simulations by van der Mheen et al. (2020a), where particles beach with a 50 % chance if they come within 8 km of a coastline. Figure adapted from van der Mheen et al. (2020a). Please note that country boundaries are obtained from outputs of scientific computer simulations and may not represent political and/or geographical realities.

Plastics can potentially also be transported from the subtropical IO garbage patch into the southern Atlantic and Pacific oceans. Van der Mheen et al. (2019) and Dobler et al. (2019) showed that plastics transported by Stokes drift and/or wind can move from the IO garbage patch past South Africa into the South Atlantic Ocean. Maes et al. (2018) suggested that there is also a “super convergence pathway” connecting the southern IO to the South Pacific Ocean. Their particle tracking simulation results showed particles being transported eastwards close to the southern Australian coastline. However, these results are potentially in contradiction to the westward-flowing FC in this region (Middleton and Cirano, 2002; Wijeratne et al., 2018), and so the existence of a super convergence pathway between the southern IO and the South Pacific Ocean along the southern Australian coast still needs further investigation.

Finally, plastics can also beach in the southern IO. Countries along the eastern African coast, Madagascar, Mauritius, Réunion, the Cocos (Keeling) Islands, and Christmas Island are potentially heavily affected by beaching plastics (Fig. 6, van der Mheen, 2020). In the simulations by van der Mheen (2020), most of these plastics originated from rivers in Indonesia.

4.2.3 Transport of plastics between the northern and southern Indian Ocean

Strong ocean surface currents can act as transport barriers for buoyant objects, preventing buoyant objects from crossing these currents (Bower, 1991; Brambilla and Talley, 2006). Due to strong equatorial currents, ocean surface drifters do not tend to cross the Equator (Maximenko et al., 2012). Therefore, it has been suggested that buoyant plastics gener-

ally remain in the hemisphere where they originally entered the ocean (Lebreton et al., 2012). However, as previously mentioned, the easterly trade winds are not steady in the IO and as a result, the NEC and SECC are not steady either. The surface waters of the IO also appear more connected between hemispheres than the other oceans (Froyland et al., 2014). Therefore, van der Mheen (2020) suggested that plastics may not remain in their original hemisphere in the IO. Particle tracking simulation results by van der Mheen (2020) showed that buoyant plastics could cross from the northern IO into the southern IO as they are transported by the SJC along the Sumatran coastline (see an example of this happening in Fig. 5f). This mainly occurred during the Second Inter-Monsoon in their simulations.

Van der Mheen et al. (2020b) confirmed that ocean surface drifters mainly cross from the northern IO into the southern IO on the eastern side of the IO basin, along the Sumatran coastline. Drifters mainly crossed during the Second Inter-Monsoon and the NE monsoon season. Van der Mheen et al. (2020) also showed that ocean surface drifters crossing from the southern IO into the northern IO do so along the western side of the basin, predominantly during the SW monsoon season. Drifters cross the Equator along the Somalian coast, likely as they are transported by the SC, which is directed northeastwards during the SW monsoon.

Based on these results, it seems possible for buoyant plastics to cross the Equator in the IO. It is, however unclear how frequently this occurs.

4.3 Sinking and ingestion

4.3.1 Sinking

Sinking and settling of plastics on the seafloor due to fragmentation and biofouling may be a major sink of plastic debris in the ocean (Koelmans et al., 2017). Based on deep-sea sediment core samples between 500–1000 m depth in the southwest IO, Woodall et al. (2014) estimated that 4 billion fibres per square kilometre were present in the IO but did not report a mass estimate. Ingested plastics by deep-sea fauna in the IO (Taylor et al., 2016) are also evidence that plastics sink to the seafloor. However, no evidence of the total size of this sink currently exists, and the understanding of the exact processes of biofouling, fragmentation, and sinking as well as the timescales on which these occur is limited.

However, the IO is one of the most productive regions in the global oceans due to intense upwelling during the southwest monsoon (Qasim, 1977). This high surface productivity results in a high export flux of organic particles from the euphotic zone to the deep sea (Ittekkot et al., 1996; Gupta et al., 1997). As a result of this high productivity, biofouling of plastic debris may occur rapidly in the IO. As a result, sinking of plastics due to biofouling may be particularly relevant in the IO.

4.3.2 Ingestion

Ingestion of plastics can occur at the ocean surface, in the water column, and on the seafloor. Estimates of plastic ingestion by vertebrates (van Franeker, 2011; Davison and Asch, 2011) indicate that the global ingestion of plastics could be on the same order of magnitude as the amount of plastics accumulating in subtropical garbage patches (van Sebille et al., 2015). However, plastic ingestion is generally considered only a temporary and not a permanent sink of marine plastic debris.

Throughout the IO (Fig. 2b), multiple studies have sampled ingested plastics in a variety of different fauna: benthic invertebrates (Taylor et al., 2016; Naidu et al., 2018), sessile invertebrates (Thushari et al., 2017), fishes (Ismail et al., 2018; Karthik et al., 2018; Baalkhuyar et al., 2018; Crutchett et al., 2020; McGregor and Strydom, 2020; Sparks et al., 2020), including large sharks (Cliff et al., 2002), seabirds (Cherel et al., 2017; Cartraud et al., 2019), turtles (Hoarau et al., 2014), bivalves (Naidu, 2019), and corals (Saliu et al., 2019; Patti et al., 2020). Recorded ingestion rates varied widely between species, from only approximately 0.4 % of large sharks sampled (Cliff et al., 2002) to up to 90 % of fish sampled (Sparks and Immelman, 2020).

These sampling studies are both relatively few and relatively recent, so no estimates can be given about the total amount of plastic ingested by marine fauna in the IO or any trends in plastic ingestion. Cherel et al. (2017) did find that the wandering albatross chicks they investigated at Crozet and Kerguelen Islands had ingested low plastic loads compared to albatross chicks in the North Pacific Ocean. Crutchett et al. (2020) found low plastic ingestion levels in sardines compared to global levels. They also suggested that sampling plastics in globally common fishes, such as sardines, is a good way to compare and monitor ingestion rates between different locations around the world. In contrast, Hoarau et al. (2014) found that turtles in the southwest IO had a similar occurrence of ingested plastic as in the Mediterranean, southwest Pacific, and Gulf of Mexico but a much higher amount of ingested plastic by number, volume, and weight.

5 Impacts

Plastic debris can harm the marine environment, communities, and economies in many different ways. The review by Law (2017) discusses the different harmful effects of plastic debris as well as existing evidence for their impact in detail. This section briefly discusses the harmful effects of plastic debris and provides more detail about the potential impacts specific to the IO and IO rim countries. In Sect. 6.4, we also provide an overview of the different policies and initiatives that are emerging in IO rim countries to prevent, mitigate, or clean up plastic debris in the IO.

5.1 Entanglement

Approximately 100×10^6 t of marine fishes have been landed globally each year since 2000, approximately 13 % of which originated from the IO (Pauly and Zeller 2016). The IO supports a range of marine ecosystem services including numerous and expansive fishing ventures with 7 %, or 1×10^6 t, of IO landings considered to be large commercial pelagic fishes, e.g. tuna and billfishes. The commercial pelagic fisheries of the IO alone are worth over USD 1300 million annually. ALDFG becomes “ghost gear”, which continues to entangle and injure or catch wildlife, for decades or centuries as it slowly degrades, increasing the pressure on marine wildlife. Entangled and caught fauna of ALDFG are initially sea turtles, marine mammals, sharks, and large predators for the first few months, later including crustaceans when gear fragments into smaller parts (Macfadyen et al., 2009; Wilcox et al., 2015; Stelfox et al., 2016).

ALDFG affects the tourism industry through a multitude of factors, including the removal of iconic marine species and wildlife in general and contributes up to 90 % of shoreline debris and degrading of the perceived beauty of an area (Gunn et al., 2010; Lachmann et al., 2017). Ghost gear is also considered to be destructive to both natural habitats, i.e. coral reefs, and manufactured objects, i.e. boats (Laist and Liffman, 2000; Gunn et al., 2010). These devaluations decrease benefits from coral reef tourism (USD 36 billion) (Spalding et al., 2017) and global exports of aquarium fish (USD 15 billion) (Raja et al., 2019). Data from genetic analyses of Olive Ridley turtles entangled in ghost nets in the Maldives showed that the individual turtles originated from populations nesting in India and Sri Lanka (Stelfox et al., 2020b). This shows that impacts on charismatic marine species that drive tourism can simultaneously impact multiple economies in the IO rim. Whilst regional data are difficult to obtain, 60 % of aquarium fish are exported from developing nations (Raja et al., 2019). Thus, the adverse cultural, societal, and sustenance impacts of ghost fishing may disproportionately affect IO rim developing nations (Wong, 2011; Guillotreau et al., 2012; Lachmann et al., 2017).

Many of the countries bordering the IO are small island developing states (SIDSs) where more than 70 % of the population lives along the coast and is highly dependent on marine ecosystem services (Canales et al., 2017). Recent interviews of fishers by Richardson et al. (2021), which included fishers from Indonesia along the IO rim, showed that the main reasons for gear loss reported were bad weather and interactions with wild life, respectively. Illegal and deliberate gear discard on the other hand was reportedly low. Furthermore, over half of the fishermen interviewed across the world reported being “concerned” or “very concerned” about ALDFG, whereby economic losses scored highest (54 %) as an issue of concern followed by environmental harm (41 %). The reported loss prevention strategies that scored highest were gear maintenance and training crew in gear management, which provide

clear avenues for targeted programmes to educate and raise awareness around ALDFG in low-income fisheries, such as in many IO rim economies (Richardson et al., 2021). A single ghost net may decrease fishing revenues by USD 20 000 annually (Lachmann et al., 2017) and can consume as much as 30 % of landings within a single fishery (Laist and Liffman, 2000). Although ALDFG is thought to diminish in catching efficiency over time, the catching potential of gill nets was estimated to be 20 %–30 % of original levels for long periods and at least 5 % after more than 2 years of abandonment (Macfadyen et al., 2009; Stelfox et al., 2016). Some catch rates were estimated to be 81 kg d⁻¹ for tangle nets over 1.5 years, an average of 64 kg d⁻¹ for deep-water gill nets over 45 d and 0.6 kg d⁻¹ for traps left for 3–6 months (Macfadyen et al., 2009).

5.2 Ingestion

As described in Sect. 4.3.2, multiple studies in the IO have found ingested plastic in different marine species. However, as described in the review by Law (2017), confirmed cases of ingestion by themselves do not determine the impact of plastic debris ingestion. Even confirmed cases of individual deaths are not readily translated to impact on species or ecosystem levels. Recently, Roman et al. (2021) determined which types of plastic debris contribute the most to mortality in marine megafauna (cetaceans, pinnipeds, sea turtles, and seabirds). In the IO plastic ingestion has been studied in sea turtles (Hoarau et al., 2014) and seabirds (Cherel et al., 2017; Cartraud et al., 2019).

For turtles, Roman et al. (2021) concluded that film-like plastics, plastic fragments, and hard plastics were most likely to cause death. Although Hoarau et al. (2014) did not find a relationship between the mortality of loggerhead turtles in the southwest IO and the type of plastic debris they had ingested, they did find a correlation between the length of debris and mortality. They also suggested that loggerhead turtles have a relatively high tolerance to the ingestion of plastic debris, only 9 % of approximately 450 dead turtles were thought to have been killed directly by the ingestion of plastic.

Roman et al. (2021) found that ingestion of hard plastics was most likely to cause death for sea birds. Cherel et al. (2017) sampled the stomach contents of live wandering albatross chicks on the Kerguelen and Crozet Islands in the IO. They found that approximately 50 % of chicks had ingested plastic fragments. However, the plastic content was relatively low and Cherel et al. (2017) suggested that this was unlikely to cause any significant deleterious effects. On Réunion and Juan de Nova, Cartraud et al. (2019) determined the plastic ingestion of deceased petrels and shearwaters. They could find no correlation between the muscular condition of the birds and the amount of plastic ingested. However, all of the birds likely died as a result of injury after being attracted and disoriented by urban light.

6 Emerging policies and initiatives on plastic in Indian Ocean rim countries

Of the top 20 countries ranked by mass in estimated mismanaged plastic waste (MMPW), nine are located along the IO rim (Indonesia, India, Thailand, Malaysia, Bangladesh, South Africa, India, Pakistan, and Myanmar). Plastic waste generated by these countries amounts to ~ 15 % of the world's total MMPW (Jambeck et al., 2015), putting pressure on these countries to address the issue of MMPW. Local and national actions have been the primary approach for mitigating plastic pollution (Vince and Hardesty, 2016). The choice of measures at a national or local level is left to the IO rim country administrations – in line with the principle of subsidiarity. For example, some countries have refundable deposit schemes for bottles (Schuyler et al., 2018). Targeted deposit schemes can help reduce littering and boost recycling and have already helped several countries achieve high collection rates for beverage containers (Lavee, 2010; Dace et al., 2013; Schuyler et al., 2018).

Several African countries have introduced measures to address plastic bag pollution. In August 2017, Kenya introduced a total ban on plastic bags, banning all plastic bags' use, manufacture, and importation for commercial and household packaging. Anyone found in violation is subject to a fine of approximately USD 20 000–40 000 and/or 1 to 4 years imprisonment, making this ban the toughest in the world. Inland countries such as Botswana, introduced a levy on plastic bags in 2010, while Eritrea banned plastic bags in 2005. Rwanda banned plastic bags in 2008 as part of its Vision 2020 plan for sustainability. In 2007, Uganda introduced a ban of lightweight plastic bags which came into effect that year but was never implemented. Tanzania introduced a ban in 2006 and in South Africa, a bag levy was introduced in 2004, although they were never banned completely.

In India, plastic pollution is being fought at various levels, by state governments, NGOs, and individuals. Initiatives range from the banning of plastic to beach clean-ups. Several Indian states have banned or regulated the use of plastic, but India still struggles to manage its huge plastic waste. For microplastics and their presence in consumer goods, regulations are still under development. The National Green Tribunal in January 2017 asked the Union Government to test leading cosmetics brands for microplastics. The Bureau of Indian Standards (BIS) classified microbeads as “unsafe” for use in cosmetic products and banned microbeads in cosmetics in October 2017, but this was only implemented in 2020.

7 Case study: the *X-Press Pearl* nurdle spill off Sri Lanka

A container ship, the *X-Press Pearl*, registered in Singapore, was transporting cargo from Jebel Ali (United Arab Emirates) to Colombo (Sri Lanka) via Hamad Port (Qatar)

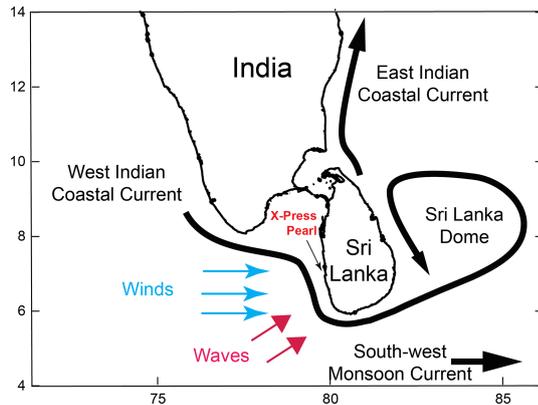


Figure 7. Location of the *X-Press Pearl* and the major currents in the region. The wind and wave directions experienced during the incident are also shown.



Figure 8. Image of nurdles on the beach directly to the east of the spill from the *X-Press Pearl* (location in Fig. 7).

and Hazira (India). The ship departed the port of Hazira on 15 May 2021 and arrived in Colombo on 19 May and was anchored off Colombo Port ~ 9.5 km offshore in a water depth of 21 m (Fig. 7). On 22 May, the sound of an explosion was heard in cargo hold no. 2 and the ship was then seen to be on fire. By 24 May, the fire had intensified and was spreading toward the aft of the vessel. A louder explosion was heard on 25 May and all personnel were evacuated. The fire burned for 13 d. On 2 June, efforts to move the ship into deeper waters failed, with the aft portion sinking to the seabed. The vessel was transporting 1486 containers, with a mixed variety of cargo, but included 78 t of plastic nurdles (low-density polyethylene pellets) in three containers released into the ocean.

7.1 Oceanographic setting

Ocean currents around Sri Lanka during May are such that they flow south along the west coast of India (West Indian coastal current; Fig. 4), cross the Gulf of Mannar, and flow south and east along the west and south coasts of Sri Lanka, respectively (see also de Vos et al., 2014; Su et al., 2021).

There is a recirculation, the SLD, to the east of the island (Figs. 4 and 7). The SLD connects the water from the west coast to the east coast, with currents flowing south along both coasts (Fig. 7). During the initial period of the incident, there were strong onshore winds ($> 10 \text{ m s}^{-1}$) and swell waves (height ~ 2 m). This was associated with the beginning of the southwest monsoon and the currents were flowing to the south along the west coast of Sri Lanka at the location of the accident. On 22 May, tropical cyclone Yaas formed in the northern Bay of Bengal and propagated to the north and crossing the coast to the north of Visakhapatnam (India). Examination of the water level records along the east coast of India and at Trincomalee (Sri Lanka) indicated the formation of a continental shelf wave (Huyer, 1990) subsequent to the cyclone making landfall (Eliot and Pattiaratchi, 2010) and propagating southward and clockwise around Sri Lanka. As a consequence of the continental shelf wave, the sub-tidal water level at Trincomalee started to increase whilst the mean water level at Colombo was decreasing. This resulted in the currents reversing from that due to the monsoon and flowing east and north along the south and west coasts of Sri Lanka, respectively, after 30 May (Fig. 2). The winds decreased to $< 5 \text{ m s}^{-1}$ after 2 June.

7.2 Impact

Nurdles are small plastic pellets used as raw materials for the manufacture of virtually anything that is plastic (e.g. from plastic bags and bottles to automobile parts) and are by definition classified as microplastics as their size is < 5 mm. For comparison, the potential discharge of 78 t of nurdles released by the *X-Press Pearl* is the second-worst release of nurdles from shipping accidents. This includes the 150 t released into Hong Kong harbour in 2012 during a typhoon, and 49 t released into Durban harbour in South Africa in 2017 (Schumann et al., 2019). On 25 May, large quantities of nurdles were washed up on the beaches closest to where the *X-Press Pearl* was anchored (Fig. 8). These were transported onshore because of the strong winds and high waves present at that time. Subsequently nurdles were found along the whole of the west of Sri Lanka (<https://storymaps.arcgis.com/stories/054f6746857242128298d53d220d76f0>, last access: 28 December 2021).

We used the Regional Ocean Modelling System (ROMS) configured to the oceans around Sri Lanka (de Vos et al., 2014; Su et al., 2021) to provide surface currents together with the Lagrangian particle tracking model Ichthyop (Lett et al., 2008) to simulate the transport of buoyant nurdles along the coast of Sri Lanka (Fig. 9). The model results indicated movement of the nurdle plume southwards with the prevailing currents and by 27 May the plume had extended to the southwest corner of the Island (Fig. 9b). On 29–31 May, the nurdle plume had detached from the coast and was moving offshore to the southeast (Fig. 9c and d). The currents off the southeast of the island was getting stronger and were flow-

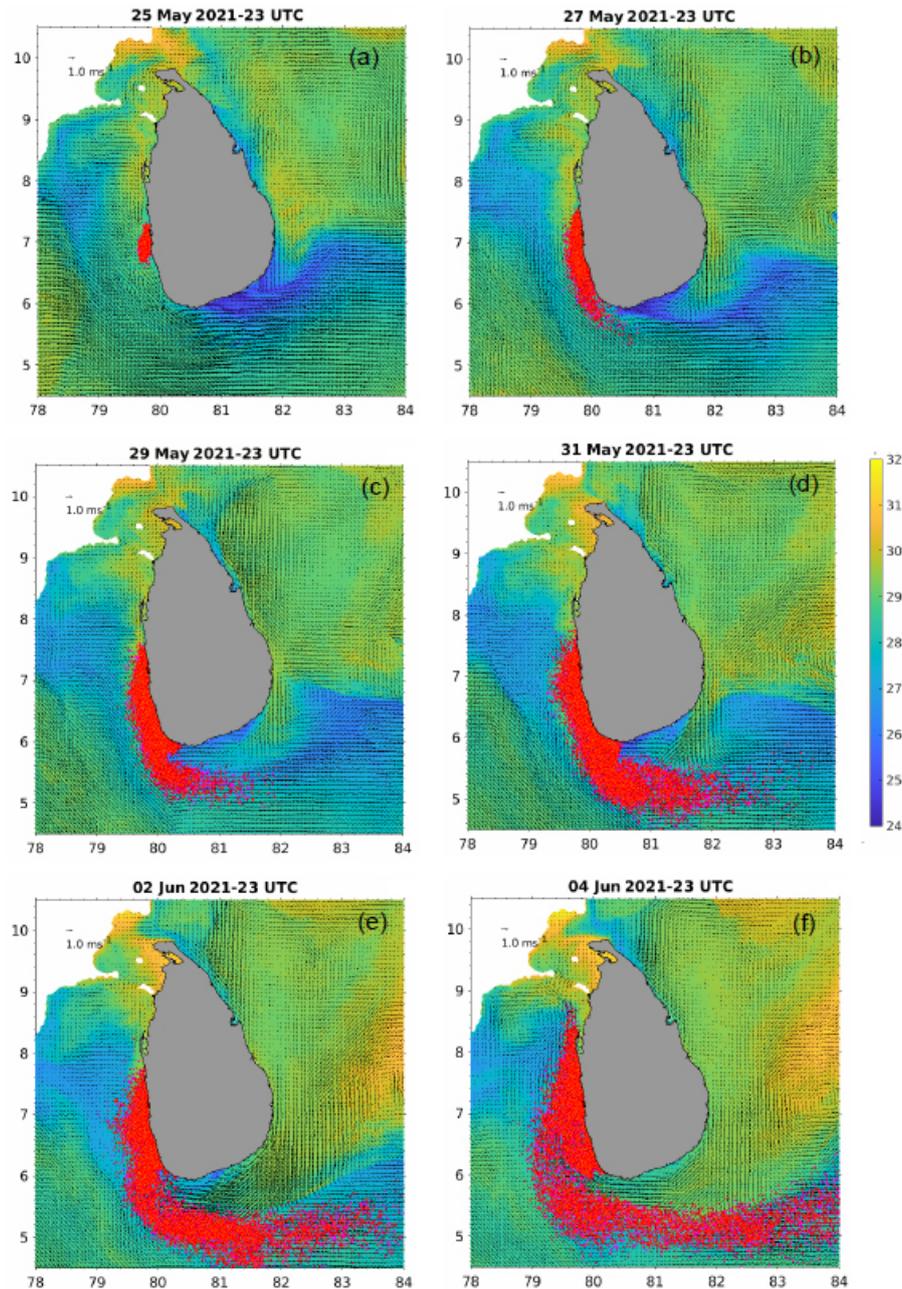


Figure 9. Time series of the simulated pathway of nurdles from 25 May to 4 June 2021 at 48 h intervals. The nurdles are shown as red dots, colours are sea surface temperature ($^{\circ}\text{C}$), and arrows represent surface velocity vectors.

ing to the south. By 2 June, this current has further strengthened and has curved around the island flowing to the east (Fig. 9e). By 4 June the currents along the west coast has reversed with the nurdle plume closest to the coast moving north (Fig. 9f). The nurdles that were detached from the coast on 29 May continued to be transported offshore and subsequently to the east. A budget of the simulated nurdles indicated that on 11 June 32 % of the nurdles released would have beached, 28 % would have been within the model domain

(Fig. 9), and 40 % would have exited the model domain. A feature of the simulations is that the nurdles were contained at distinct boundaries representing the fronts separating cold and warm water as described in Sect. 4.2. Also the reversal of the surface currents on 2 June and transport northwards along the Sri Lanka coastline was associated with the continental shelf wave that was due to tropical cyclone Yaas (see Sect. 7.1). With time, the nurdles that leave the model domain will make landfall along the countries in the northern Indian

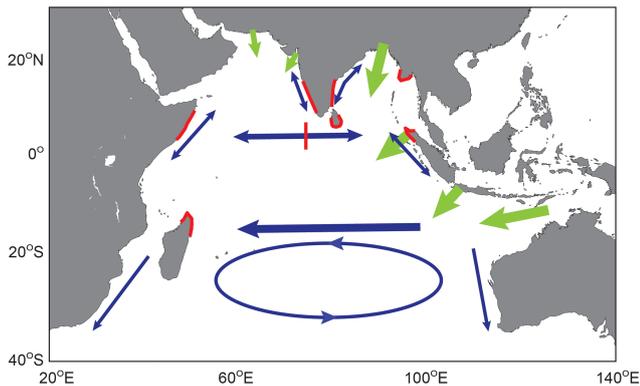


Figure 10. Schematic showing major plastic sources (green arrows), plastic transport pathways (blue arrows), and major beaching locations of plastics (red regions) in the Indian Ocean.

Ocean (e.g. Indonesia, India, Maldives, Somalia) because of the reversing monsoon currents in the region (see Fig. 5) and will be a visible pollutant on the beaches for many decades to come.

8 Summary

The Indian Ocean covers $\sim 20\%$ of the Earth's surface and, compared to other oceans has unique physical characteristics (northern landmass, connection to the Pacific Ocean through Indonesia and connection to the Atlantic Ocean past South Africa, Monsoon) that strongly influence the circulation patterns and therefore the fate of plastics. This paper examined the sources, transport, sinks, and impacts of plastics in the IO. The main sources of plastics are rivers (terrestrial sources), with the majority input in the northern Indian Ocean, particularly the Bay of Bengal (Fig. 10). There are also land-based sources from Indonesia into the southern Indian Ocean. Inputs from Australia are negligible due to the absence of rivers and low population density in Western Australia. In the northern IO, reversing monsoon-driven currents transport plastic material between the Bay of Bengal and the Arabian Sea with many plastics making landfall in Somalia, the Maldives, Sri Lanka, the east and west coasts of India, Myanmar, and western Sumatra (Fig. 10). There is no garbage patch located in the northern IO; however, in the southern IO there is evidence for a garbage patch, but it is not well defined, with very few measurements indicating that it may extend across the whole length of the ocean basin. The main beaching region in the southern IO is the coast of northern Madagascar. There could also be leakage of material into the southern Atlantic Ocean garbage patch past South Africa.

Knowledge gaps

The main gap in knowledge for the IO is the scarcity of data in both the surface and deeper ocean. Only a few data points

are available for the whole IO including along coastlines and the deeper ocean (Fig. 2). This is a very large constraint, which needs addressing. As demonstrated in this paper, we can simulate the transport of plastics using numerical models, but confirmation of these results is of paramount importance. However, the simulations pinpoint regions that require attention in future investigations. For example, it is unclear which type of simulated garbage patches (leaky with low concentrations and on the western side of the IO basin or stable with high concentrations spanning the entire width of the IO basin) best represents the behaviour of the garbage patch in the subtropical southern IO. For the same reasons, it is also unknown how much plastic is contained in the subtropical IO garbage patch as well as how long it is likely to remain in this garbage patch (both as a result of escaping the IO into the South Atlantic Ocean and as a result of other factors such as sinking and ingestion of plastics).

This paper's main sources of plastics were derived from rivers mainly in the northern Indian Ocean. There is most likely transport of plastics from southeast Asia through the Indonesian Throughflow (Sect. 4.1). However, these sources are currently undocumented and need to be investigated. Individual events such as the *X-Press Pearl* ship incident off Sri Lanka are also responsible for large input of plastics.

Considering the importance of increasing plastic pollution, the dynamic pathways and their fate in the marine environment of the Indian Ocean needs further attention. Further, studies on the ingestion of plastic particles by marine biota and their residence time in seafood in the marine environment will also be useful for food quality and the ecosystem's overall health.

The understanding of the exact processes of biofouling, fragmentation, and sinking as well as the timescales on which these occur is limited. As far as we are aware, no studies have focussed on this in relation to the IO.

Code availability. The model code and the output of the ROMS runs presented in this article are available from the first author on request.

Data availability. No data sets were used in this article.

Author contributions. CP and MvdM developed the conceptualization and CP provided the resources. The original draft was prepared by CP and MvdM with contributions from all the authors. SW undertook the numerical simulations of nurdles originating from the *X-Press Pearl* incident. Review and editing were undertaken by all the co-authors, who have agreed on the published version of the paper.

Competing interests. The contact author has declared that neither they nor their co-authors have any competing interests.

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References

- Abayomi, O. A., Range, P., Al-Ghouti, M. A., Obbard, J. P., Almeer, S. H., and Ben-Hamadou, R.: Microplastics in coastal environments of the Arabian Gulf, *Mar. Pollut. Bull.*, 124, 181–188, <https://doi.org/10.1016/j.marpolbul.2017.07.011>, 2017.
- Abbasi, S., Keshavarzi, B., Moore, F., Turner, A., Kelly, F. J., Dominguez, A. O., and Jaafarzadeh, N.: Distribution and potential health impacts of microplastics and microrubbers in air and street dusts from Asaluyeh County, Iran, *Environ. Pollut.*, 244, 153–164, <https://doi.org/10.1016/j.envpol.2018.10.039>, 2019.
- Akhbarizadeh, R., Moore, F., Keshavarzi, B., and Moeinpour, A.: Microplastics and potentially toxic elements in coastal sediments of Iran's main oil terminal (Khark Island), *Environ. Pollut.*, 220, 720–731, <https://doi.org/10.1016/j.envpol.2016.10.038>, 2017.
- Aliabad, M. K., Nassiri, M., and Kor, K.: Microplastics in the surface seawaters of Chabahar Bay, Gulf of Oman (Makran coasts), *Mar. Pollut. Bull.*, 143, 125–133, <https://doi.org/10.1016/j.marpolbul.2019.04.037>, 2019.
- Andrady, A. L.: Microplastics in the marine environment, *Mar. Pollut. Bull.*, 62, 1596–1605, <https://doi.org/10.1016/j.marpolbul.2011.05.030>, 2011.
- Arun Kumar, A., Sivakumar, R., Sai Rutwik Reddy, Y., Bhagya Raja, M. V., Nishanth, T., and Revanth, V.: Preliminary study on marine debris pollution along Marina beach, Chennai, India, *Reg. Stud. Mar. Sci.*, 5, 35–40, <https://doi.org/10.1016/j.rsma.2016.01.002>, 2016.
- Baalkhuyur, F. M., Bin Dohaish, E. J. A., Elhalwagy, M. E. A., Alikunhi, N. M., AlSuwailam, A. M., Røstad, A., Coker, D. J., Berumen, M. L., and Duarte, C. M.: Microplastic in the gastrointestinal tract of fishes along the Saudi Arabian Red Sea coast, *Mar. Pollut. Bull.*, 131, 407–415, <https://doi.org/10.1016/j.marpolbul.2018.04.040>, 2018.
- Baltar, F., Currie, K., Stuck, E., Roosa, S., and Morales, S. E.: Oceanic fronts: transition zones for bacterioplankton community composition, *Environ. Microbiol. Rep.*, 8, 132–138, <https://doi.org/10.1111/1758-2229.12362>, 2016.
- Barnes, D. K. A.: Natural and plastic flotsam stranding in the Indian Ocean, in: *The Effects of Human Transport on Ecosystems: Cars and Planes, Boats and Trains*, edited by: Davenport, J. and Davenport, J. L., Royal Irish Academy, Dublin, 193–205, 2004.
- Barnes, D. K. A., Galgani, F., Thompson, R. C., and Barlaz, M.: Accumulation and fragmentation of plastic debris in global environments, *Philos. T. Roy. Soc. B.*, 364, 1985–1998, <https://doi.org/10.1098/rstb.2008.0205>, 2009.
- Beal, L. M., de Ruijter, W. P. M., Biastoch, A., and Zahn R.: On the role of the Agulhas system in ocean circulation and climate, *Nature*, 472, 429–436, <https://doi.org/10.1038/nature09983>, 2011.
- Beaumont, N. J., Aanesen, M., Austen, M. C., Börger, T., Clark, J. R., Cole, M., Hooper, T., Lindeque, P. K., Pascoe, C., and Wyles, K. J.: Global ecological, social and economic impacts of marine plastic, *Mar. Pollut. Bull.*, 142, 189–195, <https://doi.org/10.1016/j.marpolbul.2019.03.022>, 2019.
- Belkin, I. M. and Cornillon, P. C.: Fronts in the world ocean's large marine ecosystems, in: *ICES CM 2007/D:21, International Council for the Exploration of the Sea*, Copenhagen, Denmark, 1–33, 2007.
- Belkin, I. M., Cornillon, P. C., and Sherman, K.: Fronts in large marine ecosystems, *Prog. Oceanogr.*, 81, 223–236, <https://doi.org/10.1016/j.pocean.2009.04.015>, 2009.
- Bergmann, M., Peeken, I., Beyer, B., Krumpfen, T., Primpke, S., Tekman, M. B., and Gerdt, G.: Vast quantities of microplastics in Arctic Sea ice – a prime temporary sink for plastic litter and a medium of transport, in: *Fate and Impact of Microplastics in Marine Ecosystems*, edited by: Baztan, J., Jorgensen, B., Pahl, S., Thompson, R. C., and Vanderlinden, J.-P., Elsevier, Amsterdam, 75–76, 2017.
- Bond, A. L., Montevocchi, W. A., Guse, N., Regular, P. M., Garthe, S., and Rail, J.-F.: Prevalence and composition of fishing gear debris in the nests of northern gannets (*Morus bassanus*) are related to fishing effort, *Mar. Pollut. Bull.*, 64, 907–911, <https://doi.org/10.1016/j.marpolbul.2012.03.011>, 2012.
- Bouwman, H., Evans, S. W., Cole, N., Choong Kwet Yive, N. S., and Kylin, H.: The flip-or-flop boutique: marine debris on the shores of St Brandon's rock, an isolated tropical atoll in the Indian Ocean, *Mar. Environ. Res.*, 114, 58–64, <https://doi.org/10.1016/j.marenvres.2015.12.013>, 2016.
- Bower, A. S.: A simple kinematic mechanism for mixing fluid parcels across a meandering jet, *J. Phys. Oceanogr.*, 21, 173–180, [https://doi.org/10.1175/1520-0485\(1991\)021<0173:ASKMFM>2.0.CO;2](https://doi.org/10.1175/1520-0485(1991)021<0173:ASKMFM>2.0.CO;2), 1991.
- Bowman, M. J. and Esaias, W. E.: Coastal jets, fronts, and phytoplankton patchiness, *Elsevier Oceanogr. Ser.*, 19, 255–268, [https://doi.org/10.1016/S0422-9894\(08\)70846-8](https://doi.org/10.1016/S0422-9894(08)70846-8), 1977.
- Brambilla, E. and Talley, L. D.: Surface drifter exchange between the North Atlantic subtropical and subpolar gyres, *J. Geophys. Res.-Oceans*, 111, C07026, <https://doi.org/10.1029/2005JC003146>, 2006.

- Breivik, O. and Allen, A. A.: An operational search and rescue model for the Norwegian Sea and the North Sea, *J. Mar. Syst.*, 69, 99–113, <https://doi.org/10.1016/j.jmarsys.2007.02.010>, 2008.
- Brink, K. H.: Upwelling fronts: implications and unknowns, *S. Afr. J. Mar. Sci.*, 5, 3–9, <https://doi.org/10.2989/025776187784522315>, 1987.
- Brunner, K., Kukulka, T., Proskurowski, G., and Law, K. L.: Passive buoyant tracers in the ocean surface boundary layer: 2. Observations and simulations of microplastic marine debris, *J. Geophys. Res.-Oceans*, 120, 7559–7573, <https://doi.org/10.1002/2015JC010840>, 2015.
- Canales, N., Atteridge, A., and Stuesson, A.: Climate finance for the Indian Ocean and African small island developing states, Working paper 2017-11, Stockholm Environment Institute, Stockholm, 2017.
- Cartraud, A. E., Le Corre, M., Turquet, J., and Tourmetz, J.: Plastic ingestion in seabirds of the western Indian Ocean, *Mar. Pollut. Bull.*, 140, 308–314, <https://doi.org/10.1016/j.marpolbul.2019.01.065>, 2019.
- Castillo, A. B., Al-Maslamani, I., and Obbard, J. P.: Prevalence of microplastics in the marine waters of Qatar, *Mar. Pollut. Bull.*, 111, 260–267, <https://doi.org/10.1016/j.marpolbul.2016.06.108>, 2016.
- Chen, M., Pattiaratchi, C. B., Ghadouani, A., and Hanson, C.: Influence of storm events on chlorophyll distribution along the oligotrophic continental shelf off south-western Australia, *Front. Mar. Sci.*, 7, 287, <https://doi.org/10.3389/fmars.2020.00287>, 2020.
- Cherel, Y., Xavier, J. C., de Grissac, S., Trouvé, C., and Weimerskirch, H.: Feeding ecology, isotopic niche, and ingestion of fishery-related items of the wandering albatross *Diomedea exulans* at Kerguelen and Crozet Islands, *Mar. Ecol. Prog. Ser.*, 565, 197–215, <https://doi.org/10.3354/meps11994>, 2017.
- Cliff, G., Dudley, S. F., Ryan, P. G., and Singleton, N.: Large sharks and plastic debris in KwaZulu-Natal, South Africa, *Mar. Freshwater Res.*, 53, 575–581, <https://doi.org/10.1071/MF01146>, 2002.
- Cole, K. L., Macdonald, D. G., Kakoulaki, G., and Hetland, R. D.: River plume source-front connectivity, *Ocean Model.*, 150, 101571, <https://doi.org/10.1016/j.ocemod.2020.101571>, 2020.
- Courteney-Jones, W., Quinn, B., Ewins, C., Gary, S. F., and Narayanaswamy, B. E.: Microplastic accumulation in deep-sea sediments from the Rockall Trough, *Mar. Pollut. Bull.*, 154, 111092, <https://doi.org/10.1016/j.marpolbul.2020.111092>, 2020.
- Cózar, A., Echevarria, F., Gonzalez-Gordillo, J. I., Irigoien, X., Ubeda, B., Hernandez-Leon, S., Palma, A. T., Navarro, S., Garcia de Lomas, J., Ruiz, A., Fernandez de Puellas, M. L., and Duarte, C. M.: Plastic debris in the open ocean, *P. Natl. Acad. Sci. USA*, 111, 10239–10244, <https://doi.org/10.1073/pnas.1314705111>, 2014.
- Crutchett, T., Paterson, H., Ford, B. M., and Speldewinde, P.: Plastic ingestion in sardines (*Sardinops sagax*) from Frenchman Bay, Western Australia, highlights a problem in a ubiquitous fish, *Front. Mar. Sci.*, 7, 526, <https://doi.org/10.3389/fmars.2020.00526>, 2020.
- Dace, E., Pakere, I., and Blumberga, D.: Evaluation of economic aspects of the deposit-refund system for packaging in Latvia, *Manage. Environ. Qual.*, 24, 311–329, <https://doi.org/10.1108/14777831311322631>, 2013.
- D’Asaro E. A., Shcherbina, A. Y., Klymak, J. M., Molemaker, J., Novelli, G., Guigand, C. M., Haza, A. C., Haus, B. K., Ryan, E. H., Jacobs, G. A., Huntley, H. S., Laxague, N. J. M., Chen, S. Judt, F., McWilliams, J. C., Barkan, R., Kirwan, A. D., Poje, A. C., and Özgökmen, T. M.: Ocean convergence and the dispersion of flotsam, *P. Natl. Acad. Sci. USA*, 115, 1162–1167, <https://doi.org/10.1073/pnas.1718453115>, 2018.
- Davison, P. and Asch, R. G.: Plastic ingestion by mesopelagic fishes in the North Pacific subtropical gyre, *Mar. Ecol. Prog. Ser.*, 432, 173–180, <https://doi.org/10.3354/meps09142>, 2011.
- de Vos, A., Pattiaratchi, C. B., and Wijeratne, E. M. S.: Surface circulation and upwelling patterns around Sri Lanka, *Biogeosciences*, 11, 5909–5930, <https://doi.org/10.5194/bg-11-5909-2014>, 2014.
- Dobler, D., Huck, T., Maes, C., Grima, N., Blanke, B., Martinez, E., and Arduin, F.: Large impact of Stokes drift on the fate of surface floating debris in the South Indian Basin, *Mar. Pollut. Bull.*, 148, 202–209, <https://doi.org/10.1016/j.marpolbul.2019.07.057>, 2019.
- Duhec, A. V., Jeanne, R. F., Maximenko, N., and Hafner, J.: Composition and potential origin of marine debris stranded in the western Indian Ocean on remote Alphonse Island, Seychelles, *Mar. Pollut. Bull.*, 96, 76–86, <https://doi.org/10.1016/j.marpolbul.2015.05.042>, 2015.
- Egger, M., Sulu-Gambari, F., and Lebreton, L.: First evidence of plastic fallout from the north Pacific garbage patch, *Sci. Rep.*, 10, 7495, <https://doi.org/10.1038/s41598-020-64465-8>, 2020.
- Eliot, M. and Pattiaratchi, C.: Remote forcing of water levels by tropical cyclones in southwest Australia, *Cont. Shelf Res.*, 30, 1549–1561, <https://doi.org/10.1016/j.csr.2010.06.002>, 2010.
- Engler, R. E.: The complex interaction between marine debris and toxic chemicals in the ocean, *Environ. Sci. Technol.*, 46, 12302–12315, <https://doi.org/10.1021/es3027105>, 2012.
- Eriksen, M., Lebreton, L. C. M., Carson, H. S., Thiel, M., Moore, C. J., Borerro, J. C., Galgani, F., Ryan, P. G., and Reisser, J.: Plastic pollution in the world’s oceans: more than 5 trillion plastic pieces weighing over 250,000 tons afloat at sea, *PLoS ONE*, 9, e111913, <https://doi.org/10.1371/journal.pone.0111913>, 2014.
- Findlater, J.: A major low-level air current near the Indian Ocean during the northern summer, *Q. J. Roy. Meteorol. Soc.*, 95, 362–380, <https://doi.org/10.1002/qj.49709540409>, 1969.
- Freinkel, S.: A brief history of plastic’s conquest of the world, *Sci. Am.*, available at: <https://www.scientificamerican.com/article/a-brief-history-of-plastic-world-conquest/> (last access: 21 September 2021), 2011.
- Frias, J. P. G. L. and Nash, R.: Microplastics: finding a consensus on the definition, *Mar. Pollut. Bull.*, 138, 145–147, <https://doi.org/10.1016/j.marpolbul.2018.11.022>, 2019.
- Froyland, G., Stuart, R. M., and van Sebille, E.: How well-connected is the surface of the global ocean?, *Chaos*, 24, 033126, <https://doi.org/10.1063/1.4892530>, 2014.
- Geyer, R., Jambeck, J. R., and Law, K. L.: Production, use, and fate of all plastics ever made, *Sci. Adv.*, 3, e1700782, <https://doi.org/10.1126/sciadv.1700782>, 2017.
- Gigault, J., ter Halle, A., Baudrimont, M., Pascal, P. Y., Gauffre, F., Phi, T. L., El Hadri, H., Grassl, B., and Reynaud, S.: Current opinion: what is a nanoplastic? *Environ. Pollut.*, 235, 1030–1034, <https://doi.org/10.1016/j.envpol.2018.01.024>, 2018.

- Gordon, A. L.: The browniest retroreflection, *Nature*, 421, 904–905, <https://doi.org/10.1038/421904a>, 2003.
- Gregory, M. R.: Environmental implications of plastic debris in marine settings – entanglement, ingestion, smothering, hangers-on, hitch-hiking and alien invasions, *Philos. T. Roy. Soc. B*, 364, 2013–2025, <https://doi.org/10.1098/rstb.2008.0265>, 2009.
- Guillotreau, P., Campling, L., and Robinson, J.: Vulnerability of small island fishery economies to climate and institutional changes, *Curr. Opin. Environ. Sustain.*, 4, 287–291, <https://doi.org/10.1016/j.cosust.2012.06.003>, 2012.
- Gunn, R., Hardesty, B. D., and Butler, J.: Tackling ‘ghost nets’: local solutions to a global issue in northern Australia, *Ecol. Manage. Restor.*, 11, 88–98, <https://doi.org/10.1111/j.1442-8903.2010.00525.x>, 2010.
- Guptha, M. V. S., Curry, W. B., Ittekkot, V., and Muralinath, A. S.: Seasonal variation in the flux of planktic foraminifera: sediment trap results from the Bay of Bengal, northern Indian Ocean, *J. Foramin. Res.*, 27, 5–19, <https://doi.org/10.2113/gsjfr.27.1.5>, 1997.
- Hajbane, S., Calmanovici, B., Reisser, J., Jolly, A., Summers, V., Ferrari, F., Ghadouani, A., and Pattiaratchi, C. B.: Coastal garbage patches: evidence for significant accumulations of buoyant plastics along shallow coastal fronts at Ashmore Reef (Pulau Pasir) Marine Park, *Front. Mar. Sci.*, 8, 613399, <https://doi.org/10.3389/fmars.2021.613399>, 2021.
- Hartmann, N. B., Thorsten, H., Thompson, R. C., Hassellöv, M., Verschoor, A., Daugaard, A. E., Rist, S., Karlsson, T., Brennholt, N., Cole, M., Herrling, M. P., Hess, M. C., Ivleva, N. P., Lusher, A. L., and Wagner, M.: Are we speaking the same language? Recommendations for a definition and categorization framework for plastic debris, *Environ. Sci. Technol.*, 53, 1039–1047, <https://doi.org/10.1021/acs.est.8b05297>, 2019.
- Hoarau, L., Ainley, L., Jean, C., and Ciccione, S.: Ingestion and defecation of marine debris by loggerhead sea turtles, *Caretta caretta*, from by-catches in the South-West Indian Ocean, *Mar. Pollut. Bull.*, 84, 90–96, <https://doi.org/10.1016/j.marpolbul.2014.05.031>, 2014.
- Hoorweg, D. and Bhada-Tata, P.: What a waste: a global review of solid waste management, Urban development series knowledge papers no. 15, World Bank, Washington, DC, available at: <https://openknowledge.worldbank.org/handle/10986/17388> (last access: 21 September 2021), 2012.
- Huyer, A.: Shelf circulation, in: vol. 9A, *The Sea: Ocean Engineering Science*, edited by: Mehaute, B. L. and Hanes, D., Wiley Interscience, New York, 423–466, 1990.
- Imhof, H. K., Sigl, R., Brauer, E., Feyl, S., Giesemann, P., Klink, S., Leupolz, K., Löder, M. G. J., Löscher, L. A., Missun, J., Muszynski, S., Ramsperger, A. F. R. M., Schrank, I., Speck, S., Steibl, S., Trotter, B., Winter, I., and Laforsch, C.: Spatial and temporal variation of macro-, meso- and microplastic abundance on a remote coral island of the Maldives, Indian Ocean, *Mar. Pollut. Bull.*, 116, 340–347, <https://doi.org/10.1016/j.marpolbul.2017.01.010>, 2017.
- Ismail, M. R., Lewaru, M. W., and Prihadi, D. J.: Microplastics ingestion by fish in the Biawak Island, *World Sci. News*, 106, 230–237, 2018.
- Isobe, A., Kubo, K., Tamura, Y., Kako, S., Nakashima, E., and Fujii, N.: Selective transport of microplastics and mesoplastics by drifting in coastal waters, *Mar. Pollut. Bull.*, 89, 324–330, <https://doi.org/10.1016/j.marpolbul.2014.09.041>, 2014.
- Isobe, A., Buenaventura, N. T., Chastain, S., Chavanich, S., Cózar, A., DeLorenzo, M., Hagmann, P., Hinata, H., Kozlovskii, N., Lusher, A. L., Martí, E., Michida, Y., Mu, J., Ohno, M., Potter, G., Ross, P. S., Sagawa, N., Shim, W. J., Song, Y. K., Takada, H., Tokai, T., Torii, T., Uchida, K., Vassilenko, K., Viyakarn, V., and Zhang, W.: An interlaboratory comparison exercise for the determination of microplastics in standard sample bottles, *Mar. Pollut. Bull.*, 146, 831–837, <https://doi.org/10.1016/j.marpolbul.2019.07.033>, 2019.
- Ittekkot, V., Nair, R. R., Honjo, S., Ramaswamy, V., Bartsch, M., Manginini, S., and Desai B. N.: Enhanced particle fluxes in Bay of Bengal induced by injection of fresh water, *Nature*, 351, 385–387, <https://doi.org/10.1038/351385a0>, 1996.
- Jambeck, J. R., Geyer, R., Wilcox, C., Siegler, T. R., Perryman, M., Andrady, A., Narayan, R., and Law, K. L.: Plastic waste inputs from land into the ocean, *Science*, 347, 768–771, <https://doi.org/10.1126/science.1260352>, 2015.
- Jamieson, A. J., Brooks, L. S. R., Reid, W. D. K., Piertney, S. B., Narayanaswamy, B. E., and Linley, T. D.: Microplastics and synthetic particles ingested by deep-sea amphipods in six of the deepest marine ecosystems on Earth, *Roy. Soc. Open Sci.*, 6, 180667, <https://doi.org/10.1098/rsos.180667>, 2019.
- Jayasiri, H. B., Purushothaman, C. S., and Vennila, A.: Plastic litter accumulation on high-water strandline of urban beaches in Mumbai, India, *Environ. Monit. Assess.*, 185, 7709–7719, <https://doi.org/10.1007/s10661-013-3129-z>, 2013.
- Joseph, P. V. and Sijikumar, S.: Intraseasonal variability of the low-level jet stream of the Asian summer monsoon, *J. Climate*, 17, 1449–1458, [https://doi.org/10.1175/1520-0442\(2004\)017<1449:IVOTLJ>2.0.CO;2](https://doi.org/10.1175/1520-0442(2004)017<1449:IVOTLJ>2.0.CO;2), 2004.
- Karati, K. K., Vineetha, G., Raveendran, T. V., Muraleedharan, K. R., Habeebrehman, H., Philson, K. P., and Achuthankutty, C. T.: River plume fronts and their implications for the biological production of the Bay of Bengal, Indian Ocean, *Mar. Ecol. Prog. Ser.*, 597, 79–98, <https://doi.org/10.3354/meps12607>, 2018.
- Karthik, R., Robin, R. S., Purvaja, R., Ganguly, D., Anandavelu, I., Raghuraman, R., Hariharan, G., Ramakrishna, A., and Ramesh, R.: Microplastics along the beaches of south-east coast of India, *Sci. Total Environ.*, 645, 1388–1399, <https://doi.org/10.1016/j.scitotenv.2018.07.242>, 2018.
- Kaza, S., Yao, L. C., Bhada-Tata, P., and Van Woerden, F.: What a waste 2.0: a global snapshot of solid waste management to 2050, Urban development, World Bank, Washington, DC, IGO, available at: <https://openknowledge.worldbank.org/handle/10986/30317> (last access: 28 December 2021), 2018.
- Khalik, W. M. A. W. M., Ibrahim, Y. S., Tuan Anuar, S., Govindasamy, S., and Baharuddin, N. F.: Microplastics analysis in Malaysian marine waters: a field study of Kuala Nerus and Kuantan, *Mar. Pollut. Bull.*, 135, 451–457, <https://doi.org/10.1016/j.marpolbul.2018.07.052>, 2018.
- Koelmans, A. A., Kooi, M., and van Sebille, E.: All is not lost: deriving a top-down mass budget of plastic at sea, *Environ. Res. Lett.*, 12, 1–9, <https://doi.org/10.1088/1748-9326/aa9500>, 2017.
- Kooi, M., Reisser, J., Slat, B., Ferrari, F. F., Schmid, M. S., Cunsolo, S., Brambini, R., Noble, K., Sirks, L., Linders, T. E. W., Schoeneich-Argent, R. I., and Koelmans, A. A.: The effect of particle properties on the depth profile of buoyant plastics in the

- ocean, *Sci. Rep.*, 6, 33882, <https://doi.org/10.1038/srep33882>, 2016.
- Kooi, M., Nes, E. H., Scheffer, M., and Koelmans, A. A.: Ups and downs in the ocean: effects of biofouling on vertical transport of microplastics, *Environ. Sci. Technol.*, 51, 7963–7971, <https://doi.org/10.1021/acs.est.6b04702>, 2017.
- Koongolla, J. B., Andrady, A. L., Terney Pradeep Kumara, P. B., and Gangabadage, C. S.: Evidence of microplastics pollution in coastal beaches and waters in southern Sri Lanka, *Mar. Pollut. Bull.*, 137, 277–284, <https://doi.org/10.1016/j.marpolbul.2018.10.031>, 2018.
- Kosore, C., Ojwang, L., Maghanga, J., Kamau, J., Kimeli, A., Omukoto, J., Ngisiag'e, N., Mwaluma, J., Ong'ada, H., Magori, C., and Ndirui, E.: Occurrence and ingestion of microplastics by zooplankton in Kenya's marine environment: first documented evidence, *Afr. J. Mar. Sci.*, 40, 225–234, <https://doi.org/10.2989/1814232X.2018.1492969>, 2018.
- Kukulka, T. and Brunner, K.: Passive buoyant tracers in the ocean surface boundary layer: 1. Influence of equilibrium wind-waves on vertical distributions, *J. Geophys. Res.-Oceans*, 120, 3837–3858, <https://doi.org/10.1002/2014JC010487>, 2015.
- Kukulka, T., Proskurowski, G., Moret-Ferguson, S., Meyer, D. W., and Law, K. L.: The effect of wind mixing on the vertical distribution of buoyant plastic debris, *Geophys. Res. Lett.*, 39, L07601, <https://doi.org/10.1029/2012GL051116>, 2012.
- Lachmann, F., Almroth, B. C., Baumann, H., Broström, G., Corvellec, H., Gipperth, L., Hassellöv, M., Karlsson, T., and Nilsson, P.: Marine plastic litter on small island developing states (SIDS): impacts and measures, Report no. 2017:4, Swedish Institute for the Marine Environment, University of Gothenburg, Gothenburg, 2017.
- Laist, D. W. and Liffman, M.: Impacts of marine debris: research and management needs, in: Proceedings of the International Marine Debris Conference on Derelict Fishing Gear and the Ocean Environment, edited by: McIntosh, N., Simonds, K., Donohue, M., Brammer, C., Manson, S., and Carbajal, S., 6–11 August 2000, Honolulu, 344–357, 2000.
- Lambert, E., Le Bars, D., and de Ruijter, W. P. M.: The connection of the Indonesian throughflow, south Indian Ocean counter-current and the Leeuwin current, *Ocean Sci.*, 12, 771–780, <https://doi.org/10.5194/os-12-771-2016>, 2016.
- Lavee, D.: A cost-benefit analysis of a deposit-refund program for beverage containers in Israel, *Waste Manage.*, 30, 338–345, <https://doi.org/10.1016/j.wasman.2009.09.026>, 2010.
- Lavers, J. L. and Bond, A. L.: Exceptional and rapid accumulation of anthropogenic debris on one of the world's most remote and pristine islands, *P. Natl. Acad. Sci. USA*, 114, 6052–6055, <https://doi.org/10.1073/pnas.1619818114>, 2017.
- Lavers, J. L., Dicks, L., Dicks, M. R., and Finger, A.: Significant plastic accumulation on the Cocos (Keeling) Islands, Australia, *Sci. Rep.*, 9, 7102, <https://doi.org/10.1038/s41598-019-43375-4>, 2019.
- Law, K. L.: Plastics in the marine environment, *Annu. Rev. Mar. Sci.*, 9, 205–229, <https://doi.org/10.1146/annurev-marine-010816-060409>, 2017.
- Laxague, N. J. M., Ozgokmen, T. M., Haus, B. K., Novelli, G., Shcherbina, A., Sutherland, P., Guigand, C. M., Lund, B., Mancha, S., Alday, M., and Molemaker, J.: Observations of near-surface current shear help describe oceanic oil and plastic transport, *Geophys. Res. Lett.*, 45, 245–249, <https://doi.org/10.1002/2017GL075891>, 2017.
- Lebreton, L., Egger, M., and Slat, B.: A global mass budget for positively buoyant microplastic debris in the ocean, *Sci. Rep.*, 9, 12922, <https://doi.org/10.1038/s41598-019-49413-5>, 2019.
- Lebreton, L. C., Greer, S. D., and Borrero, J. C.: Numerical modelling of floating debris in the world's oceans, *Mar. Pollut. Bull.*, 64, 653–661, <https://doi.org/10.1016/j.marpolbul.2011.10.027>, 2012.
- Lebreton, L. C. M., van der Zwet, J., Damsteeg, J.-W., Slat, B., Andrady, A., and Reisser, J.: River plastic emissions to the world's oceans, *Nat. Commun.*, 8, 15611, <https://doi.org/10.1038/ncomms15611>, 2017.
- Lett, C., Verley, P., Mullon, C., Parada, C., Brochier, T., Penven, P., and Balnke, P.: A Lagrangian tool for modelling ichthyoplankton dynamics, *Environ. Model. Softw.*, 23, 1210–1214, <https://doi.org/10.1016/j.envsoft.2008.02.005>, 2008.
- Lobelle, D. and Cunliffe, M.: Early microbial biofilm formation of marine plastic debris, *Mar. Pollut. Bull.*, 62, 197–200, <https://doi.org/10.1016/j.marpolbul.2010.10.013>, 2011.
- Long, M., Moriceau, B., Gallinari, M., Lambert, C., Huvet, A., Raffray, J., and Soudant, P.: Interactions between microplastics and phytoplankton aggregates: impact on their respective fates, *Mar. Chem.*, 175, 39–46, <https://doi.org/10.1016/j.marchem.2015.04.003>, 2015.
- Luketina, D. A. and Imberger, J.: Turbulence and entrainment in a buoyant surface plume, *J. Geophys. Res.-Oceans*, 94, 12619–12636, <https://doi.org/10.1029/JC094iC09p12619>, 1989.
- Lutjeharms, J. R. E. (Ed.): The Agulhas Current retroreflection, in: The Agulhas Current, Springer, Switzerland, 151–207, 2006.
- Macfadyen, G., Huntington, T., and Cappell, R.: Abandoned, lost or otherwise discarded fishing gear, UNEP Regional Seas Reports and Studies no. 185, FAO Fisheries and Aquaculture Technical Paper no. 523, FAO, Rome, 2009.
- Madzena, A. and Lasiak, T.: Spatial and temporal variations in beach litter on the Transkei coast of South Africa, *Mar. Pollut. Bull.*, 34, 900–907, [https://doi.org/10.1016/S0025-326X\(97\)00052-0](https://doi.org/10.1016/S0025-326X(97)00052-0), 1997.
- Maes, C., Grima, N., Blanke, B., Martinez, E., Paviet-Salomon, T., and Huck, T.: A surface “superconvergence” pathway connecting the south Indian Ocean to the subtropical south Pacific gyre, *Geophys. Res. Lett.*, 45, 1915–1922, <https://doi.org/10.1002/2017GL076366>, 2018.
- Maximenko, N., Hafner, J., and Niiler, P.: Pathways of marine debris derived from trajectories of Lagrangian drifters, *Mar. Pollut. Bull.*, 65, 51–62, <https://doi.org/10.1016/j.marpolbul.2011.04.016>, 2012.
- McGregor, S. and Strydom, N. A.: Feeding ecology and microplastic ingestion in *Chelon richardsonii* (Mugilidae) associated with surf diatom *Anaulus australis* accumulations in a warm temperate South African surf zone, *Mar. Pollut. Bull.*, 158, 111430, <https://doi.org/10.1016/j.marpolbul.2020.111430>, 2020.
- Meijer, L. J. J., van Emmerink, T., van der Ent, R., and Lebreton, L.: More than 1000 rivers account for 80 % of global riverine plastic emissions into the ocean, *Sci. Adv.*, 7, eaaz5803, <https://doi.org/10.1126/sciadv.aaz5803>, 2021.
- Meikle, J. L.: American Plastic: A Cultural History, first paperback edition, Rutgers University Press, New Jersey, USA, 1997.

- Menezes, V. V., Phillips, H. E., Schiller, A., Bindoff, N. L., Domingues, C. M., and Vianna, M. L.: South Indian countercurrent and associated fronts, *J. Geophys. Res.-Oceans*, 119, 6763–6791, <https://doi.org/10.1002/2014JC010076>, 2014.
- Middleton, J. F. and Cirano, M.: A northern boundary current along Australia's southern shelves: the Flinders current, *J. Geophys. Res.*, 107, 1–11, <https://doi.org/10.1029/2000JC000701>, 2002.
- Miyao, Y. and Isobe, A.: A combined balloon photography and buoy-tracking experiment for mapping surface currents in coastal waters, *J. Atmos. Ocean. Tech.*, 33, 1237–1250, <https://doi.org/10.1175/JTECH-D-15-0113.1>, 2016.
- Monteiro, R. C., do Sul, J. A. I., and Costa, M. F.: Plastic pollution in islands of the Atlantic Ocean, *Environ. Pollut.*, 238, 103–110, <https://doi.org/10.1016/j.envpol.2018.01.096>, 2018.
- Morris, R. J.: Plastic debris in the surface waters of the South Atlantic, *Mar. Pollut. Bull.*, 11, 164–166, [https://doi.org/10.1016/0025-326X\(80\)90144-7](https://doi.org/10.1016/0025-326X(80)90144-7), 1980.
- Nahas, E. L., Pattiaratchi, C. B., and Ivey, G. N.: Processes controlling the position of frontal systems in Shark Bay, Western Australia, *Estuar. Coast Shelf Sci.*, 65, 463–474, <https://doi.org/10.1016/j.ecss.2005.06.017>, 2005.
- Naidu, S. A.: Preliminary study and first evidence of presence of microplastics and colorants in green mussel, *Perna viridis* (Linnaeus, 1758), from southeast coast of India, *Mar. Pollut. Bull.*, 140, 416–422, <https://doi.org/10.1016/j.marpolbul.2019.01.024>, 2019.
- Naidu, S. A., Ranga Rao, V., and Ramu, K.: Microplastics in the benthic invertebrates from the coastal waters of Kochi, south-eastern Arabian Sea, *Environ. Geochem. Health*, 40, 1377–1383, <https://doi.org/10.1007/s10653-017-0062-z>, 2018.
- Naidu, S. A., Mawii, L., Ranga Rao, V., Anitha, G., Mishra, P., Narayanaswamy, B. E., Anil Kumar, V., Ramana Murthy, M. V., and Gupta, G. V. M.: Characterization of plastic debris from surface waters of the eastern Arabian Sea–Indian Ocean, *Mar. Pollut. Bull.*, 169, 112468, <https://doi.org/10.1016/j.marpolbul.2021.112468>, 2021.
- Naji, A., Nuri, M., and Vethaak, A. D.: Microplastics contamination in molluscs from the northern part of the Persian Gulf, *Environ. Pollut.*, 235, 113–120, <https://doi.org/10.1016/j.envpol.2017.12.046>, 2018.
- Nel, H. A. and Froneman, P. W.: A quantitative analysis of microplastic pollution along the south-eastern coastline of South Africa, *Mar. Pollut. Bull.*, 101, 274–279, <https://doi.org/10.1016/j.marpolbul.2015.09.043>, 2015.
- Nor, N. H. M. and Obbard, J. P.: Microplastics in Singapore's coastal mangrove ecosystems, *Mar. Pollut. Bull.*, 79, 278–283, <https://doi.org/10.1016/j.marpolbul.2013.11.025>, 2014.
- Obbard, R. W., Sadri, S., Wong, Y. Q., Khitun, A. A., Baker, I., and Thompson, R. C.: Global warming releases microplastic legacy frozen in Arctic Sea ice, *Earth's Future*, 2, 315–320, <https://doi.org/10.1002/2014EF000240>, 2014.
- O'Donnell, J., Marmorino, G. O., and Trump, C. L.: Convergence and downwelling at a river plume front, *J. Phys. Oceanogr.*, 28, 1481–1495, [https://doi.org/10.1175/1520-0485\(1998\)028<1481:CADAAR>2.0.CO;2](https://doi.org/10.1175/1520-0485(1998)028<1481:CADAAR>2.0.CO;2), 1998.
- Owen, R. W.: Fronts and eddies in the sea: mechanisms, interactions and biological effects, in: *Analysis of Marine Ecosystems*, edited by: Longhurst, A. R., Academic Press, London, 197–233, 1981.
- Palastanga, V., van Leeuwen, P. J., Schouten, M. W., and de Ruijter, W. P. M.: Flow structure variability in the subtropical Indian Ocean: instability of the south Indian Ocean countercurrent, *J. Geophys. Res.*, 112, C01001, <https://doi.org/10.1029/2005JC003395>, 2007.
- Patti, T. B., Fobert, E. K., Reeves, S. E., and da Silva, K. B.: Spatial distribution of microplastics around an inhabited coral island in the Maldives, Indian Ocean, *Sci. Total Environ.*, 748, 141263, <https://doi.org/10.1016/j.scitotenv.2020.141263>, 2020.
- Pattiaratchi, C. and Woo, M.: The mean state of the Leeuwin current system between North West Cape and Cape Leeuwin, *J. Roy. Soc. West. Aust.*, 92, 221–241, 2009.
- Pattiaratchi, C. B., James, A. E., and Collins, M. B.: Island wakes and headland eddies: a comparison between remotely sensed data and laboratory experiments, *J. Geophys. Res.-Oceans*, 92, 783–794, <https://doi.org/10.1029/JC092iC01p0078319>, 1987.
- Pauly, D. and Zeller, D.: Catch reconstructions reveal that global marine fisheries catches are higher than reported and declining, *Nat. Commun.*, 7, 10244, <https://doi.org/10.1038/ncomms10244>, 2016.
- Peeken, I., Primpke, S., Beyer, B., Gütermann, J., Katlein, C., Krumpfen, T., Bergmann, M., Hehemann, L., and Gerdt, G.: Arctic sea ice is an important temporal sink and means of transport for microplastic, *Nat. Commun.*, 9, 1505, <https://doi.org/10.1038/s41467-018-03825-5>, 2018.
- Pham, C. K., Pereira, J. M., Frias, J. P., Ríos, N., Carriço, R., Juliano, M., and Rodríguez, Y.: Beaches of the Azores archipelago as transitory repositories for small plastic fragments floating in the North-East Atlantic, *Environ. Pollut.*, 263, 114494, <https://doi.org/10.1016/j.envpol.2020.114494>, 2020.
- PlasticsEurope: Plastics – the facts 2019: an analysis of European plastics production, demand and waste data, Technical report, PlasticsEurope, available at: <https://plasticseurope.org/knowledge-hub/plastics-the-facts-2019/> (last access: 28 December 2021), 2019.
- Qasim, S. Z.: Contribution of zooplankton in food chains of some warm water environment, in: *Proceedings of the Symposium on Warm Water Zooplankton*, National Institute of Oceanography, Goa, 700–708, 1977.
- Qui, Y. and Yu, W.: Behaviour of the Wrytki jet observed with surface drifting buoys and satellite altimeter, *Geophys. Res. Lett.*, 36, 1–5, <https://doi.org/10.1029/2009GL039120>, 2009.
- Raja, K., Aanand, P., Padmavathy, S., and Sampathkumar, J. S.: Present and future market trends of Indian ornamental fish sector, *Int. J. Fish Aquat. Stud.*, 7, 6–15, 2019.
- Raynaud, J.: Valuing plastic: the business case for measuring, managing and disclosing plastic use in the consumer goods industry, UNEP, available at: <https://wedocs.unep.org/handle/20.500.11822/25302> (last access: 28 December 2021), 2014.
- Readman, J. W., DeLuna, F., Ebinghaus, R., Guzman, A. N., Price, A. R. G., Readman, E. E., Sheppard, A. L. S., Sleight, V. A., Sturm, R., Thompson, R. C., Tonkin, A., Wolschke, H., Wright, R. J., and Sheppard, C. R. C.: Contaminants, pollution and potential anthropogenic impacts in Chagos/BIOT, in: *Coral Reefs of the United Kingdom Overseas Territories*, Coral Reefs of the World, vol. 4, edited by: Sheppard, C. R. C., Springer, Dordrecht, https://doi.org/10.1007/978-94-007-5965-7_21, 2013.
- Reddy, M. S., Basha, S., Adimurthy, S., and Ramachandriah, G.: Description of the small plastics fragments in

- marine sediments along the Alang-Sosiya ship-breaking yard, India, *Estuar. Coast. Shelf Sci.*, 68, 656–660, <https://doi.org/10.1016/j.ecss.2006.03.018>, 2006.
- Reisser, J., Shaw, J., Wilcox, C., Hardesty, B. D., Proietti, M., Thums, M., and Pattiaratchi, C.: Marine plastic pollution in waters around Australia: characteristics, concentrations, and pathways, *PLoS ONE*, 8, e80466, <https://doi.org/10.1371/journal.pone.0080466>, 2013.
- Reisser, J., Slat, B., Noble, K., du Plessis, K., Epp, M., Proietti, M., de Sonnevile, J., Becker, T., and Pattiaratchi, C.: The vertical distribution of buoyant plastics at sea: an observational study in the North Atlantic gyre, *Biogeosciences*, 12, 1249–1256, <https://doi.org/10.5194/bg-12-1249-2015>, 2015.
- Richardson, K., Asmutis-Silvia, R., Drinkwin, J., Gilardi, K. V., Giskes, I., Jones, G., O'Brien, K., Pragnell-Raasch, H., Ludwig, L., Antonelis, K., and Barco, S.: Building evidence around ghost gear: global trends and analysis for sustainable solutions at scale, *Mar. Pollut. Bull.*, 138, 222–229, <https://doi.org/10.1016/j.marpolbul.2018.11.031>, 2019a.
- Richardson, K., Hardesty, B. D., and Wilcox, C.: Estimates of fishing gear loss rates at a global scale: a literature review and meta-analysis, *Fish Fish.*, 20, 1218–1231, <https://doi.org/10.1111/faf.12407>, 2019b.
- Richardson, K., Hardesty, B. D., Vince, J. Z., and Wilcox, C.: Global causes, drivers, and prevention measures for lost fishing gear, *Front. Mar. Sci.*, 8, 690447, <https://doi.org/10.3389/fmars.2021.690447>, 2021.
- Richardson, P. L.: Drifting in the wind: leeway error in shipdrift data, *Deep-Sea Res.*, 44, 1877–1903, [https://doi.org/10.1016/S0967-0637\(97\)00059-9](https://doi.org/10.1016/S0967-0637(97)00059-9), 1997.
- Robin, R. S., Karthik, R., Purvaja, R., Ganguly, D., Anandavelu, I., Mugilarasan, M., and Ramesh, R.: Holistic assessment of microplastics in various coastal environmental matrices, southwest coast of India, *Sci. Total Environ.*, 703, 134947, <https://doi.org/10.1016/j.scitotenv.2019.134947>, 2020.
- Rochman, C. M., Hoh, E., Hentschel, B. T., and Kaye, S.: Long-term field measurements of sorption of organic contaminants to five types of plastic pellets: implications for plastic marine debris, *Environ. Sci. Technol.*, 47, 1646–1654, <https://doi.org/10.1021/es303700s>, 2012.
- Röhrs, J., Christensen, K. H., Hole, L. R., Brostrom, G., Drivdal, M., and Sundby, S.: Observation-based evaluation of surface wave effects on currents and trajectory forecasts, *Ocean Dynam.*, 62, 1519–1533, <https://doi.org/10.1007/s10236-012-0576-y>, 2012.
- Roman, L., Schuyler, Q., Wilcox, C., and Hardesty, B. D.: Plastic pollution is killing marine megafauna, how do we prioritize policy to reduce mortality?, *Conserv. Lett.*, 14, e12781, <https://doi.org/10.1111/conl.12781>, 2021.
- Roy, R., Chitari, R., Kulkarni, V., Krishna, M. S., Sarma, V. V. S. S., and Anil, A. C.: CHEMTAX-derived phytoplankton community structure associated with temperature fronts in the northeastern Arabian Sea, *J. Mar. Syst.*, 144, 81–91, <https://doi.org/10.1016/j.jmarsys.2014.11.009>, 2015.
- Ryan, P. G.: The origin and fate of artefacts stranded on islands in the African sector of the Southern Ocean, *Environ. Conserv.*, 14, 341–346, <https://doi.org/10.1017/S0376892900016854>, 1987.
- Saliu, F., Montano, S., Garavaglia, M. G., Lasagni, M., Seveso, D., and Galli, P.: Microplastic and charred microplastic in the Faafu Atoll, Maldives, *Mar. Pollut. Bull.*, 136, 464–471, <https://doi.org/10.1016/j.marpolbul.2018.09.023>, 2018.
- Saliu, F., Montano, S., Leoni, B., Lasagni, M., and Galli, P.: Microplastics as a threat to coral reef environments: detection of phthalate esters in neuston and scleractinian corals from the Faafu Atoll, Maldives, *Mar. Pollut. Bull.*, 142, 234–241, <https://doi.org/10.1016/j.marpolbul.2019.03.043>, 2019.
- Sari, D. A. A., Suryanto, Sudarwanto, A. S., Nugraha, S., and Utomowati, R.: Reduce marine debris policy in Indonesia, *Earth Environ. Sci.*, 724, 012118, <https://doi.org/10.1088/1755-1315/724/1/012118>, 2021.
- Sarijan, S., Azman, S., Said, M. I. M., Andu, Y., and Zon, N. F.: Microplastics in sediment from Skudai and Tebrau river, Malaysia: a preliminary study, *MATEC Web Conf.*, 250, 06012, <https://doi.org/10.1051/mateconf/201825006012>, 2018.
- Sarkar, K., Aparna, S. G., Dora, S., and Shankar, D.: Seasonal variability of sea-surface temperature fronts associated with large marine ecosystems in the north Indian Ocean, *J. Earth Syst. Sci.*, 128, 1–8, <https://doi.org/10.1007/s12040-018-1045-x>, 2018.
- Sarma, V. V. S. S., Delabehra, H. B., Sudharani, P., Remya, R., Patil, J. S., and Desai, D.V.: Variations in the inorganic carbon components in the thermal fronts during winter in the northeastern Arabian Sea, *Mar. Chem.*, 169, 16–22, <https://doi.org/10.1016/j.marchem.2014.12.009>, 2015.
- Sathish, N., Jeyasanta, K. I., and Patterson, J.: Abundance, characteristics and surface degradation features of microplastics in beach sediments of five coastal areas in Tamil Nadu, India, *Mar. Pollut. Bull.*, 142, 112–118, <https://doi.org/10.1016/j.marpolbul.2019.03.037>, 2019.
- Schmidt, C., Krauth, T., and Wagner, S.: Export of plastic debris by rivers into the sea, *Environ. Sci. Technol.*, 51, 12246–12253, <https://doi.org/10.1021/acs.est.7b02368>, 2017.
- Schmidt, C., Krauth, T., and Wagner, S.: Correction to export of plastic debris by rivers into the sea, *Environ. Sci. Technol.*, 52, 927, <https://doi.org/10.1021/acs.est.7b06377>, 2018.
- Schott, F. A. and McCreary, J. P.: The monsoon circulation of the Indian Ocean, *Prog. Oceanogr.*, 51, 1–123, [https://doi.org/10.1016/S0079-6611\(01\)00083-0](https://doi.org/10.1016/S0079-6611(01)00083-0), 2001.
- Schott, F. A., Xie, S. P., and McCreary, J. P.: Indian Ocean circulation and climate variability, *Rev. Geophys.*, 47, RG1002, <https://doi.org/10.1029/2007RG000245>, 2009.
- Schumann, E. H., MacKay, C. F., and Strydom, N. A.: Nurdle drifters around South Africa as indicators of ocean structures and dispersion, *S. Afr. J. Sci.*, 115, 5372, <https://doi.org/10.17159/sajs.2019/5372>, 2019.
- Schuyler, Q., Hardesty, B. D., Lawson, T. J., Opie, K., and Wilcox, C.: Economic incentives reduce plastic inputs to the ocean, *Mar. Policy*, 96, 250–255, <https://doi.org/10.1016/j.marpol.2018.02.009>, 2018.
- Serra-Gonçalves, C., Lavers, J. L., and Bond, A. L.: Global review of beach debris monitoring and future recommendations, *Environ. Sci. Technol.*, 53, 12158–12167, <https://doi.org/10.1021/acs.est.9b01424>, 2019.
- Seth, C. K. and Shriwastav, A.: Contamination of Indian sea salts with microplastics and a potential prevention strategy, *Environ. Sci. Pollut. Res.*, 25, 30122–30131, <https://doi.org/10.1007/s11356-018-3028-5>, 2018.
- Shankar, D. and Shetye, S. R.: On the dynamics of the Lakshadweep high and low in the southeastern Arabian Sea, *J. Geophys. Res.*

- Oceans, 102, 12551–12562, <https://doi.org/10.1029/97JC00465>, 1997.
- Sharples, J. and Simpson, J. H.: Shelf Sea and Shelf Slope Fronts, in: *Encyclopaedia of Ocean Sciences*, 2nd Edn., edited by: Steele, J. H., Academic Press, Cambridge, Massachusetts, USA, 391–400, <https://doi.org/10.1016/B978-012374473-9.00623-8>, 2020.
- Sheavly, S. B. and Register, K. M.: Marine debris & plastics: environmental concerns, sources, impacts and solutions, *J. Polym. Environ.*, 15, 301–305, <https://doi.org/10.1007/s10924-007-0074-3>, 2007.
- Shenoi, S. S. C., Shankar, D., and Shetye, S. R.: Remote forcing annihilates barrier layer in southeastern Arabian Sea, *Geophys. Res. Lett.*, 31, L05307, <https://doi.org/10.1029/2003GL019270>, 2004.
- Simpson, J. H. and Hunter, J. R.: Fronts in the Irish Sea, *Nature*, 250, 404–406, <https://doi.org/10.1038/250404a0>, 1974.
- Slip, D. J. and Burton, H. R.: Accumulation of fishing debris, plastic litter, and other artefacts, on Heard and Macquarie islands in the Southern Ocean, *Environ. Conserv.*, 18, 249–254, <https://doi.org/10.1017/S0376892900022177>, 1991.
- Spalding, M., Burke, L., Wood, S. A., Ashpole, J., Hutchison, J., and zu Ermgassen, P.: Mapping the global value and distribution of coral reef tourism, *Mar. Policy*, 82, 104–113, <https://doi.org/10.1016/j.marpol.2017.05.014>, 2017.
- Sparks, C. and Immelman, S.: Microplastics in offshore fish from the Agulhas Bank, South Africa, *Mar. Pollut. Bull.*, 156, 111216, <https://doi.org/10.1016/j.marpolbul.2020.111216>, 2020.
- Sprintall, J., Wijffels, S. E., Molcard, R., and Jaya, I.: Direct estimates of the Indonesian throughflow entering the Indian Ocean: 2004–2006, *J. Geophys. Res.-Oceans*, 114, 1–19, <https://doi.org/10.1029/2008JC005257>, 2009.
- Sruthy, S. and Ramasamy, E. V.: Microplastic pollution in Vembanad Lake, Kerala, India: the first report of microplastics in lake and estuarine sediments in India, *Environ. Pollut.*, 222, 315–322, <https://doi.org/10.1016/j.envpol.2016.12.038>, 2016.
- Stelfox, M., Hudgins, J., and Sweet, M.: A review of ghost gear entanglement amongst marine mammals, reptiles and elasmobranchs, *Mar. Pollut. Bull.*, 111, 6–17, <https://doi.org/10.1016/j.marpolbul.2016.06.034>, 2016.
- Stelfox, M., Lett, C., Reid, G., Souch, G., and Sweet, M.: Minimum drift times infer trajectories of ghost nets found in the Maldives, *Mar. Pollut. Bull.*, 154, 111037, <https://doi.org/10.1016/j.marpolbul.2020.111037>, 2020a.
- Stelfox, M., Burian, A., Shanker, K., Rees, A. F., Jean, C., Willson, M. S., Manik, N. A., and Sweet, M.: Tracing the origin of olive ridley turtles entangled in ghost nets in the Maldives: a phylogeographic assessment of populations at risk, *Biol. Conserv.*, 245, 108499, <https://doi.org/10.1016/j.biocon.2020.108499>, 2020b.
- Stramma, L., Fischer, J., and Schott, F.: The flow field off southwest India at 8° N during the southwest monsoon of August 1993, *J. Mar. Res.*, 54, 55–72, <https://doi.org/10.1357/0022240963213448>, 1996.
- Su, D., Wijeratne, S., and Pattiaratchi, C. B.: Monsoon influence on the island mass effect around the Maldives and Sri Lanka, *Front. Mar. Sci.*, 8, 645672, <https://doi.org/10.3389/fmars.2021.645672>, 2021.
- Syakti, A. D., Bouhroum, R., Hidayati, N. V., Koenawan, C. J., Boulkamh, A., Sulisty, I., Lebarillier, S., Akhlus, S., Doumenq, P., and Wong-Wah-Chung, P.: Beach macro-litter monitoring and floating microplastic in a coastal area of Indonesia, *Mar. Pollut. Bull.*, 122, 217–225, <https://doi.org/10.1016/j.marpolbul.2017.06.046>, 2017.
- Talley, L. D., Pickard, G. L., Emery, W. J., and Swift, J. H.: *Descriptive Physical Oceanography: An Introduction*, 6th Edn., Academic Press, New York, 2011.
- Taylor, M. L., Gwinnett, C., Robinson, L. F., and Woodall, L. C.: Plastic microfibre ingestion by deep-sea organisms, *Sci. Rep.*, 6, 33997, <https://doi.org/10.1038/srep33997>, 2016.
- Thushari, G. G. N., Senevirathna, J. D. M., Yakupitiyage, A., and Chavanich, S.: Effects of microplastics on sessile invertebrates in the eastern coast of Thailand: an approach to coastal zone conservation, *Mar. Pollut. Bull.*, 124, 349–355, <https://doi.org/10.1016/j.marpolbul.2017.06.010>, 2017.
- Tomczak, M. and Godfrey, J. S.: *Regional Oceanography: An Introduction*, 2nd Edn., Daya Publishing House, Delhi, 2003.
- Uneputty, P. A. and Evans, S. M.: Accumulation of beach litter on islands of the Pulau Seribu Archipelago, Indonesia, *Mar. Pollut. Bull.*, 34, 652–655, [https://doi.org/10.1016/S0025-326X\(97\)00006-4](https://doi.org/10.1016/S0025-326X(97)00006-4), 1997.
- Van Cauwenberghe, L., Vanreusel, A., Mees, J., and Janssen, C. R.: Microplastic pollution in deep-sea sediments, *Environ. Pollut.*, 182, 495–499, <https://doi.org/10.1016/j.envpol.2013.08.013>, 2013.
- van der Mheen, M.: Chapter 6: Escape pathways for buoyant debris from the Indian Ocean, in *Transport and Accumulation of Buoyant Marine Plastic Debris in the Indian Ocean*, PhD thesis, the University of Western Australia, Perth, Australia, <https://doi.org/10.26182/x6a8-9r24>, 2020.
- van der Mheen, M., Pattiaratchi, C., and van Sebille, E.: Role of Indian Ocean dynamics on accumulation of buoyant debris, *J. Geophys. Res.-Oceans*, 124, 2571–2590, <https://doi.org/10.1029/2018JC014806>, 2019.
- van der Mheen, M., Pattiaratchi, C., Cosoli, S., and Wandres, M.: Depth-dependent correction for wind-driven drift current in particle tracking applications, *Front. Mar. Sci.*, 7, 305, <https://doi.org/10.3389/fmars.2020.00305>, 2020a.
- van der Mheen, M., van Sebille, E., and Pattiaratchi, C. B.: Beaching patterns of plastic debris along the Indian Ocean rim, *Ocean Sci.*, 16, 1317–1336, <https://doi.org/10.5194/os-16-1317-2020>, 2020b.
- van Franeker, J. A.: Reshape and relocate: seabirds as transformers and transporters of microplastics, NOAA Technical Memorandum, NOS-OR&R, NOAA, 278–280, available at: https://marinedebris.noaa.gov/sites/default/files/publications-files/TM_NOS-ORR_38.pdf (last access: 28 December 2021), 2011.
- van Sebille, E., Wilcox, C., Lebreton, L., Maximenko, N., Hardesty, B. D., van Franeker, J. A., Eriksen, M., Siegel, D., Galgani, F., and Law, K. L.: A global inventory of small floating plastic debris, *Environ. Res. Lett.*, 10, 124006, <https://doi.org/10.1088/1748-9326/10/12/124006>, 2015.
- van Sebille, E., Aliani, S., Law, K. L., Maximenko, N., Alsina, J. M., Bagaev, A., Bergmann, M., Chapron, B., Chubarenko, I., Cózar, A., Delandmeter, P., Egger, M., Fox-Kemper, B., Garaba, S. P., Goddijn-Murphy, L., Hardesty, B. D., Hoffman, M. J., Isobe, A., Jongedijk, C. E., Kaandorp, M. L. A., Khatmullina, L., Koelmans, A. A., Kukulka, T., Laufkötter, C., Lebreton, L.,

- Lobelle, D., Maes, C., Martinez-Vicente, V., Maqueda, M. A. M., Poulain-Zarcos, M., Rodriguez, E., Ryan, P. G., Shanks, A. L., Shim, W. J., Suaria, G., Thiel, M., van den Bremer, T. S., and Wichmann, D.: The physical oceanography of the transport of floating marine debris, *Environ. Res. Lett.*, 15, 023003, <https://doi.org/10.1088/1748-9326/ab6d7d>, 2020.
- Veerasingam, S., Mugilarasan, M., Venkatachalapathy, R., and Vethamony, P.: Influence of 2015 flood on the distribution and occurrence of microplastic pellets along the Chennai coast, India, *Mar. Pollut. Bull.*, 109, 196–204, <https://doi.org/10.1016/j.marpolbul.2016.05.082>, 2016a.
- Veerasingam, S., Saha, M., Suneel, V., Vethamony, P., Rodrigues, A. C., Bhattacharyya, S., and Naik, B. G.: Characteristics, seasonal distribution and surface degradation features of microplastic pellets along the Goa coast, India, *Chemosphere*, 159, 496–505, <https://doi.org/10.1016/j.chemosphere.2016.06.056>, 2016b.
- Vince J. and Hardesty, B. D.: Plastic pollution challenges in marine and coastal environments: from local to global governance, *Restor. Ecol.*, 25, 123–128, <https://doi.org/10.1111/rec.12388>, 2016.
- Weiss, L., Ludwig, W., Heussner, S., Canals, M., Ghiglione, J. F., Estournel, C., Constant, M., and Kerhevé, P.: The missing ocean plastic sink: gone with the rivers, *Science*, 373, 107–111, <https://doi.org/10.1126/science.abe0290>, 2021.
- Wijeratne, S., Pattiaratchi, C., and Proctor, R.: Estimates of surface and subsurface boundary current transport around Australia, *J. Geophys. Res.-Oceans*, 123, 3444–3466, <https://doi.org/10.1029/2017JC013221>, 2018.
- Wilcox, C., Heathcote, G., Goldberg, J., Gunn, R., Peel, D., and Hardesty, B. D.: Understanding the sources and effects of abandoned, lost, and discarded fishing gear on marine turtles in northern Australia, *Conserv. Biol.*, 29, 198–206, <https://doi.org/10.1111/cobi.12355>, 2015.
- Wolanski, E. and Hamner, W. M.: Topographically controlled fronts in the ocean and their biological influence, *Science*, 241, 177–181, <https://doi.org/10.1126/science.241.4862.177>, 1988.
- Wong, P. P.: Small island developing states, *Wiley Interdisciplin. Rev. Clim. Change*, 2, 1–6, <https://doi.org/10.1002/wcc.84>, 2011.
- Woodall, L. C., Sanchez-Vidal, A., Canals, M., Paterson, G. L. J., Coppock, R., Sleight, V., Calafat, A., Rogers, A. D., Narayanaswamy, B. E., and Thompson, R. C.: The deep sea is a major sink for microplastic debris, *Roy. Soc. Open Sci.*, 1, 140317, <https://doi.org/10.1098/rsos.140317>, 2014.
- Woodson, C. B. and Litvin, S. Y.: Ocean fronts drive marine fishery production and biogeochemical cycling, *P. Natl. Acad. Sci. USA*, 112, 1710–1715, <https://doi.org/10.1073/pnas.1417143112>, 2015.
- World Bank: What A Waste Global Database, available at: <https://datacatalog.worldbank.org/dataset/what-waste-global-database>, last access: 13 August 2021.
- Wyrtki, K.: An equatorial jet in the Indian Ocean, *Science*, 181, 262–264, <https://doi.org/10.1126/science.181.4096.262>, 1973.
- Zhang, H.: Transport of microplastics in coastal seas, *Estuar. Coast. Shelf Sci.*, 199, 74–86, <https://doi.org/10.1016/j.ecss.2017.09.032>, 2017.